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Canadian Science Advisory Secretariat

## SCCS

Research Document 2009/081
Document de recherche 2009/081

# Review of DFO Science information for Atlantic salmon (Salmo salar) populations in the Southern Upland region of Nova Scotia 

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# Correct citation for this publication: <br> La présente publication doit être citée comme suit : 

Gibson, A.J.F., H.D. Bowlby, D.L. Sam, and P.G. Amiro. 2010. Review of DFO Science information for Atlantic salmon (Salmo salar) populations in the Southern Upland region of Nova Scotia. DFO Can. Sci. Advis. Sec. Res. Doc. 2009/081. vi + 83 p.


#### Abstract

The purpose of this document is to review available information for Atlantic salmon populations in the Southern Upland region of Nova Scotia in support of an assessment of the extinction risk by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC). As such, annually collected monitoring data on all life stages of salmon from two index river populations, as well as adult and juvenile abundance and distribution data from other populations in the Southern Upland, were assessed relative to conservation requirements and other reference values. Trends in abundance and distribution, changes to life-history parameter values, overall population viability, and threats to population persistence were quantified where possible.

For the two index populations (LaHave and St. Mary's rivers), current abundance of all life stages is low and populations have undergone declines in excess of $50 \%$ over the previous three generations. Mortality rates of immature and adult salmon at sea are high, and there is some indication that freshwater productivity has declined over the last 20 years. Surveys to assess the status of surrounding populations relative to the index populations suggest that current adult and juvenile abundance is critically low in most rivers, and that population extirpations may have occurred. For rivers in which salmon were found, juvenile abundance has declined markedly over the last eight years. Furthermore, estimated declines in adult escapement, as determined by changes in recreational catch or monitoring at fishways, exceed $95 \%$ in some rivers.

Estimates of current life-history parameter values for salmon in the index rivers indicate that populations may be viable at low overall population size. However, the maximum reproductive rates of the populations are extremely low, indicating little capacity for a compensatory response to further decreases in population size. This places populations at risk from the cumulative effects of environmental variability or stochastic events. In terms of recovery potential, increases to freshwater productivity have the potential to enhance long-term population viability but will not recover populations to levels above the conservation requirement without an increase in survival at sea.


Acidification is known to have substantially reduced the capacity of rivers in the Southern Upland to produce salmon and populations in highly acidified rivers are likely extirpated. Other threats to Atlantic salmon in the Southern Upland are thought to affect a large proportion of populations, yet their impact on spawner abundance in any specific population tends to be low or uncertain. However, equilibrium analyses suggest that population viability could be enhanced via increased freshwater productivity, although the extent to which habitat restoration efforts can increase freshwater production is not known. Given the current trends in salmon abundance in the Southern Upland, any actions to promote recovery will have to be timely to be effective.

## RÉSUMÉ

Le présent document a pour but d'examiner l'information disponible au sujet des populations de saumon atlantique dans la région des hautes-terres du sud de la Nouvelle-Écosse en appui à une évaluation du risque d'extinction par le Comité sur la situation des espèces en péril au Canada (COSEPAC). Par conséquent, on a évalué les données recueillies annuellement à tous les stades de vie du saumon dans deux populations de rivières-repères, ainsi que les données sur l'abondance et la répartition des adultes et des juvéniles issus d'autres populations des hautes-terres du sud, à la lumière des exigences de conservation et d'autres valeurs de référence. Dans la mesure du possible, on a quantifié les tendances relatives à l'abondance et à la répartition, les changements apportés aux valeurs des paramètres du cycle vital des poissons, la viabilité de la population et les menaces à la pérennité de la population.

Pour les deux populations-repères (rivières LaHave et Ste-Marie), l'abondance actuelle à tous les stades de vie est faible et les populations ont enregistré des diminutions de plus de 50 p .100 par rapport aux trois générations précédentes. Le taux de mortalité des saumons immatures et adultes en mer est élevé, et tout semble indiquer que le rendement en eau douce a diminué au cours des 20 dernières années. Les relevés visant à évaluer la situation des populations environnantes par rapport aux populations-repères suggèrent que l'abondance actuelle de saumons adultes et juvéniles est dangereusement faible dans la plupart des rivières, et on craint que des extirpations se soient produites. Pour les rivières où l'on a repéré des saumons, l'abondance de juvéniles avait diminué considérablement au cours des huit dernières années. De plus, les diminutions estimées des échappements des saumons adultes, déterminées par les changements observés dans les prises de pêche récréative ou la surveillance effectuée dans les passes à poissons, surpassaient 95 p .100 dans certaines rivières.

Les estimations des valeurs des paramètres du cycle vital des saumons des rivières-repères indiquent que les populations pourraient être viables malgré une taille globale réduite de la population. Cependant, le taux de reproduction maximal des populations est extrêmement faible, ce qui révèle une capacité limitée à déployer un mécanisme d'adaptation en vue d'éviter d'autres diminutions de la taille de la population. Cela pose un risque pour les populations confrontées aux effets cumulatifs de la variabilité environnementale ou aux phénomènes stochastiques. Relativement au potentiel de rétablissement, les hausses de rendement en eau douce auraient le potentiel d'améliorer la viabilité à long terme des populations, mais cela ne permettrait pas de rétablir les populations à des niveaux supérieurs aux exigences de conservation fixées sans une augmentation du taux de survie en mer.

On sait que l'acidification a contribué à diminuer considérablement la capacité des rivières des hautes-terres du sud à produire des saumons et que les populations présentes dans les rivières hautement acidifiées risquent de disparaître. On pense que d'autres menaces à la survie du saumon atlantique dans les hautes-terres du sud touchent une vaste proportion des populations, et pourtant leur incidence sur l'abondance des saumons reproducteurs dans toute population tend à être faible ou incertaine. Cependant, les analyses d'équilibre suggèrent que la viabilité de la population pourrait être améliorée par le biais d'une hausse du rendement en eau douce, bien que la mesure selon laquelle les efforts de remise en état de l'habitat puissent accroître le rendement en eau douce ne soit pas connue. Étant donné les tendances actuelles observées relativement à l'abondance de saumon dans les hautes-terres du sud, toute mesure visant à promouvoir le rétablissement devra être opportune pour être efficace.

### 1.0 INTRODUCTION

This document contains an assessment of the current status, abundance trends and dynamics of salmon populations throughout the Southern Upland region of Nova Scotia. The Southern Upland includes all rivers along the Eastern Shore and southwestern portion of the province that drain into the Atlantic Ocean (Figure 1). It has been divided into two Salmon Fishing Areas (SFAs) for management purposes: SFA 20 (Eastern Shore) and SFA 21 (Southwest Nova Scotia).

Owing to the dominant geological substrate types of the region (Roland 1982), rivers in the Southern Upland are characterized by organic acid-stained water and are typically low in dissolved minerals, which make them less productive than more mineral-rich rivers (Watt 1987). In addition, the region has been extensively impacted by sulfate deposition (acid precipitation), which has lowered the pH in many rivers. At a mean annual pH below 5.1 , salmon production is considered unstable and only remnant populations may persist (LaCroix 1985). Interspersed within the Southern Upland are limestone-rich soils (drumlins) that provide some rivers with less-acidified water.

Within the previous century, 63 rivers in the Southern Upland are known to have supported anadromous Atlantic salmon populations (DFO and MNRF 2008). As of 1986, at least 14 of these rivers were heavily acidified ( $\mathrm{pH}<4.7$ ) and were no longer able to support salmon (Amiro et al. 2000, Watt 1987). A further 20 rivers were partially acidified ( pH ranges from 4.7 to 5.0 ) and were thought to support only remnant populations.

Atlantic salmon commercial fisheries within the Maritimes Region were closed in 1985, and inriver closures of recreational fisheries for the Southern Upland began in 1998. In 2007, retention-angling fisheries for small salmon ( $<63 \mathrm{~cm}$ fork length) were restricted to four heavily acidified rivers (East River, Sheet Harbour; Mersey; Jordan; and Clyde). Five other eastern and southern shore rivers were open to a catch-and-release fishery of at least 45 days duration. Aboriginal communities have respected these conservation initiatives and generally restricted their harvests to hatchery adipose-clipped small salmon from nine rivers in Nova Scotia using methods that facilitate the live release of wild fish.

Supplementation through captive breeding and rearing has been used to enhance Atlantic salmon populations in the Maritimes Region for over a century. However, recent assessments in the Southern Upland have shown continued declines (relative to the 1980s) in the abundance of both the wild and enhanced components of salmon populations. In the cases of the acidified Liscomb, Medway, East and Tusket rivers, population enhancement of smolts did not sustain adult escapement, presumably because of low marine survival coupled with high freshwater acidity (Amiro et al. 2000).

Monitoring effort since 1986 has been focused on low- or non-acidified rivers ( $\mathrm{pH}>5.0$ ) in the Southern Upland, with the St. Mary's River in SFA 20 and the LaHave River in SFA 21 being chosen as index rivers for long-term population monitoring (Amiro et al. 2000). The status for most, if not all, low- or non-acidified rivers is expected to be similar to or worse than that of the index rivers (O'Neil et al. 1998, Amiro et al. 2000). Inference of trends obtained from the index rivers to other rivers throughout the Southern Upland is made based on recreational catch and effort data, adult count data from fish ladders, and region-wide electrofishing data.

Additional information about these populations, as well as previous assessment documents, can be found in Canadian Science Advisory Secretariat documents published by the Department of

Fisheries and Oceans (DFO) in Ottawa, the most recent being: Amiro et al. (2006) and DFO (2008).

### 2.0 LIFE-HISTORY CHARACTERISTICS OF SOUTHERN UPLAND SALMON

Wild Atlantic salmon have a complex hierarchical population structure, and their nearly precise homing to natal streams restricts gene flow among fish at different spawning locations. This can lead to local adaptation and dissimilarities in life-history characteristics among fish inhabiting geographically distant or environmentally distinct rivers. As such, anadromous Atlantic salmon populations returning to rivers in the Southern Upland exhibit a range of life-history characteristics, with differences in growth, maturation, run timing, and sex ratio among populations (Hutchings and Jones 1998, O'Connell et al. 2006).

In general, adults return to Southern Upland rivers as first-time spawners after one or two winters at sea and most enter the rivers throughout the spring (May/June) and summer (July/August) months. Timing is partially determined by river flow. Spawning takes place in late October or November, and spawned-out adults (kelts) either return to sea immediately or remain in fresh water until the following spring (O'Connell et al. 2006). Juveniles emerge from the gravel in early spring and typically spend three years in fresh water before undergoing smoltification. The abundance of juvenile salmon in fresh water is regulated by densitydependent factors, although the timing of density dependence is variable among populations (Gibson 2006). No evidence for density-dependent survival has been found in the marine environment, presumably because survival in the marine environment is not resource-limited for Southern Upland populations.

### 3.0 PRESENT STATUS AND RECENT TRAJECTORIES IN ABUNDANCE, RANGE AND NUMBER OF POPULATIONS

### 3.1 St. Mary's River

### 3.1.1 Background

The St. Mary's River is one of the major river systems in SFA 20 and consists of two main branches: the West Branch and the East Branch. In general, water in the East Branch is less acidified than water in the West Branch because the underlying soils are rich in base minerals; however, there is variation in pH among tributaries throughout the system. In addition, geological changes to the course of the river system over time (Roland 1982) have resulted in differences between the two branches in terms of the distribution of salmon habitat and the resulting productivity. Anecdotal reports also note a difference in the life-history characteristics of the returning adult population, with a higher proportion of large multi-sea-winter (MSW) fish historically returning to the East Branch (O'Neil and Harvie 1995).

In the St. Mary's River, approximately $80 \%$ of Atlantic salmon juveniles spend two years in fresh water and migrate to sea as two-year-old smolts (Amiro et al. 2006). Smolts exit the river in May, while adults return to the river predominantly during mid-spring and early summer (June/July/August). Historically, adult returns to the system were characterized by a relatively high proportion of two-sea-winter (2SW) and three-sea-winter (3SW) salmon (14\% and 9\%, respectively), of which approximately $60 \%$ were female (Marshall 1986), although one-seawinter (1SW) salmon are thought to have been more prevalent in the West Branch. More recent assessments have shown a reduction in the number of virgin 3SW adults returning to the river
(O'Neil and Harvie 1995). Current assessments in the West Branch indicate that most salmon are maturing after one winter at sea (e.g., 92\%; Amiro et al. 2006). The length-fecundity relationship (Marshall 1986) calculated from adult broodstock collections during the period 1972 to 1985 is:

$$
\text { Fecundity }=340.832 \mathrm{e}^{0.0389 * \text { Fork Length }} \text {. }
$$

## Conservation Requirement

Although alternate calculations exist (O'Connell et al. 1997), the conservation requirement used for recent assessments on the St. Mary's River is 7.4 million eggs, which is equivalent to approximately 3,155 adult salmon (O'Neil et al. 1998, Amiro et al. 2006). This egg requirement is based on the estimated habitat area suitable for juvenile production in the river and a target egg deposition rate of $2.4 \mathrm{eggs} / \mathrm{m}^{2}$ (Marshall 1986).

### 3.1.2 Status

The annual status of salmon in the St. Mary's River is assessed using a combination of approaches including fishery-independent adult counts, smolt counts and juvenile surveys. Prior to 1996, adult escapement estimates were derived from recreational catches and the mean annual exploitation rate imported from the LaHave River (O'Neil et al. 1998). However, riverspecific escapement estimates have been calculated since 1997 using mark-recapture seining experiments to estimate adult abundance (Table 1). Although attempts were made to sample both the East and West branches of the river, sufficient marks and recaptures could only be obtained for salmon in the West Branch. Recent seining efforts have focused on the West Branch exclusively. From 2002 to 2005, high water levels after the initial seining date precluded a second seining attempt to complete the mark-recapture experiment. Therefore, it was necessary to assume that the efficiency was equal to the mean catchability calculated for 1997 to 2001. Catchability for the marking pass is the number of fish caught on the marking pass divided by the adult population estimate. In past assessments, an attempt was made to scale the West Branch results up to the entire river by dividing the West Branch estimate by 0.55 , which is the proportion of habitat available in the West Branch compared to that of the entire river (Amiro 1993, Amiro et al. 2000). The possibility that environmental factors like water quality, discharge, or substrate distribution differentially affect habitat production capacities among the West Branch, East Branch and Main River is not accounted for when production is scaled solely based on habitat area. For this reason, the most recent assessments have reported adult abundance estimates for the West Branch only.

## Adults

The mark-recapture seining experiment on the West Branch typically takes place at the end of September and beginning of October each year, once the majority of adult salmon have entered the river. Three separate pools are seined over two days during both the marking and the recapture trips using a knotless nylon net ( $\sim 100 \mathrm{ft}$. x 8 ft .). All salmon caught are sampled for biological characteristics (sex, length, scale and DNA samples), and fish from the three pools are given a different mark during the marking pass (used to evaluate the degree of mixing among pools). Other fish species present (generally brook trout and white sucker) are counted but population estimates are not attempted. Approximately two weeks separate the marking trip from the recapture trip to allow for random mixing of marked fish in the population.

In 2008, a total of 30 salmon were marked, 63 were captured and 4 were recaptured, giving a corrected Petersen estimate of escapement of 397 salmon (95\% C.I. $=175-781$ ). This estimate
represents the first notable increase in escapement in 5 years (Table 1). Based on the scale samples taken from captured fish, $91 \%$ of the population was 1 SW, $7 \%$ was 2 SW and $2 \%$ were repeat spawners (Table 2). No 3SW fish were captured.

The estimated escapement in 2008 for the West Branch of the St. Mary’s River ( 397 fish) is approximately $23 \%$ of the conservation requirement for the West Branch (55\% of 3,155 salmon), which is the highest value recorded within the last 4 years (Table 5).

## Smolts

Smolt monitoring on the St. Mary's River uses a rotary screw trap (RST) anchored off Glenelg Bridge on the West Branch. The RST is typically deployed in late April and is fished daily until early June. All smolts captured have biological samples taken (length and scale samples) and all other fish species are counted. A mark-recapture experiment, where all captured smolts are marked and released upstream of the RST, is used to calculate the efficiency of the RST. Smolt population estimates (2005-2008) are calculated based on a corrected Petersen estimate of the total number of marked and recaptured smolts.

In 2008, the RST was deployed on April 30, but some mechanical problems prevented sampling until May 8 . In total, 485 smolts were captured, and, of these, 15 fish were tagged, indicating that they had been captured a second time. This gives a population estimate of $15,217(95 \%$ C.I. $=9,451-24,154)$ smolts and a capture efficiency of $3.1 \%$ for the smolt wheel. These values are slightly below the 2007 estimates of 16,110 ( $95 \%$ C.I. $=12,735-20,835$ ) smolts with a capture efficiency of $5.4 \%$. Of the smolts sampled in $2008(n=254), 91 \%$ were age-2 and $9 \%$ were age- 3 . On average, age- 2 smolts were approximately 2 cm smaller than age-3 smolts, with mean fork lengths of 14.5 cm and 16.4 cm , respectively.

The estimated area of the juvenile habitat in the St. Mary's River is $3,078,000 \mathrm{~m}^{2}$ (Marshall 1986), $55 \%\left(1,692,900 \mathrm{~m}^{2}\right)$ of which is located in the West Branch. Based on this value, smolt production in 2008 was 0.90 smolts per $100 \mathrm{~m}^{2}$. This estimate is low relative to the previous 2 years (Table 3) and to the reference value of 3.8 smolts $/ 100 \mathrm{~m}^{2}$ sometimes used for Atlantic salmon (Symons 1979). Given the productive habitat area in the West Branch of the St. Mary's River, 3.8 smolts $/ 100 \mathrm{~m}^{2}$ would place the natural production capacity for the branch at 64,330 smolts annually.

## Juveniles

Electrofishing sampling sites on the St. Mary's River were chosen using a random stratified design, where gradient strata were sampled in proportion to the amount of juvenile habitat (based on stream gradient and distance from the river mouth) (Amiro et al. 1989, Amiro 1993). Between 9 and 37 sites throughout the St. Mary's River have been electrofished annually, typically starting in July and finishing in August of each year. Provided that a sufficient number of juveniles were initially captured, salmon abundance within a site was estimated using Petersen mark-recapture methods, and catchability was estimated by dividing the number of marked fish by the number caught during the recapture pass (Amiro et al. 1989). Annual mean catchability was used to estimate fish abundance for sites where mark-recapture was not possible. The densities of three age classes, age-0, age-1, and age-2+ (age-2 plus the small proportion of individuals that remained in the river for longer than two years) were estimated for each site based on ages determined by reading scales.

In 2008, the mean densities in the entire St. Mary's River were calculated based on data from 12 sites. Estimated overall age-0, age-1 and age-2+ densities were 6.1, 2.5 and 0.3 fish per
$100 \mathrm{~m}^{2}$, respectively, and were relatively consistent between the two branches (Table 4). Fry (age-0) and age-1 parr densities were lower in 2008 than in 2007, but the density of age-2 parr was approximately double the 2007 value (Figure 2).

The mean fry density observed in 2008 is consistent with the predicted linear relationship between estimated adult salmon returns for the years 1993-2007 and subsequent fry density in the following year (Figure 3). The similarity between the predicted and observed estimate of fry density for 2008 suggests that adult escapement in 2007 (203 fish) and 2008 fry densities are consistent with each other.

### 3.2 LaHave River

### 3.2.1 Background

The LaHave River drains approximately $1,670 \mathrm{~km}^{2}$ of the Southern Upland region of Nova Scotia, and is one of the largest watersheds in SFA 21. It contains 113 lakes with a total surface area of 7,515 ha, and consists of five major sub-drainages: West Branch, North Branch, Ohio River, North River and the Main Stem (Gray et al. 1989). Throughout its length, the LaHave River contains several natural and manmade barriers to salmon migration. One of the larger obstacles is a natural waterfall named Morgans Falls, which is presently the site of a hydroelectric facility built in 1995. Morgans Falls is on the Main Stem of the LaHave River and is downstream of the Ohio and North river sub-drainages. Prior to the 1960s, Atlantic salmon had limited access to the watershed upstream of Morgans Falls. In the late-1960s, a fishway was constructed to bypass the falls and DFO began a stocking program to enhance the developing salmon run. Counts of adult salmon returning to the fishway began in 1970.

The smolt enhancement program on the LaHave River ran from 1969 to 2003. The first broodstock were taken from the nearby Medway River, and the first hatchery-reared smolts were released above Morgans Falls in 1971. From 1971 to 2005 (excluding 1982), the LaHave River was stocked annually with hatchery-reared smolts. From 1972 to 2003, all broodstock were collected at the Morgans Falls fishway. From 1996 to 2003, broodstock selection was proportional to the wild and hatchery components of the returning population (Amiro et al. 2006). The proportion of adults of hatchery origin contributing to annual egg deposition has ranged from $0 \%$ to $94 \%$. Overall, 1SW salmon (wild and hatchery) contribute approximately 1,240 eggs per fish annually, while 2 SW salmon contribute an average of 5,120 eggs per fish. Despite differences in the spawning escapement among large (> 63 cm fork length) and small ( $\leq 63 \mathrm{~cm}$ fork length) salmon, each size class contributes approximately $50 \%$ of the total annual egg deposition above Morgans Falls (Amiro, unpublished data).

## Conservation Requirement

The annual assessment on the LaHave River is based primarily on an estimate of the egg deposition upstream of Morgans Falls, determined from the salmon counts and biological data collected at the Morgans Falls fishway. This value is compared to the conservation requirement for that area. There is some uncertainty of the effects of acidification on the provision of appropriate conservation requirements for the LaHave River above Morgans Falls. Cutting and Grey (1984) calculated the required number of spawners for the entire LaHave River ( 3,312 fish) based on the target egg deposition rate of $2.4 \mathrm{eggs} / \mathrm{m}^{2}$, an estimated rearing area of $2,046,228 \mathrm{~m}^{2}$, and an average fecundity of 1,482 eggs per fish. The current conservation requirement of 1.96 million eggs for the LaHave River upstream of Morgans Falls assumes the same fecundity per fish and target egg deposition rate but is based on the assumption that only $40 \%$ of the rearing area is usable (Amiro et al. 1996). This interim conservation requirement has
been used to assess the status of the population in recent years (Amiro et al. 2006), although alternate calculations for both rearing area (Amiro et al. 1996) and conservation requirements (O'Connell et al. 1997) exist.

### 3.2.2 Status

Upstream-migrating adult salmon have been counted at the Morgans Falls fishway since 1970, and downstream migrating smolts have been counted each May since 1996 (Amiro et al. 2006). Currently, scale samples are taken from all wild adults in the fishway and from every fifth smolt for aging purposes. The collection of adult broodstock for enhancement ceased at Morgans Falls in 2003, with the last release of hatchery-reared smolts occurring in 2005.

## Adults

The total count of adult salmon at the Morgans Falls fishway in 2008 was 691 fish (593 small and 98 large salmon), none of which were of hatchery origin (Figure 4). This is consistent with values for total returns observed since 1997, but represents a notable increase from returns in 2007. Age and spawning history were determined using scale samples from 684 of the captured adults, indicating that $86 \%$ of the population were 1 SW salmon, $13 \%$ were 2 SW salmon and $1 \%$ were repeat spawners (Table 6).

The estimated egg deposition above Morgans Falls was 1,078,475 eggs in 2008, equating to $55 \%$ of the conservation requirement (Figure 6, Figure 7). Because all adults captured at Morgans Falls from 1970 to 2008 were aged, it was possible to do a cohort analysis to calculate the number of adult recruits per adult spawner. Recent values of this statistic are negative, which demonstrates that successive generations of the salmon population above Morgans Falls in the LaHave River have not replaced themselves since the 1985 escapement year (Figure 8).

## Smolts

Similar to the St. Mary's River, captured smolts are marked and released upstream of Morgans Falls for a mark-recapture experiment. A corrected Petersen estimate of the total number of marked and recaptured smolts is used to estimate population size. The catchability of the downstream fishway is estimated as the proportion of recaptured to marked fish. In 2008, a total of 14,450 wild smolts $(90 \%$ C.I. $=13,500-15,500)$ were estimated to have migrated from above Morgans Falls, a $41 \%$ decline from the 2007 value and less than the 1996-2007 mean of 16,589 smolts (Table 7).

In 2008, approximately 20\% of the smolts that were captured at Morgans Falls were measured and aged ( 1,239 individuals). Of these, $87 \%$ were age- $2,13 \%$ were age- 3 , and $0.3 \%$ were age4. Mean fork lengths of age-2, age-3, and age-4 smolts were 16.4 cm (range: 13.0-20.1 cm), 18.9 cm (range: $16.0-23.5 \mathrm{~cm}$ ), and 23.7 cm (range: 20.6-28.5 cm), respectively. These values are higher than those for smolts found in the St. Mary's River.

Smolt production in 2008 was 0.55 smolts per $100 \mathrm{~m}^{2}$, less than the long-term mean (19962007) of 0.64 smolts per $100 \mathrm{~m}^{2}$ (Table 7). Egg-to-smolt survival was $1.3 \%$, similar to the longterm mean (1996-2007) of $1.2 \%$. Both of these parameters indicate relatively low freshwater production of juvenile salmon in 2008.

The ratio between smolt production and subsequent adult returns provides an estimate of the return rate of smolts (indicative of at-sea survival). For the LaHave River above Morgans Falls, return rates have ranged from $1.13 \%$ to $7.95 \%$ for 1 SW adults and $0.11 \%$ to $0.86 \%$ for 2 SW
adults (Table 7). The estimate of the return rate of wild smolts emigrating from above Morgans Falls in 2007 to 1SW returns in 2008 was $2.33 \%$, less than the long-term mean of $2.8 \%$. The estimate of the return rate of wild smolts emigrating in 2006 to 2 SW returns in 2008 was $0.4 \%$, a value identical to the long-term mean.

## Juveniles

Electrofishing site selection and mark-recapture methods for estimating juvenile densities are the same as on the St. Mary's River (see Section 3.1.2). The number of electrofishing sites fished annually on the LaHave River has ranged from 7 to 30. In 2008, a total of 8 electrofishing sites, three located above Morgans Falls and five located below, were surveyed (Table 8). All juvenile salmon captured were marked during the initial electrofishing pass at each site, and 5 of the 8 sites were revisited for the recapture pass. The density of each age class (age-0, age-1 and age-2+) at each 2-pass site was calculated using a Petersen estimate (Gibson et al. 2003, Gibson and Amiro 2003), and density at each single-pass site was estimated based on total catch multiplied by the 2-year (2007 and 2008) mean estimate of efficiency for age-1 parr on an electrofishing pass. Unfortunately, density at one site in 2008 could not be calculated using either method because the area that was electrofished was not recorded.

Parr density (age-1, age-2, and older juveniles combined) in 2008 was higher below Morgans Falls than above, with mean parr densities of 13.9 and 4.5 parr per $100 \mathrm{~m}^{2}$, respectively (Appendix 1, Figure 5). The 2008 mean parr density above Morgans Falls was the lowest estimate since 1984, yet the 2008 mean parr density below Morgans Falls was the highest estimate since 1999.

### 3.3 Abundance Trends and Information for other Southern Upland Rivers

### 3.3.1 2008 Electrofishing Survey

In 2008, a region-wide electrofishing survey for Atlantic salmon juveniles and other fish species was undertaken in the Southern Upland. A catchability of $42.8 \%$ for salmon (Gibson et al. 2003) was used to calculate density at single-pass electrofishing sites and a depletion experiment was used to calculate efficiency at multi-pass sites. A total of 143 sites were surveyed in 51 rivers, with between 1 and 12 sites fished per river (Figure 9, Appendix 2). Considering only the first pass of each survey, 143,385 seconds of shocking effort was applied over $98,019 \mathrm{~m}^{2}$ of habitat, resulting in the capture of 3,474 fish, 977 of which were Atlantic salmon (Appendix 2). Salmon juveniles were captured at 52 of the 143 sites ( $36.4 \%$ ) and were found in 20 of the 51 rivers surveyed ( $39.2 \%$ ). American eel were the most commonly captured species ( 1,555 fish), followed by juvenile salmon (994 fish), and then by brook trout ( 333 fish).

Where present, the observed densities of juvenile salmon ranged from 0.3 to 33.9 fish per $100 \mathrm{~m}^{2}$ (Figure 10). Observed densities of fry (age-0) ranged from 0.3 to 28.0 fish per $100 \mathrm{~m}^{2}$ and of parr (age-1 and age-2) ranged from 0.2 to 16.2 fish per $100 \mathrm{~m}^{2}$, with the highest values being recorded on the Musquodoboit River (Table 9). In six rivers, only one life stage was found (either fry or parr), yet it is likely that additional effort or alternate site selection would have resulted in the capture of the other life stage in the system. In rivers where both life stages were found, mean age-0 densities (range: 0.04-10.3 fish $/ 100 \mathrm{~m}^{2}$ ) were typically higher than age-1 and older densities (range: 0.04-7.5 fish/100 $\mathrm{m}^{2}$ ). In general, the mean density of either age class was much lower than Elson's norm ( 30 age- 0 fish $/ 100 \mathrm{~m}^{2}$ and 24 age- 1 and older fish/ $100 \mathrm{~m}^{2}$ ), values that have been used as a reference for juvenile production in fresh water (Elson 1967, Elson 1975).

### 3.3.2 Comparison of the 2000 and 2008 Electrofishing Surveys

Electrofishing surveys in the Southern Upland during 2000 (Figure 11, Appendix 3) and 2008 (Figure 9, Appendix 2) were similar in terms of total effort and coverage. Marginally more sites were completed in 2008 (143 versus 128), yet one less river was visited ( 51 versus 52). Total shocking time was slightly greater in 2008 ( 143,385 seconds versus 104,331 seconds), but the total area surveyed on the first pass at each site was lower ( $98,019 \mathrm{~m}^{2}$ versus $128,842 \mathrm{~m}^{2}$ ). However, less than half as many fish were captured on the first pass in $2008(3,474)$ as in 2000 $(7,825)$, including approximately one quarter as many salmon (977 versus 3,733). In 2000, juvenile Atlantic salmon were found in $54 \%$ of the rivers (28 of 52) rather than $39 \%$ (20 of 51) as in 2008.

At sites were juvenile salmon were captured, the total juvenile salmon density in 2000 ranged from 0.1 to 99.6 fish per $100 \mathrm{~m}^{2}$ (Figure 12), which is approximately 3 times higher than maximum densities at a site in 2008. Observed densities of the total number of fry ranged from 0.1 to 86.3 fish per $100 \mathrm{~m}^{2}$, and of parr ranged from 0.1 to 31.2 fish per $100 \mathrm{~m}^{2}$ in 2000 , with the highest values recorded on the Musquodoboit River (Table 10). Overall, the mean density of age-0 juveniles declined from 5.0 to 1.2 fish per $100 \mathrm{~m}^{2}$ between 2000 and 2008, while the mean density of age-1 and older parr decreased from 3.5 to 0.9 fish per $100 \mathrm{~m}^{2}$. In addition, juvenile salmon were absent in 7 sites and 2 rivers in 2008, where they were previously found in 2000 (Figure 13).

Of the sites surveyed in both years ( $n=74$ ), total juvenile density decreased in 43\% ( $n=32$ ) and increased in $8 \%(n=6)$ (Figure 14). The remainder of the sites $(n=36)$ had recorded densities of zero for both years. Any increase from the 2000 density was very small (as shown by the proximity of the points to the $1: 1$ line) while declines tended to be quite large. When the data were separated by age class, similar results were obtained. Any increase in fry (age-0) density was relatively small, while the declines were much larger, the most extreme example being a site with $>80$ fry per $100 \mathrm{~m}^{2}$ in 2000, but < 5 fry per $100 \mathrm{~m}^{2}$ in 2008 (Figure 15). The trend was not as dramatic for parr (age-1 and older), but the most extreme example still shows a reduction in density from $>30$ parr per $100 \mathrm{~m}^{2}$ in 2000 to $<6$ parr per $100 \mathrm{~m}^{2}$ in 2008 (Figure 16).

### 3.3.3 Recreational Catch and Effort

Catch and effort data from the annual recreational salmon fishery have been collected using a license-stub return program since 1983. After the close of the fishing season, anglers send in their stubs during autumn and winter. Preliminary estimates of the season's catch and effort are provided the following spring, and estimates are finalized during the next year.

In 2007, the majority of rivers were closed to angling, except for 5 rivers in SFA 20 and 9 rivers in SFA 21 which were open from June 1 to July 15 (Appendix 4). Total fishing effort in 2007 was concentrated in 2 rivers: the St. Mary's in SFA 20 (597 rod-days) and the LaHave in SFA 21 ( 497 rod-days), with 297 adult salmon caught on the St. Mary's and 117 caught on the LaHave (Appendix 4). Catches and effort on nearly all other rivers were an order of magnitude lower. Of the 8 other rivers open to angling in SFA 21, at least one salmon was caught on 3 rivers and the highest number caught was 13. In 2007, salmon were caught on all of the other 4 rivers open to angling in SFA 20, but the greatest number caught was only 35 fish. The majority of fish captured were small salmon (74\%) and a total of five of these were retained (i.e., removed from the river).

To summarize changes in reported recreational catch and effort (Appendix 5, Appendix 6), log mean catch and effort for each of the time periods of: 1988-1993, 1994-1999 and 2000-2007 were compared with the log mean from 1983-1987. Increases or decreases in catch or effort were summarized in terms of a percent change between the two time periods. In this way, it was possible to demonstrate progressive changes in effort and catch over time. Each time period corresponds to roughly one generation for Atlantic salmon in SFAs 20 and 21, except for the most recent time period (2000-2007) that was grouped together because of the scarcity of data (many rivers were closed to angling in 1998).

For the majority of rivers in SFA 20, a comparison of the reported recreational catch during the 1983-1987 time period with the 2000-2007 time period shows that the reported catch has declined, often by > 95\% (Figure 17). Four of the 5 rivers open to angling in 2007 show a >90\% decline in recreational catch. Concurrent with the decline in reported catch has been a decline in reported effort on most rivers, which dropped by nearly $100 \%$ before the rivers were closed to angling (Figure 18). One exception in terms of catch and fishing effort has been St. Francis Harbour River (labeled Saint Francis on Figures 17 and 18), which has seen an increase in recreational fishing effort and catch in recent years compared with the 1983-1987 time period. Although the increase appears substantial, it represents a change from 2 salmon caught in 4 rod-days in 1983 (with no salmon captured from 1984 to 1987), to 7 salmon caught in 3 roddays in 2007 (the river was closed to recreational fishing from 2000 to 2006) (Appendix 5).

For the majority of rivers in SFA 21, reported recreational catch has declined over time. Of the 9 rivers that remained open to angling in 2007, all of them have shown a $>75 \%$ reduction in recreational catch between 1983-1987 and 2000-2007 (Figure 19). In conjunction, fishing effort has declined by $>95 \%$ in most rivers (Figure 20), with the exception being the Sackville River where a stocking program has been in place since the 1980s. Both recreational catch and effort have increased on the Sackville River since the stocking program began (Figures 19 and 20). However, the mean catch and fishing effort for the most recent time period (2000-2007) is low, at 3.4 salmon caught in 34.4 rod-days (Appendix 6 ).

In general, any decline in reported recreational effort lagged behind the decline in reported recreational catch. For example, in the early 1990s the distribution of recreational effort in SFA 20 changed slightly, with approximately the same number of rivers showing an increase as a decrease (Figure 18). However, with the exception of four rivers, recreational catches declined between the 1983-1987 and 1988-1993 time periods (Figure 17). This pattern is also evident for rivers in SFA 21, where the river-specific decline in recreational catch tends to be greater than the decline in fishing effort for a given time period.

### 3.3.4 Trends in Adult Salmon Abundance in the Southern Upland

Adult abundance time series data are available for four populations in the Southern Upland. These are the St. Mary's and LaHave populations, described above, as well as counts of adult salmon ascending fishways on the East River, Sheet Harbour and the Liscomb River. Counting facilities on the East River, Sheet Harbour and the Liscomb River are relatively close to the mouth of their respective river system so the counts are thought to be representative of the majority of returns in each year. On the Liscomb River, returns were predominantly 1SW adults and counts peaked in 1987 when 1,614 1SW and 88 MSW fish returned to the river (Table 11). Beginning in 1990, annual counts declined progressively until the last year of monitoring in 1999, when only nine 1SW and no MSW adults returned to the river. Counts in East River, Sheet Harbour were low relative to the other rivers in the Southern Upland, and generally ranged from 25-100 1SW and < 10 MSW adults annually (Table 11). However, in the most recent 9 years, the highest count has been 3 salmon. The low counts in more recent years
relative to the rest of the time series suggest that salmon populations in the Liscomb River and in East River, Sheet Harbour have substantially declined.

When assessing the status of wildlife in Canada, COSEWIC uses trends in abundance over 10 years or 3 generations (whichever is longer) as one of the criteria for determining the extent to which a species is at risk of extinction. Here, two different methods were used to evaluate declines over the previous 3 generations for Southern Upland populations following the approach of Gibson et al. (2006). The first was to calculate the extent of the decline as the ratio of the population size in two 5 -year time periods separated by 15 years (or as close to 15 years as possible given the available monitoring data). The model was fit using maximum likelihood and confidence intervals for the step function were calculated using likelihood ratios. The second method estimated decline rates using a log-linear model, which was fit to the data using least-squares after transformation onto the log scale.

Declines in total escapement for 4 rivers in the Southern Upland (St. Mary's, LaHave, Liscomb, and East River Sheet Harbour) could be assessed using the models described above. In all cases, similar decline rates were found using the 2 models (c.f. 15-Year and Step function in Table 12 for St. Mary's, LaHave, and East River, Sheet Harbour; and 10-Year and Step function for Liscomb) and the confidence intervals did not straddle zero, indicating that the declines were significantly different from zero (Table 12). Decline rates on the St. Mary's (slope $=-0.080$ ) and LaHave rivers (slope $=-0.056$ ) were much lower than those predicted for East River, Sheet Harbour (slope $=-0.243$ ) or Liscomb River (slope $=-0.458$ ). Over the last 10 years when data were available, the salmon population in the Liscomb River was predicted to have declined by $>99 \%$ (Figure 21). The predicted population decline over the last three generations of salmon in East River, Sheet Harbour is also extreme, at > 97\% (Figure 22). Although decline rates have been lower on the LaHave and St. Mary's, the populations are still predicted to have declined by $>56 \%$ (Figure 23) and $>69 \%$ (Figure 24), respectively, over the previous three generations.

### 4.0 POPULATION DYNAMICS

The population dynamics of Southern Upland salmon were analyzed using a population model consisting of two parts: a freshwater production model that provides estimates of the expected smolt production as a function of egg deposition, and an egg-per-smolt model that provides estimates of the rate at which smolts produce eggs throughout their lives. These components are combined via an equilibrium analysis that provides estimates of the abundance at which the population would stabilize if the input parameters remained unchanged. This combined model can also be used to evaluate how population size would be expected to change in response to changes in carrying capacity, survival, or life stage transition probabilities.

### 4.1 Freshwater Production Model

The values of several demographic parameters were estimated, including freshwater carrying capacity, age- and stage-specific survival rates, smoltification probabilities and maturation rates using the model of Gibson et al. (2009). The model begins with the annual egg deposition in the river, and it uses estimates of age-specific survival and smoltification probabilities to link egg deposition to smolt production:

Let $P_{t, 0}$ be the number of age-0 salmon (fry) in year $t$. The relationship between the number of age-0 salmon in the spring and egg production during the previous fall is modeled as:

$$
P_{t, 0}=E g g_{t-1}\left(1-M_{E g g}\right),
$$

where $E g g_{t}$ is egg deposition in year $t$, and $M_{E g g}$ is the density-independent mortality of eggs and newly emerged fry (i.e., all mortality that occurs from the time of egg deposition to the time of sampling by electrofishing the following summer).

For Atlantic salmon, density dependence is known to occur in fresh water (Chaput and Jones 2006, Gibson 2006), but little or no evidence exists for its presence in the marine environment (Jonsson et al. 1998. Armstrong 2005, Gibson 2006), presumably because marine survival is not resource-limited. Therefore, density dependence is modeled as density-dependent survival in fresh water. This is the equivalent of assuming that the carrying capacity of the freshwater habitat limits production and that, as a result of competition for limited resources, survival decreases as cohort size increases. In a comparison of nine populations, Gibson (2006) found that the timing of density dependence in fresh water varied among populations. In the case of the St. Mary's and LaHave populations, it occurred between age-0 and age-1. No evidence of overcompensation was found in his analyses, so density dependence is incorporated into the model using a Beverton-Holt function (Hilborn and Walters 1992).

In year $t$, the number of age-1 parr, $P_{t, 1}$, is given by:

$$
P_{t, 1}=\frac{\alpha P_{t-1,0}}{1+\frac{\alpha P_{t-1,0}}{R_{a s y} h}}\left(1-j_{1}\right),
$$

where $\alpha$ is the slope at the origin of the Beverton-Holt function and describes the maximum survival rate between age-0 and age-1, $R_{a s y}$ is the asymptotic density of age-1 parr (number per $100 \mathrm{~m}^{2}$ habitat units), $h$ is the number of habitat units available to the population, and $j_{1}$ is the probability that a fish emigrates as a smolt at age-1. Thus, $R_{\text {ass }} h$ is the carrying capacity of the river for wild age-1 parr. The model is formulated this way because the electrofishing data, used as indices of the abundance of parr, are reported as a density (number per $100 \mathrm{~m}^{2}$ ), whereas the total number of parr in the river are of interest here.

For a given age $a$, the number of age-2 and older parr is determined by the number of parr in the cohort from the previous year, the annual mortality rate of parr $M_{\text {Parr }}$ (assumed to be density-independent), and the age-specific probability of smoltification $j_{a}$ :

$$
P_{t, a}=P_{t-1, a-1}\left(1-M_{P a r r}\right)\left(1-j_{a}\right) .
$$

Wild age-1 smolts have not been observed in the St. Mary's River, and all parr have undergone smoltification by age 3. The number of smolts in year $t$ of age $a\left(S_{t, a}\right)$ is then:

$$
S_{t, a}=P_{t-a, a-1}\left(1-M_{\text {Parr }}\right) j_{a}
$$

### 4.1.1 Data Available to Parameterize the Freshwater Production Model

St. Mary's River

For the St. Mary's River, data inputs for the life-history model were exclusively from monitoring on the West Branch of the river. This was done to avoid any uncertainty in the model associated with scaling up abundance estimates calculated from the West Branch to the entire river based on habitat area. The observed data that were used in the life-history model were: (1) smolt counts and biological characteristics (age and size distribution) from the years 2005 to 2008 (Table 3), (2) mark-recapture electrofishing survey data from which juvenile density by age class was estimated from 1990 to 2008 (Table 13), (3) recreational catch estimates by size class from a volunteer license-stub return program for the years 1984 to 1996 (Table 14), and (4) markrecapture seining estimates of adult abundance in the West Branch from 1997 to 2008 (Table 1).

The life-history model was fit using data spanning 3 overlapping time periods: 1984-2008, 19932008 and 1997-2008. The first time period corresponds to the years when egg deposition data could be calculated either from the recreational catch (with an assumed catch rate equal to the median catch rate observed on the LaHave River) or from mark-recapture experiments and the length-fecundity relationship for adults on the St. Mary's River. The second time period encompasses the previous 3 generations of Atlantic salmon in the St. Mary's River. The third time period corresponds to the years during which fishery-independent data are available to estimate escapement. Given the overlap in the data, parameter estimates obtained for each time period are correlated and cannot be directly compared with one another (i.e., to test for changes in parameter values over time). However, it is useful to present multiple analyses of the dataset and to look at how conclusions drawn from each differ with respect to salmon status in the St. Mary's River.

## LaHave River

For the LaHave River, estimates of annual egg deposition were available for the period 1973 to 2008 (Table 15), and estimates of the number and age composition of emigrating smolts were available for the years 1996 to 2008 (Table 6). When applied to this population, the life-history model above was simplified into a Beverton-Holt stock-recruit function describing freshwater production from egg to smolt. Given that smolt data have only recently been collected, the estimated freshwater carrying capacity and maximum rate of population increase from egg to smolt describe current freshwater conditions (post-1996) in the LaHave River.

### 4.1.2 Parameter Estimation

## St. Mary's River

For the St. Mary's River, parameter estimates were calculated by simultaneously fitting the model to the observed data using maximum likelihood, by minimizing the value of an objective function, O.F.V. (Quinn and Deriso 1999). The O.F.V. equals the sum of the negative loglikelihoods for the juvenile electrofishing data ( $\ell_{\text {electro }}$ ), the smolt age-frequency data ( $\ell_{\text {age }}^{\text {smolt }}$ ), the egg deposition data ( $\ell_{\text {egg }}$ ) and the smolt count data ( $\ell_{\text {smolt }}$ ). Lognormal error structures (Myers et al. 1995) were used for all likelihoods except the smolt age-frequency data, for which a multinomial likelihood (Quinn and Deriso 1999) was used. Following the approach by Gibson et al. (2009), the mean standard deviation of the likelihood functions ( $\sigma$ ) was set at 0.33:

$$
\begin{gathered}
\ell_{\text {electro }}=\sum_{a}\left(-n \ln \sigma_{\text {elect }} \sqrt{2 \pi}-\sum_{t} P_{t, a}^{\text {obs }}-\frac{1}{2 \sigma_{\text {electro }}^{2}} \sum_{t}\left(\ln P_{t, a}^{\text {obs }}-\ln \left(P_{t, a} / h\right)\right)^{2}\right), \\
\ell_{\text {smolt }}=-n \ln \sigma_{\text {smolt }} \sqrt{2 \pi}-\sum_{t} S_{t}^{\text {obs }}-\frac{1}{2 \sigma_{\text {smolt }}^{2}} \sum_{t}\left(\ln S_{t}^{\text {obs }}-\ln S_{t}\right)^{2}, \\
\ell_{\text {egg }}=-n \ln \sigma_{\text {egg }} \sqrt{2 \pi}-\sum_{t} E g g_{t}^{\text {obs }}-\frac{1}{2 \sigma_{\text {egg }}^{2}} \sum_{t}\left(\ln E g g_{t}^{\text {obs }}-\ln E g g_{t}\right)^{2}, \\
\ell_{\text {age }}^{\text {smolt }}=\sum_{t} \log \left(\frac{n_{\text {smolt }, t}!}{\left(x_{\text {smolt }, t, 1}!\right)\left(x_{\text {smolt }, t, 2}!\right) \ldots . .\left(x_{\text {smolt } t, r, r}!\right)} p_{\text {smolt }, t, 1,1}^{x_{\text {solt }} \ldots . . p_{\text {smolt }, t, r}}{ }^{x_{\text {smolt }, r}}\right), \\
O . F . V .=-\left(\ell_{\text {egg }}+\ell_{\text {smolt }}+\ell_{\text {electro }}+\ell_{\text {age }}^{\text {smolt }}\right) .
\end{gathered}
$$

The model was programmed using AD Model Builder (Fournier 1996). AD Model Builder (ADMB) uses the C++ auto-differentiation library for rapid fitting of complex non-linear models, has Bayesian and profile likelihood capabilities, and is designed specifically for fitting these types of models.

## LaHave River

For the LaHave River, the stock-recruit function was fit to the egg deposition and smolt production data using maximum likelihood (with a log-normal error distribution) in S-Plus.

### 4.1.3 Results

## St. Mary's River

In setting up the model, a decision was required on how best to treat the habitat parameter, $h$. For the St. Mary's River, total habitat area (i.e., the amount of habitat with a gradient greater than 0.12 \% ( $3,985,400 \mathrm{~m}^{2}$, Amiro 1993)) had been estimated based on measurements from orthophoto maps (aerial photographs corrected for bias from which actual distances can be measured), as well as from in-stream survey measurements (3,078,500 m²; MacEachern 1954 cited in Marshall 1986). The latter value has been used in previous assessments for calculating the conservation requirement for the St. Mary's River (Amiro et al. 2006). Calculating $h$ from the in-stream habitat measurements as compared to the orthophoto map measurements leads to a significant reduction of the O.F.V. function, indicating a better fit of the life-history model to the data. Additionally, the biological plausibility of the resulting estimate for parr mortality ( $M_{\text {Parr }}$ ) and the habitat carrying capacity for age-1 parr ( $R_{\text {asy }}$ ) were improved. Therefore, $h$ is described in the results as $1,693,200 \mathrm{~m}^{2}(55 \%$ of $3,078,500)$.

In total, 5 parameters were estimated from the freshwater production model ( $M_{\text {Egg }}, M_{\text {Parr }}, \alpha, R_{\text {asy }}, j_{2}$ ), and the model was fit to data from each of the 3 time periods: 19842008, 1993-2008 and 1997-2008. Parameter estimates obtained from the freshwater production model (Table 16) were biologically plausible (e.g., mortality estimates between 0 and 1 ) and the standard errors of the estimates were not large relative to the means, indicating reasonable model fits. The observed electrofishing data series and the predicted densities of parr (Figure 25) both show declining trends for each time period and in all juvenile age classes.

Scatterplots of the abundance of Atlantic salmon within a cohort in sequential age classes (Figure 26) illustrate the asymptotic behaviour (characteristic of density dependence) for age-1 at relatively low densities of both age-0 and age-1 fish. Estimated relationships appear to fit the data reasonably well. The predicted sizes of the smolt run from the model intersect the data (Figure 27), although residuals are greater than those for the younger age classes. This was expected given that process error would accumulate as fish get older.

It is interesting to note that the most pronounced difference among the 3 time periods modeled was the change in carrying capacity for age-1 parr. Estimated carrying capacity for age-1 parr was 47.1 fish $/ 100 \mathrm{~m}^{2}$ of habitat based on the 1984-2008 time period, which dropped to 17.2 for data from the 1993-2008 time period, and to 10.8 for the 1997-2008 time period (Table 16). The standard deviations of all parameter estimates tended to overlap, which is an indication that changes may not be statistically significant among time periods. However, such a pronounced change in freshwater carrying capacity is suggestive of a decline in freshwater habitat quality over time, and warrants further investigation to determine if the change is real.

Given that the predicted abundances of each age class were similar to the observed data series, it is likely that the life-history model is accurately predicting life-history parameter values. Therefore, it was possible to use the predicted smolt abundance series and the observed 1SW adult returns to calculate return rates to the St. Mary's River over time. These were calculated for each of the 3 time periods (1984-2008, 1993-2008 and 1997-2008) and were compared with return rates observed on the LaHave River from 1996 to 2007 (Figure 28). The 3 series of return rates predicted for the St. Mary's track each other closely and, in general, follow the trends observed in smolt return rates to the LaHave River (i.e., increasing or decreasing in the same year). The mean, maximum and minimum return rates predicted for the 1997-2008 time period were chosen as the input series for the marine component of the equilibrium model (see Section 4.2.1). Using this data series had two advantages: (1) information on the age distribution and spawning history of returning adults could be used to estimate the 2SW return rate, and (2) the level of uncertainty in the abundance estimates for all age classes was higher in the earlier years of data collection (1984-1997), which would have increased the uncertainty in any parameters estimated from these data.

## LaHave River

Two parameter values were estimated from the LaHave freshwater production model: the maximum rate of smolt production ( $\breve{\alpha}=0.032$ ) and the carrying capacity of the river for smolt ( $\breve{R}_{\text {asy }}=25,001$ ). These values represent current freshwater conditions in the LaHave River given that smolt aging data (necessary to calculate total abundance from a single egg cohort) were only available from 1996-2008.

### 4.2 Egg-per-smolt Model and Equilibrium Modeling

### 4.2.1 Egg-per-smolt Model

The egg-per-smolt model includes terms for the probability of maturing at sea-age-1, sizespecific fecundity, and post-spawning survival (survival between spawning events). The number of eggs produced by a smolt throughout its life (EPS) is given by:

$$
E P S=\sum_{c=1}^{2} E g g_{c},
$$

where

$$
\begin{gathered}
E g g_{1}=\left(1-M_{\text {Sea }}\right)\left(m_{1}\right)\left(f_{1}+\sum_{1}^{\max p}\left(1-M_{\text {Adult }}\right)^{p} f_{2}\right), \\
E g g_{2}=\left(1-M_{\text {Sea }}\right)^{c}\left(1-m_{1}\right) \sum_{0}^{\max p}\left(1-M_{\text {Adult }}\right)^{p} f_{2}
\end{gathered}
$$

where $c$ is the number of years spent at sea prior to maturity, $M_{\text {Sea }}$ is the annual mortality rate of immature salmon at sea, $m_{1}$ is the probability of maturing after one winter at sea, $f_{c}$ is either the fecundity of first-time spawners $\left(f_{1}\right)$ or older salmon $\left(f_{2}\right), p$ is the number of previous spawnings, and $M_{\text {Adult }}$ is the adult mortality rate. As written above, all repeat spawning is sequential and a maximum of 3 sequential spawning events is assumed per individual (based on observed data for this population). The parameter $M_{\text {Sea }}$ is a composite parameter that includes all sources of mortality in the marine environment, and the parameter $M_{\text {Adult }}$ includes mortality associated with overwintering in fresh water as well as with post-spawning downstream migration.

To describe survival at sea and maturity, the survival rate for immature salmon is assumed to be the same during the first and second years at sea and the model is written in terms of a survival and a maturity parameter, calculated from the observed proportions of 1SW and 2SW salmon returning each year. For the St. Mary's River, data on the ages of returning adults exist from 1997-2008, corresponding to years when seining took place. Using the predicted smolt abundance series for 1997-2008 from the life-history model, the mean, minimum and maximum return rates from 1997 to 2007 were converted into survival and maturity parameters for the population. For the LaHave River, return rates from 1996-2007 could be calculated from observed data exclusively (given that smolts had been monitored at the fishway since 1996). Similar to the St. Mary's, the mean, minimum and maximum return rates were converted to survival and maturity parameters for the population (Table 17). For each river, these values represent the range of variation in marine conditions experienced by the population in recent years.

To estimate $M_{\text {Adult }}$, a cohort analysis on the spawning history of adults sampled in the St. Mary's or LaHave rivers was used to calculate the survival rate of repeat spawners. Data from the previous 5 years (2002-2007) were used to ensure that the estimates represent current conditions. For the St. Mary's River, the mean percentage of adults that returned to spawn in consecutive years was $3.7 \%$, leading to an adult mortality estimate of $96.3 \%$. For the LaHave River, the mean percentage of adults returning to spawn in consecutive years was $5.5 \%$, giving an adult mortality estimate of $94.5 \%$.

### 4.2.2 Equilibrium Calculations

Equilibrium modeling splits the life cycle of a species into two (or more) parts and determines the population size at which the productivity of both halves of the lifecycle are balanced (Gibson and Myers 2003). The equilibrium point is the size that the population will tend towards if the life-history parameter values used as model inputs do not change.

Equilibrium models are ideal for assessing the recovery potential of diadromous fish like Atlantic salmon because their life cycle is naturally split into two parts (at the egg and smolt life stages, roughly equivalent to freshwater production and marine growth and survival). The freshwater
production model (smolt-per-egg model) gives the relationship between egg deposition and survival through to the smolt life stage. The marine production model (egg-per-smolt model) gives the rate at which smolts produce eggs throughout their lives. The equilibrium population size occurs where the rate at which eggs produce smolts equals the inverse of the rate at which smolts produce eggs.

For the St. Mary's and LaHave rivers, the parameter values used in the smolt-per-egg model (freshwater production component) were derived from the life-history model described above (Section 4.1). Parameter values for the egg-per-smolt model (EPS) are derived from observed data using the equations below, following the approach outlined in Gibson et al. (2009).

The number of eggs produced by the number of smolts $(S)$ in year $t$ is:

$$
E g g_{t}=E P S \cdot S_{t}
$$

Equilibrium numbers of eggs and recruitment levels (denoted with asterisks) are found by solving this equation for $S$, and substituting the result in the freshwater production model (Quinn and Deriso 1999):

$$
\frac{E g g^{*}}{E P S}=\frac{\breve{\alpha} E g g^{*}}{1+\frac{\widetilde{\alpha} E g g^{*}}{\widetilde{R}_{\text {asy }}}} .
$$

Note that here, $\breve{\alpha}$ and $\breve{R}_{\text {asy }}$ have been rescaled for the St. Mary's River to represent the maximum survival rate from egg to smolt and the asymptotic recruitment level for smolt (these parameters were estimated directly for the LaHave River). The equilibrium spawning biomass (Egg*) is then:

$$
E g g^{*}=\frac{(\breve{\alpha} E P S-1) \breve{R}_{\text {asy }}}{\breve{\alpha}}
$$

and the equilibrium number of smolts $\left(S^{*}\right)$ is found by substituting Egg* into the freshwater production model:

$$
S^{*}=\frac{\breve{\alpha} E g g^{*}}{1+\frac{\breve{\alpha} E g g^{*}}{\bar{R}_{\text {asy }}}} .
$$

### 4.2.3 Results

## St. Mary's River

Under recent freshwater conditions in the St. Mary's River (1997-2008), the maximum number of smolts produced per egg ( $\breve{\alpha}$ ) was estimated to be 0.044 with an asymptotic population size ( $\breve{R}_{\text {asy }}$ ) of 61,360 smolts. This latter estimate equates to approximately 3.6 smolts per $100 \mathrm{~m}^{2}$ of habitat. Over 3 generations (1993-2008), the estimated maximum number of smolts produced per egg is slightly lower, at 0.039 (owing to a lower value for $\alpha$ for age-1 parr), but the asymptotic smolt population size is substantially higher, at $123,863\left(7.3\right.$ smolts $/ 100 \mathrm{~m}^{2}$ of habitat). For the longest time period (1984-2008), the maximum number of smolts produced per egg is even lower, at 0.03 , but the asymptotic smolt population size is 234,038 smolts
( 13.8 smolts $/ 100 \mathrm{~m}^{2}$ of habitat). The latter estimate of smolt production is very high and is unlikely to be representative of the St. Mary's River. High-grading during electrofishing site selection during the earlier time period could bias this estimate upwards. However, the other 2 estimates are closer to rates of smolt production that have been observed in rivers in the past (Symons 1979), and are more likely to represent freshwater production potential in the St. Mary's River.

At the mean return rates predicted for 1SW and 2SW adults on the St. Mary's River, lifetime egg production per smolt was 36.1 eggs. At the minimum return rates, this value dropped to 13.1 eggs, while at the maximum return rates the value increased to 64.9 eggs. Given the freshwater production estimates above, one smolt would have to produce between 22 and 33 eggs (the inverse of 0.044 and 0.03 smolts produced per egg, respectively) throughout its life to ensure an equilibrium population size greater than zero. As such, the population could decline to extinction in the absence of human intervention due to natural variation in return rates, particularly if there is any further downturn in marine survival or freshwater productivity. Furthermore, the equilibrium model for the St. Mary's River predicts a very low maximum lifetime reproductive rate for the population. Multiplying the maximum rate of population increase for smolt ( $\breve{\alpha}$ ) by the number eggs produced per smolt (at the mean return rates for adults) gives a maximum lifetime reproductive rate of 1.59 .

For the 1997-2008 freshwater data series, small equilibrium population sizes are predicted when return rates equal the mean or maximum predicted values (Figure 29; intersection of lines (b) and (C), or lines (c) and (C), respectively). However, the predicted population sizes are well below $50 \%$ of the conservation requirement for the river, suggesting that the population is likely to remain small. Even at the two higher levels of freshwater production predicted from the 19932008 and 1984-2008 data series, there is little difference in equilibrium population size at the minimum or mean return rates (Figure 29, intersection of lines (a) and origin, or lines (b) and (A), (B) or (C)). However, the predicted equilibrium increases from approximately 3 million eggs, to 5 million, to 7 million when return rates to the St. Mary's are high and freshwater production increases (Figure 29, intersection of lines (c) and (C), (B) and (A), respectively). This suggests that any improvement to freshwater habitat (i.e., through restoration activities) could promote population growth and help restore salmon abundance in the St. Mary's River, but a decrease in marine mortality is necessary to restore populations to levels above the conservation requirement. None of the 3 modeled scenarios predict population sizes in excess of the conservation requirement, which suggests that population recovery to levels above the conservation requirement on the St. Mary's River will be difficult to achieve even with a change in at-sea survival. However, increases in freshwater productivity should increase population viability, albeit at lower population sizes.

## LaHave River

Under recent freshwater conditions in the LaHave River (1996-2008), the maximum number of smolts produced per egg ( $\breve{\alpha}$ ) was estimated to be 0.032 with an asymptotic population size ( $\breve{R}_{\text {asy }}$ ) of 25,001 smolts. This latter estimate equates to approximately 0.96 smolts per $100 \mathrm{~m}^{2}$ of habitat, based on the estimated $2,605,200 \mathrm{~m}^{2}$ of juvenile habitat available in the watershed above Morgans Falls. Compared to the St. Mary's River, this estimate of smolt production is very low and could be indicative of relatively poor freshwater productivity in the LaHave River. However, it could partially result from variability in the observed smolt and egg data series. The 2005 smolt cohort (2002 egg deposition year) was extremely small, potentially due to irregular water flow patterns in the LaHave River prior to the smolt migration. If this data point is removed
from the series, $\breve{\alpha}=0.012$ and $R_{\text {asy }}=92,389$, which increases the smolt production estimate to 3.5 smolts per $100 \mathrm{~m}^{2}$ of habitat. Given that the 2005 smolt abundance estimate is low relative to both the 2004 parr density and the number of 2006 1SW returns, these later values are likely more representative of current freshwater productivity in the LaHave River above Morgans Falls.

At the mean return rates observed for 1SW and 2SW adults on the LaHave River, lifetime egg production per smolt was 78.1 eggs. At the minimum return rates, this value decreased to 32.7 eggs, while at the maximum return rates, it increased to 206.5 eggs. Given the freshwater production estimate above, one smolt would have to produce 31.3 eggs (the inverse of 0.032 smolts per egg) throughout its life to ensure an equilibrium population size greater than zero. At the lowest observed return rates, the maximum lifetime reproductive rate for salmon in the LaHave River is estimated to be 1.05 . Values less than 1 indicate that a population cannot replace itself and is in decline, a result collaborated by the survival index calculated for the LaHave River in Section 3. At the average observed return rates, the maximum lifetime reproductive rate is 2.50 . This value is still low relative to many fish populations, but does indicate a greater resiliency to environmental fluctuations than was estimated for salmon in the St. Mary's River.

At the mean return rates observed on the LaHave River, the predicted equilibrium population size is approximately $55 \%$ of the conservation requirement if all observed smolt data are used (Figure 30, intersection of lines (b) and (B)) and is approximately $25 \%$ of the conservation requirement if the 2005 smolt cohort year is not included (Figure 30, intersection of lines (b) and (A)). An extremely small equilibrium population is predicted at the minimum return rates, but only under the lower freshwater production scenario (Figure 30, intersection of lines (a) and (B)). However, an equilibrium well in excess of the conservation requirement is predicted at the maximum return rates for both levels of freshwater production (Figure 30, intersection of lines (c) and (A), or (c) and (B)). This suggests that the population has the potential to increase or decrease in size depending on the level of mortality experienced in the marine environment. Any decline in marine mortality should lead to population recovery even in the absence of changes to freshwater productivity.

### 5.0 THREATS TO POPULATIONS

Threats to Atlantic salmon populations include any factors or activities that contribute to their decline or limit their recovery, and can affect the population at any life stage in the marine or freshwater environments. One issue known to be affecting salmon populations throughout the Southern Upland is reduced survival in the marine environment, although the factors leading to the mortality are poorly understood (Amiro et al. 2008, Gibson et al. 2008). Current mortality rates in the marine environment (as estimated in Section 4) are $96 \%$ for first-time spawners and 95\% for repeat spawners returning to the St. Mary's River, and $92 \%$ for first-time spawners and $95 \%$ for repeat spawners returning to the LaHave River. Although these mortality rates are not as extreme as those affecting populations in the inner Bay of Fundy (Gibson et al. 2008), the equilibrium analysis in Section 4 demonstrates the limiting nature of marine mortality on both current population size and future population recovery.

In the freshwater environment, sulfate deposition in the form of acid rain has lowered the pH of many rivers in the Southern Upland to the point that they may no longer be able to support viable salmon populations. The last region-wide assessment of pH was completed in 1986, where 22 rivers in the Southern Upland were classified as low- or non-acidified ( $\mathrm{pH}>5.0$ ),

20 rivers were partially acidified ( pH ranges from 4.7-5.0) and 14 rivers were heavily acidified ( $\mathrm{pH}<4.7$ ) (Watt 1987, Amiro et al. 2000). At a pH below 5.1, salmon production is considered unstable and only remnant populations may persist (LaCroix 1985). Estimates of the loss in productive potential attributable to acidification for salmon throughout the Southern Upland range from $24 \%$ (Amiro 2000) to 50\% (Watt 1989), with the majority of the impact taking place prior to 1986. Despite reductions in acid precipitation in recent years, the pH in rivers of the Southern Upland has not recovered at rates observed in other geographic areas (Watt 1987). Based on the electrofishing surveys done in 2000 and 2008, both the proportion of rivers in which juvenile salmon are present and the estimated densities of juvenile salmon are decreasing. It is likely that the low pH of many rivers is a contributing factor to these observed declines. The loss of productivity related to acidification would likely exacerbate the negative effects of low marine survival and further increase a population's vulnerability to extirpation (Amiro 2000).

Other factors with the potential to impact salmon abundance or distribution in the Southern Upland have been reviewed in a semi-quantitative manner in a recent draft conservation status report (DFO and MNRF 2009). The broad categories of threats include: directed salmon fishing, by-catch, fisheries on prey species, municipal water use, habitat alterations, aquaculture and other fish culture/stocking, military activities, scientific research, air pollutants (acid rain), ecotourism, invasive species, and ecosystem change. Impacts on salmon populations are measured in two ways: (1) the proportion of salmon populations that are likely to be influenced by a given activity, and (2) the population-level impact of a given activity on spawner abundance.

Fishing activities (including directed fisheries, incidental by-catch and fisheries on prey species) have a high cumulative effect on salmon populations given that they are thought to impact more than $30 \%$ of salmon populations throughout the Southern Upland. However, the overall impact on spawner abundance is low or uncertain. Multiple steps have been taken to reduce the impact of fisheries on salmon populations, such as permitting only catch-and-release angling, negotiating fishing agreements with Aboriginal communities and closing domestic retention fisheries. Furthermore, the impacts of fisheries on marine or freshwater habitat quality and quantity are thought to be low. Reported by-catch, illegal retention, and interception by distant commercial fisheries are low. However, any removal of pre-spawning adult salmon could have significant population-level impacts given the small size of populations throughout the Southern Upland.

Habitat alteration and water extraction, through activities such as mining, forestry, agriculture, infrastructure development and maintenance, municipal water use, and hydroelectric power generation, impacts more than $30 \%$ of salmon populations in the Southern Upland. However, the impact in terms of spawner loss is typically uncertain. Given that these activities take place predominantly in freshwater and estuarine environments, they have the potential to impact all life stages of Atlantic salmon. Furthermore, development activities that alter habitat and change the hydrology of rivers have no known positive effects on salmon populations. As such, spawner loss could be substantial.

In general, activities in the marine environment affect the majority of salmon populations in the Southern Upland, yet their impact on spawner abundance is relatively low or uncertain. This result is surprising given current marine mortality rates (> 90\%), and it suggests that further research on the cumulative nature of multiple threats leading to changes in spawner abundance is necessary. Similarly, the only factor in the freshwater environment that is known to have a high impact on spawner abundance is acid precipitation. However, recent declines in the predicted carrying capacity of the St. Mary's River (a non-acidified river) and the low production
potential of the LaHave River (a low-acidified river) suggest that factors other than acidification are reducing salmon production in fresh water throughout the Southern Upland.

### 6.0 DISCUSSION

Overall, the available data on salmon in rivers throughout the Southern Upland demonstrate that some populations are presently extirpated and that the healthiest populations are persisting at low abundance levels. This conclusion is consistent for all monitored life stages of the index populations and for the region-wide assessments of adults and juveniles. The estimated abundances of age-0, age-1, and age-2 parr and smolts are well below reference values for salmon production in freshwater, and adult abundance remains well below the conservation requirements established for the St. Mary's and LaHave rivers. The predicted decline rates for adult escapement over the last two or three generations for populations in the Southern Upland indicate declines in excess of $50 \%$ for the index rivers and in excess of $95 \%$ for other populations. Life-history modeling indicates that freshwater carrying capacity is low on both index rivers and is potentially decreasing in the St. Mary's River. Furthermore, equilibrium modeling demonstrates the limiting nature of high marine mortality and low freshwater productivity on abundance and recovery potential for populations in the Southern Upland.

Recent region-wide surveys indicate very low salmon abundance in the majority of rivers in the Southern Upland. Recreational catches in 2007 were extremely low, with $85 \%$ of the catch concentrated in two of the larger river systems: the St. Mary's and the LaHave. Fisheryindependent data on juvenile abundance and distribution corroborate the results of the recreational catch data, indicating extremely low juvenile density in the majority of rivers in the Southern Upland. No Atlantic salmon juveniles were observed in 31 of 50 rivers, and the mean density in a river system did not exceed 10.8 fry per $100 \mathrm{~m}^{2}$ or 7.5 parr per $100 \mathrm{~m}^{2}$, with the majority of values being much lower. The current distribution of salmon in the Southern Upland appears to be increasingly restricted, and actual abundance within those systems with salmon is extremely low.

Although populations in the index rivers are thought to be large relative to those in the other rivers in the Southern Upland, current monitoring indicates that abundance in the index rivers is low relative to the conservation requirement and other indicators. On the St. Mary's River, adult escapement is less than $25 \%$ of the conservation requirement and smolt production is an order of magnitude lower than that expected from a healthy salmon population. The LaHave River attained $55 \%$ of the adjusted conservation requirement in 2008, but the total count of adult salmon was one of the lowest values observed in the past 3 decades. Similarly, freshwater production in the LaHave was low (as indicated by smolt production as well as the egg-to-smolt survival rate in 2008), and the survival index suggests that the population is not able to replace itself under current conditions (i.e., it is gradually decreasing in size).

At a regional level, the recreational catch data and electrofishing surveys demonstrate the continued decline in abundance and restriction in range of Atlantic salmon populations throughout the Southern Upland. For the majority of rivers, reported recreational catch has declined steadily through time with a slight lag in the reduction in fishing effort. From 1983 to 2007, catches on most rivers dropped by more than $95 \%$, and many rivers were closed to angling in 1998 due to concerns over abundance. Similarly, when the results from the 2008 electrofishing survey were compared with those from 2000, both the estimated density of each life stage and the number of rivers in which salmon were found were significantly lower. Only $54 \%$ of rivers in the Southern Upland were found to contain salmon in 2000 (Amiro et al. 2000) and only $38 \%$ were found to contain salmon in 2008. Such a dramatic change in the estimated
density and distribution of juvenile salmon suggests that population extirpations are occurring throughout the Southern Upland and that abundance is critically low in many rivers.

The decline rates predicted over the previous three generations for salmon populations in the Liscomb River and East River, Sheet Harbour meet the criteria outlined by COSEWIC for a population designation of 'Endangered.' The salmon populations in the St. Mary's River and LaHave River meet the criteria for a designation of 'Threatened,' provided the cause of the decline is known and has ceased, or a designation of 'Endangered' if the cause of the decline is unknown and on-going. In other rivers throughout the Southern Upland, the trends in recreational catch over time and juvenile salmon abundance suggest that population declines have been as severe, if not greater (i.e., leading to extirpation), for salmon in these rivers. Although acidification has been identified as a major threat to salmon populations throughout the Southern Upland (Watt 1987), the effects have not been entirely mitigated in any river, although a liming station does exist in the West River, Sheet Harbour (SFA 20) that raises the pH of approximately $30 \%$ of the total watershed area. Other significant stressors to salmon populations in the Southern Upland include high marine mortality, which is also affecting populations throughout the inner Bay of Fundy (Amiro et al. 2008, Gibson et al. 2008, Trzcinski et al. 2004); hydroelectric power generation; and potentially reduced freshwater habitat quantity or quality (as evidenced by the decline in $R_{\text {asy }}$ over time on the St. Mary's River).

When all data sources were combined into the equilibrium model, small viable populations were predicted for the St. Mary's and LaHave rivers under current conditions. However, at the minimum return rates of 1 SW and 2 SW adults, both populations have an equilibrium size that is near zero, which suggests that natural variation in life-history parameters (particularly survival at sea) could drive the population to extinction in the absence of human intervention or a change in population vital rates. Given the low maximum lifetime reproductive rates, population growth back to the (low) equilibrium size is expected to be slow if random events further depress abundance. If marine mortality were to decline, current freshwater productivity should allow abundance to increase in excess of the conservation requirement on the LaHave River above Morgans Falls. This situation differs from that in the St. Mary's River, where increases in freshwater productivity and marine survival would be necessary before the population could significantly increase in size. Replacement rates calculated for the LaHave and St. Mary's river salmon indicate that the populations have low resiliency to environmental variability, thereby increasing their overall risk of extinction. Populations with maximum lifetime reproductive rates close to one have limited ability to increase in size following any change in abundance, while populations with rates less than one are not able to increase in size. It is expected that populations in surrounding rivers would be less likely to be viable under current conditions, given that the salmon populations in the LaHave and St. Mary's rivers are the largest.

While not fully explored here, there are major differences in the dynamics of salmon populations in the Southern Upland and inner Bay of Fundy. Within the inner Bay, smolt-to-adult return rates are roughly $0.3 \%$ (Gibson et al. 2008), or an order of magnitude lower than those reported here for Southern Upland populations. The marine survival rates observed for inner Bay of Fundy populations are so low that they cannot be offset by increased freshwater productivity. In the absence of an increase in survival at sea, populations in that region are expected to extirpate in the absence of human intervention, such as live gene banking. In contrast, the equilibrium analyses and threats information herein indicate that freshwater production for many Southern Upland populations is depressed, and that increasing freshwater productivity can increase the viability of populations in this region (although an increase in survival at sea will be required to restore populations to levels above their conservation requirement). Given the well documented effects of captive rearing on the genetic integrity of fish populations (Fraser 2008), an emphasis
on restoration of freshwater habitat or mitigation of identified threats in the marine environment should be preferred over live gene banking (see description below) as a mechanism to maintain these populations.

All indicators suggest that there is a strong likelihood for continuing decline in the abundance and distribution of Atlantic salmon in the Southern Upland. There are very few remaining populations with any appreciable abundance and of these, none are close to the conservation requirement for the river. The number of rivers with sufficient abundance of salmon remaining for recovery to occur is likely small. Within the inner Bay of Fundy, a key rescue activity was the establishment of a Live Gene Bank (O'Reilly and Doyle 2007), the goal of which was to protect remaining genetic diversity within remnant populations. Given that small viable populations are predicted at current life-history parameter values for populations in the Southern Upland, recovery actions aimed at increasing freshwater production potential or marine survival are potentially alternatives to the establishment of a Live Gene Bank. However, if current trends in the Southern Upland continue, any actions will have to be timely to be effective.

### 6.0 ACKNOWLEDGEMENTS

The data and analyses presented here were made possible only through the efforts of many individuals and groups. For monitoring completed in 2008, we thank the crew leaders from the electrofishing survey (J. Whitelaw, E. Jefferson, J. Bryan, and J. Brazner) working with DFO. In addition, the Nova Scotia Department of Agriculture and Fisheries (NSDoAF), Nova Scotia Power Inc., and the Bluenose Coastal Action Foundation (BCAF) shared their juvenile monitoring data. The annual smolt and adult monitoring programs on the St. Mary's River would not be possible without collaboration with the St. Mary's River Association (SMRA). We would also like to thank the numerous individuals who have assisted with juvenile, smolt or adult monitoring programs in the Southern Upland over the last 40 years. Additionally, we thank Carolyn Harvie and David Hardie for comments that improved this manuscript.

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### 8.0 TABLES

Table 1. Adult escapement estimates based on mark-recapture seining experiments on the West Branch of the St. Mary's River from 1997 to 2008. Estimates from years where the mark-recapture was not completed are shown in bold type, and were calculated from the average seining catchability from 1997 to 2001.

| Year | Escapement Coefficient |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Marks | Captures | Recaptures | Estimate | of Variat | Catchability |
| 1997 | 67 | 117 | 8 | 892 | 30.39 | 0.075 |
| 1998 | 152 | 268 | 37 | 1083 | 14.84 | 0.140 |
| 1999 | 38 | 82 | 8 | 360 | 29.86 | 0.106 |
| 2000 | 76 | 191 | 43 | 336 | 13.09 | 0.226 |
| 2001 | 41 | 52 | 5 | 371 | 35.59 | 0.111 |
| 2002 | 31 |  |  | 236 |  |  |
| 2003 | 95 | 4 | 3 | 722* | 20.00 | 0.754 |
| 2004 | 64 |  |  | 486 |  |  |
| 2005 | 26 |  |  | 198 |  |  |
| 2006 | 142 | 50 | 30 | 240 | 11.07 | 0.592 |
| 2007 | 112 | 107 | 59 | 203 | 8.54 | 0.551 |
| 2008 | 30 | 63 | 4 | 397 | 39.2 | 0.076 |

[^0]Table 2. Age, spawning history, and fork length of adult salmon seined from the West Branch of the St. Mary's River in 2008. The 'Age' designation gives the sea-age of salmon, followed by the age of the fish at previous spawning events (sp).

| Age | Number |  | Length (cm) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Mean |  | Maximum |  | Minimum |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 1 | 52 | 49 | 56.7 | 54.8 | 63.3 | 61.7 | 50.0 | 50.5 |
| 3 sp 1 |  | 1 |  |  |  | 73.5 |  | 73.5 |
| 2 |  | 8 |  | 74.4 |  | 79.1 |  | 71.2 |
| 3 sp 2 |  | 1 |  |  |  | 82.0 |  | 82.0 |

Table 3. The estimated annual wild smolt production (90\% C.I.) and smolt wheel efficiency on the West Branch of the St. Mary's River during 2005 to 2008.

| Year | Wheel <br> Efficiency | Abundance <br> Estimate | $90 \% \mathrm{Cl}$ | Production <br> per unit area <br> $\left(\right.$ smolts $/ 100 \mathrm{~m}^{2}$ ) |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | $0.103^{* *}$ | 7350 | 6000 | 9100 | 0.43 |
| 2006 | 0.028 | 25100 | 18700 | 40300 | 1.48 |
| 2007 | 0.054 | 16110 | 12735 | 20835 | 0.95 |
| 2008 | 0.031 | 15217 | 9451 | 24154 | 0.90 |

** two wheels were deployed side-by-side.
Approximately $55 \%$ of the juvenile habitat is thought to be contained in the West Branch.

Table 4. Summary of the electrofishing sites surveyed on the St. Mary's River in 2008, including catch and estimated density for the three age classes of juvenile salmon for the East and West branch.


## Counts at the mark run (M)

Total count at the capture run (C)
Numbers of recaptures in the capture run (R)
Numbers of mortalities (Mort)

* estimates obtained using mean age-1 efficiency from mark-recapture sites in 2007 and 2008 (0.386)

Table 5. Estimated escapement of adult Atlantic salmon relative to the conservation requirement in the West Branch of the St. Mary's River for the years 1995 to 2008.

| Year | 1SW | MSW | \% Egg <br> Conservation |
| :---: | :---: | :---: | :---: |
| 1995 | 1121 | 240 | 78 |
| 1996 | 844 | 325 | 67 |
| 1997 | 390 | 61 | 26 |
| 1998 | 1059 | 41 | 63 |
| 1999 | 307 | 83 | 22 |
| 2000 | 315 | 25 | 20 |
| 2001 | 319 | 106 | 24 |
| 2002 | 220 | 16 | 14 |
| 2003 | 600 | 122 | 42 |
| 2004 | 464 | 23 | 28 |
| 2005 | 192 | 8 | 12 |
| 2006 | 222 | 18 | 14 |
| 2007 | 182 | 23 | 12 |
| 2008 | 361 | 36 | 23 |

Table 6. Age, spawning history, and size composition of the adult salmon captured at Morgans Falls fishway on the LaHave River in 2008. Age is divided into years in freshwater (fresh) and years at sea (sea) as well as sea-age at previous spawnings (s1 or s2).

| Age |  |  |  | Fork Length (cm) |  |  |  |  | Weight (kg) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fresh | Sea | s1 | s2 | Number | Mean | Min. | Max. | Std. | Number | Mean | Min. | Max. | Std. |
|  | 1 |  |  | 6 | 55.4 | 53.2 | 58.4 | 1.7 | 6 | 1.9 | 1.6 | 2.0 | 0.2 |
| 2 | 1 |  |  | 449 | 54.9 | 48.6 | 62.5 | 2.1 | 449 | 2.0 | 1.2 | 3.2 | 0.3 |
| 3 | 1 |  |  | 135 | 55.7 | 48.0 | 62.0 | 2.3 | 135 | 2.1 | 1.2 | 2.6 | 0.3 |
|  | 2 |  |  | 2 | 71.8 | 70.8 | 72.8 | 1.0 | 2 | 4.6 | 4.3 | 4.9 | 0.3 |
| 2 | 2 |  |  | 80 | 72.1 | 65.6 | 79.8 | 2.5 | 80 | 5.0 | 3.4 | 7.0 | 0.7 |
| 3 | 2 |  |  | 5 | 73.5 | 68.8 | 76.0 | 2.5 | 5 | 5.3 | 4.4 | 5.8 | 0.5 |
| 2 | 3 | 1 |  | 7 | 72.2 | 69.3 | 74.5 | 1.8 | 7 | 5.1 | 4.0 | 6.2 | 0.7 |

Table 7. The estimated production (90\% C.I.), density and return rate of wild smolts above Morgans Falls on the LaHave River during 1996 to 2008.

| Wild smolts |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | per | Ret | rate |
| Year | Estimate | $100 \mathrm{~m}^{2}$ | 1SW | 2SW |
| 1996 | $\begin{gathered} 20,511 \\ (19,886-21,086) \end{gathered}$ | 0.79 | 1.47\% | 0.23\% |
| 1997 | $\begin{gathered} 16,550 \\ (16,000-17,100) \end{gathered}$ | 0.63 | 4.33\% | 0.43\% |
| 1998 | $\begin{gathered} 15,600 \\ (14,675-16,600) \end{gathered}$ | 0.60 | 2.04\% | 0.34\% |
| 1999 | $\begin{gathered} 10,420 \\ (9,760-11,060) \end{gathered}$ | 0.40 | 4.82\% | 0.86\% |
| 2000 | $\begin{gathered} 16,300 \\ (15,950-16,700) \end{gathered}$ | 0.63 | 1.16\% | 0.11\% |
| 2001 | $\begin{gathered} 15,700 \\ (15,230-16,070) \end{gathered}$ | 0.60 | 2.70\% | 0.59\% |
| 2002 | $\begin{gathered} 11,860 \\ (11,510-12,210) \end{gathered}$ | 0.46 | 1.95\% | 0.45\% |
| 2003 | $\begin{gathered} 17,845 \\ (8,821-26,870) \end{gathered}$ | 0.68 | 1.75\% | 0.17\% |
| 2004 | $\begin{gathered} 20,613 \\ (19,613-21,513) \end{gathered}$ | 0.79 | 1.13\% | 0.33\% |
| 2005 | $\begin{gathered} 5,270 \\ (4,670-5,920) \end{gathered}$ | 0.20 | 7.95\% | 0.54\% |
| 2006 | $\begin{gathered} 22,971 \\ (20,166-26,271) \end{gathered}$ | 0.88 | 1.48\% | 0.40\% |
| 2007 | $\begin{gathered} 24,430 \\ (23,000-28,460) \end{gathered}$ | 0.98 | 2.33\% |  |
| 2008 | $\begin{gathered} 14,450 \\ (13,500-15,500) \end{gathered}$ | 0.55 |  |  |

Table 8. Summary of the electrofishing sites surveyed on the LaHave River in 2008, including catch and estimated density for the three age classes of juvenile salmon for sites above and below Morgans Falls.


Counts at the mark run (M)
Total count at the capture run (C)
Numbers of recaptures in the capture run (R)
Numbers of mortalities (Mort)

* estimates obtained using mean age-1 efficiency from mark-recapture sites in 2007 and 2008 (0.214)

Table 9. Summary of the densities (number per $100 \mathrm{~m}^{2}$ ) of age-0 and age-1 and older Atlantic salmon estimated by electrofishing in Nova Scotia's Southern Upland rivers during 2008. An asterisk (*) denotes rivers where at least one site used a mean area value (rather than measured area) to calculate overall density. $N$ is the number of electrofishing sites.

| River | N | Age-0 |  |  |  |  | Age-1 and older |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | mean | std. dev. | min | max | median | mean | std. dev. | min | max | median |
| Annapolis | 7 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.31 | 0.45 | 0.00 | 1.18 | 0.00 |
| Annis | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Bear | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Belliveau | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Blacks Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Chegoggin | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Clyde | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| East (Chester) | 3 | 0.26 | 0.45 | 0.00 | 0.78 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| East (Lockport) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| East (St Margarets) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| East Brk (Porter's Lake) | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Ecum Secum | 4 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.40 | 4.81 | 0.00 | 9.62 | 0.00 |
| Gaspereau Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Gegogan Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Gold * | 7 | 1.23 | 2.01 | 0.00 | 4.57 | 0.00 | 2.19 | 2.88 | 0.00 | 6.20 | 0.00 |
| Granite Village Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Halfway Brk | 1 | 5.03 | NA | 5.03 | 5.03 | 5.03 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Indian | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Ingram | 4 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Jordan | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| LaHave * | 9 | 2.93 | 2.59 | 0.00 | 5.87 | 1.76 | 2.73 | 2.84 | 0.00 | 9.14 | 2.57 |
| Little West | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Martin's | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Medway * | 4 | 2.71 | 3.74 | 0.00 | 8.16 | 1.34 | 1.42 | 1.75 | 0.00 | 3.59 | 1.05 |
| Mersey | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Middle (Chester) | 2 | 0.29 | 0.05 | 0.25 | 0.32 | 0.29 | 1.85 | 0.12 | 1.77 | 1.94 | 1.85 |
| Mosher | 3 | 0.95 | 1.65 | 0.00 | 2.85 | 0.00 | 1.01 | 1.07 | 0.00 | 2.14 | 0.90 |
| Mushamush | 4 | 0.12 | 0.24 | 0.00 | 0.47 | 0.00 | 0.48 | 0.35 | 0.00 | 0.75 | 0.58 |
| Musquodoboit | 4 | 10.27 | 12.85 | 0.00 | 28.04 | 6.53 | 7.45 | 6.69 | 0.00 | 16.16 | 6.82 |
| Nine Mile | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Petite | 3 | 0.30 | 0.27 | 0.00 | 0.50 | 0.41 | 0.14 | 0.23 | 0.00 | 0.41 | 0.00 |
| Purney Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Quoddy | 4 | 0.17 | 0.35 | 0.00 | 0.69 | 0.00 | 0.29 | 0.58 | 0.00 | 1.15 | 0.00 |
| Rodney Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Roseway | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Round Hill | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Sable | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Digby) | 3 | 0.33 | 0.57 | 0.00 | 0.99 | 0.00 | 0.33 | 0.57 | 0.00 | 0.99 | 0.00 |
| Salmon (Halifax) | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Lake Major) | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Lawrencetown) | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Port Dufferin) | 2 | 1.26 | 1.79 | 0.00 | 2.53 | 1.26 | 0.14 | 0.20 | 0.00 | 0.28 | 0.14 |
| Ship Harbour | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 4.17 | NA | 4.17 | 4.17 | 4.17 |
| Smith Brk | 1 | 4.81 | NA | 4.81 | 4.81 | 4.81 | 0.44 | NA | 0.44 | 0.44 | 0.44 |
| St. Mary's | 12 | 5.33 | 4.04 | 0.00 | 11.80 | 4.11 | 1.67 | 1.22 | 0.00 | 3.35 | 1.48 |
| Tangier | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tidney | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Tusket | 8 | 0.04 | 0.12 | 0.00 | 0.35 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| West Brk (Porter's Lake) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| West River Sheet Harbour * | 7 | 0.05 | 0.12 | 0.00 | 0.33 | 0.00 | 0.04 | 0.11 | 0.00 | 0.30 | 0.00 |
| West Taylor Bay | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 10. Summary of the densities (number per $100 \mathrm{~m}^{2}$ ) of age-0 and age-1 and older Atlantic salmon estimated by electrofishing in Nova Scotia's Southern Upland rivers during 2000. An asterisk (*) denotes sites where a mean area value (rather than measured area) was used to calculate overall density. $N$ is the number of electrofishing sites.

| River | N | Age-0 |  |  |  |  | Age-1 and older |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | mean | std. dev. | min | max | median | mean | std. dev. | min | max | median |
| Annis | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Argyle | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Belliveau | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Chezzetcook | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Clyde | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Country Harbour | 2 | 15.97 | 9.36 | 9.35 | 22.59 | 15.97 | 10.32 | 0.28 | 10.12 | 10.51 | 10.32 |
| East (Chester) | 2 | 2.85 | 3.95 | 0.06 | 5.65 | 2.85 | 3.59 | 4.84 | 0.17 | 7.01 | 3.59 |
| East (Lockport) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| East (St Margarets) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| East Brk (Porter's Lake) | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| East Taylor Bay | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Ecum Secum | 1 | 0.61 | NA | 0.61 | 0.61 | 0.61 | 9.55 | NA | 9.55 | 9.55 | 9.55 |
| Gaspereau Brk | 2 | 0.63 | 0.89 | 0.00 | 1.25 | 0.63 | 2.23 | 3.15 | 0.00 | 4.45 | 2.23 |
| Gegogan Brk | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Gold | 2 | 5.79 | 5.03 | 2.23 | 9.35 | 5.79 | 5.89 | 0.29 | 5.69 | 6.09 | 5.89 |
| Indian | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Indian Harbour Lakes | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.46 | 0.80 | 0.00 | 1.38 | 0.00 |
| Ingram | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Issac's Harbour | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.62 | 0.88 | 0.00 | 1.25 | 0.62 |
| Jordan | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Kirby | 2 | 18.17 | 3.67 | 15.58 | 20.77 | 18.17 | 17.80 | 3.15 | 15.58 | 20.03 | 17.80 |
| Lahave * | 18 | 5.25 | 5.69 | 0.00 | 19.47 | 3.86 | 8.95 | 7.00 | 0.00 | 20.02 | 9.96 |
| Liscombe | 1 | 0.50 | NA | 0.50 | 0.50 | 0.50 | 7.84 | NA | 7.84 | 7.84 | 7.84 |
| Little West | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Martin's * | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Medway | 3 | 5.35 | 6.65 | 0.00 | 12.80 | 3.25 | 2.03 | 1.94 | 0.00 | 3.86 | 2.23 |
| Mersey | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Middle | 2 | 1.40 | 1.98 | 0.00 | 2.80 | 1.40 | 2.32 | 0.90 | 1.69 | 2.96 | 2.32 |
| Mosher | 2 | 6.25 | 8.46 | 0.27 | 12.24 | 6.25 | 6.01 | 5.66 | 2.00 | 10.01 | 6.01 |
| Mushamush * | 2 | 33.66 | 0.49 | 33.32 | 34.01 | 33.66 | 6.90 | 2.00 | 5.49 | 8.31 | 6.90 |
| Musquodoboit | 2 | 43.14 | 61.01 | 0.00 | 86.29 | 43.14 | 22.21 | 12.64 | 13.28 | 31.15 | 22.21 |
| New Harbour | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.13 | 0.22 | 0.00 | 0.39 | 0.00 |
| Nine Mile | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Petite * | 5 | 9.44 | 20.28 | 0.00 | 45.71 | 0.65 | 1.34 | 1.24 | 0.37 | 3.27 | 0.78 |
| Quoddy * | 3 | 0.05 | 0.08 | 0.00 | 0.14 | 0.00 | 1.35 | 0.79 | 0.73 | 2.23 | 1.08 |
| Round Hill | 2 | 0.44 | 0.62 | 0.00 | 0.88 | 0.44 | 2.01 | 0.86 | 1.40 | 2.62 | 2.01 |
| Sable | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Salmon (Digby) | 2 | 3.27 | 4.63 | 0.00 | 6.54 | 3.27 | 6.85 | 9.69 | 0.00 | 13.71 | 6.85 |
| Salmon (Halifax) | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Lake Major) | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmon (Lawrencetown) | 3 | 0.02 | 0.03 | 0.00 | 0.06 | 0.00 | 0.02 | 0.03 | 0.00 | 0.06 | 0.00 |
| Salmon (Port Dufferin) | 2 | 0.60 | 0.85 | 0.00 | 1.21 | 0.60 | 0.40 | 0.57 | 0.00 | 0.81 | 0.40 |
| Ship Harbour | 2 | 2.57 | 3.63 | 0.00 | 5.13 | 2.57 | 1.97 | 2.78 | 0.00 | 3.94 | 1.97 |
| Sissibo | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Smith Brk | 1 | 2.50 | NA | 2.50 | 2.50 | 2.50 | 10.01 | NA | 10.01 | 10.01 | 10.01 |
| St. Mary's River | 14 | 14.21 | 23.52 | 0.00 | 85.08 | 4.64 | 3.36 | 3.30 | 0.00 | 11.21 | 2.21 |
| Tangier * | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tidney | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| Tusket | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.61 | 0.06 | 2.57 | 2.66 | 2.61 |
| West Brk (Porter's Lake) | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |
| West River Sheet Harbour | 3 | 4.47 | 4.80 | 0.00 | 9.55 | 3.86 | 1.90 | 1.50 | 0.20 | 3.05 | 2.44 |
| West Taylor Bay | 1 | 0.00 | NA | 0.00 | 0.00 | 0.00 | 0.00 | NA | 0.00 | 0.00 | 0.00 |

Table 11. Total (wild + hatchery) escapement estimates for 1SW and MSW salmon in the four Southern Upland rivers on which adult monitoring has taken place. The values for LaHave above Morgans Falls; East River, Sheet Harbour; and the Liscomb River are based on total counts at a fishway. The size of the salmon observed on the East River, Sheet Harbour in 2008 was not recorded. The values for the St. Mary's River (total river) are based on recreational catch estimates from 1974 to 1996 and on markrecapture seining experiments on the West Branch (rescaled by habitat area) thereafter.

| Year | SFA 21 |  | SFA 20 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LaHave above M.F. |  | East Sheet Harbour |  | Liscomb |  | St. Mary's |  |
|  | 1SW | MSW | 1SW | MSW | 1SW | MSW | 1SW | MSW |
| 1970 | 2 | 4 | 31 |  |  |  |  |  |
| 1971 | 3 |  | 19 | 1 |  |  |  |  |
| 1972 | 17 | 2 | 111 |  |  |  |  |  |
| 1973 | 152 | 16 | 29 | 4 |  |  |  |  |
| 1974 | 471 | 21 | 87 |  |  |  | 3227 | 722 |
| 1975 | 504 | 73 | 89 | 4 |  |  | 558 | 242 |
| 1976 | 646 | 131 | 120 | 6 |  |  | 2763 | 427 |
| 1977 | 1266 | 109 | 83 | 1 |  |  | 2128 | 528 |
| 1978 | 842 | 276 | 13 | 3 |  |  | 1138 | 427 |
| 1979 | 1920 | 166 | 19 | 0 | 60 |  | 4889 | 290 |
| 1980 | 1973 | 777 | 53 | 6 | 111 |  | 12840 | 669 |
| 1981 | 3047 | 592 | 59 | 1 | 76 | 6 | 3549 | 1197 |
| 1982 | 1420 | 486 | 5 | 0 | 252 | 10 | 3213 | 268 |
| 1983 | 1156 | 313 | 59 | 3 | 520 | 15 | 4246 | 787 |
| 1984 | 2293 | 420 | 66 | 4 | 606 | 48 | 3453 | 767 |
| 1985 | 1445 | 715 | 26 | 1 | 507 | 87 | 3654 | 2853 |
| 1986 | 1724 | 662 | 9 | 2 | 736 | 117 | 7353 | 3147 |
| 1987 | 3102 | 611 | 46 | 4 | 1614 | 88 | 1744 | 1070 |
| 1988 | 3520 | 449 | 32 | 3 | 477 | 76 | 6675 | 2313 |
| 1989 | 2530 | 694 | 57 | 9 | 532 | 75 | 1715 | 1540 |
| 1990 | 2476 | 508 | 16 | 1 | 955 | 44 | 6010 | 913 |
| 1991 | 604 | 326 | 31 | 5 | 586 | 38 | 4998 | 880 |
| 1992 | 2489 | 273 | 22 | 4 | 145 | 27 | 1583 | 507 |
| 1993 | 1158 | 205 | 33 | 1 | 134 | 11 | 2697 | 1320 |
| 1994 | 848 | 247 | 17 | 2 | 134 | 10 | 462 | 100 |
| 1995 | 948 | 228 | 27 | 2 | 150 | 6 | 2038 | 437 |
| 1996 | 1130 | 196 | 11 | 1 | 85 | 9 | 1535 | 590 |
| 1997 | 449 | 131 | 4 | 1 | 27 | 1 | 709 | 110 |
| 1998 | 919 | 137 | 1 | 0 | 9 | 0 | 1926 | 74 |
| 1999 | 452 | 132 | 15 | 0 | 9 | 0 | 559 | 150 |
| 2000 | 794 | 120 | 1 | 0 |  |  | 572 | 46 |
| 2001 | 379 | 182 | 1 | 0 |  |  | 580 | 193 |
| 2002 | 1133 | 71 | 0 | 0 |  |  | 400 | 29 |
| 2003 | 437 | 207 | 1 | 0 |  |  | 1092 | 221 |
| 2004 | 638 | 122 | 1 | 0 |  |  | 843 | 41 |
| 2005 | 416 | 84 |  |  |  |  | 349 | 15 |
| 2006 | 425 | 115 |  |  |  |  | 404 | 32 |
| 2007 | 341 | 41 |  |  |  |  | 331 | 42 |
| 2008 | 593 | 98 | 3 total |  |  |  | 656 | 65 |

Table 12. Summary of declines in adult Atlantic salmon abundance (large and small size categories combined) for four rivers on Nova Scotia's Southern Upland. The regression method is a log-linear model fit via least squares. The step function is the change in the 5-year mean population size ending on the years given in the time period column (the number of years differs between the methods). The standard errors and $95 \%$ confidence intervals are in parentheses. Fifteen years correspond to about three generations. A negative value in the decline columns indicates an increasing population size. Model fits for the 15-year time period are shown in Figures 21-24.

| Fishing area | Population | Length of time series (years) | Time Period | Slope (SE) | Regression |  | Step Function |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | 1 Year decline rate (\%) | Decline over time period (\%) | Decline over time period (\%) |
| 20 | Liscomb | 10 | 1989-1999 | -0.458 (0.052) | 36.7 (29.9,42.8) | 99.3 (98.0,99.8) | 95.9 (87.3,98.7) |
| 20 | East River, SH | 15 | 1994-2008 | -0.243 (0.104) | 21.6 (3.9,36.0) | 97.4 (44.8,99.9) | 93.1 (86.1,96.3) |
| 20 | St. Mary's | 15 | 1994-2008 | -0.080 (0.032) | 7.6 (1.7,13.3) | 69.6 (22.1,88.2) | $81.8(54.2,92.7)$ |
| 21 | LaHave | 15 | 1994-2008 | -0.056 (0.018) | $5.4(2.0,8.7)$ | 56.6 (26.7,74.3) | 65.7 (43.4,78.9) |

Table 13. Mean juvenile density by age class on the West and East branches of the St. Mary's River as estimated from mark-recapture electrofishing surveys for the years 1990-2008. Surveys in years marked with asterisks were completed by the St. Mary's River Association and the ages of the juveniles captured were approximated from length-frequency information.

| Year | EAST |  |  |  | WEST |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | $0+$ parr | $\begin{aligned} & 1+ \\ & \text { parr } \\ & \hline \end{aligned}$ | $\begin{gathered} 2+ \\ \text { parr } \end{gathered}$ | N | $\begin{gathered} \hline 0+ \\ \text { parr } \\ \hline \end{gathered}$ | $\begin{aligned} & \hline 1+ \\ & \text { parr } \\ & \hline \end{aligned}$ | $\begin{gathered} 2+ \\ \text { parr } \end{gathered}$ |
| 1990* | 11 | 3.40 | 7.76 | 0.89 | 3 | 4.70 | 7.80 | 0.90 |
| 1991* | 9 | 3.41 | 5.39 | 0.56 | 5 | 25.80 | 4.20 | 0.40 |
| 1992* | 14 | 6.72 | 2.87 | 0.70 | 8 | 22.00 | 5.40 | 0.90 |
| 1993* | 3 | 34.65 | 7.71 | 0.56 | 3 | 143.70 | 10.20 | 0.60 |
| 1994* | 9 | 2.54 | 7.33 | 0.41 | 5 | 1.40 | 2.80 | 0.20 |
| 1995 | 11 | 19.99 | 4.13 | 1.00 | 4 | 16.60 | 2.61 | 0.36 |
| 1996 | 8 | 14.50 | 3.71 | 1.40 | 3 | 11.15 | 3.23 | 0.46 |
| 1997 | 7 | 32.67 | 3.01 | 0.36 | 8 | 25.22 | 10.44 | 0.80 |
| 1998 | 7 | 6.06 | 5.89 | 0.32 | 8 | 23.41 | 6.88 | 1.75 |
| 1999 | 7 | 14.29 | 1.68 | 1.18 | 8 | 12.37 | 3.44 | 1.53 |
| 2000 | 6 | 19.37 | 1.81 | 0.14 | 8 | 6.66 | 4.06 | 0.32 |
| 2001 | 4 | 24.02 | 9.51 | 0.60 | 5 | 5.91 | 5.43 | 0.71 |
| 2002 | 8 | 2.85 | 5.28 | 1.33 | 6 | 3.92 | 2.14 | 0.72 |
| 2003 | 6 | 4.85 | 2.23 | 2.58 | 6 | 4.23 | 5.27 | 0.48 |
| 2004 | 6 | 2.53 | 2.63 | 0.39 | 6 | 3.63 | 0.63 | 0.36 |
| 2005 | 5 | 13.98 | 5.23 | 1.18 | 4 | 7.72 | 5.58 | 0.87 |
| 2006 | 5 | 5.95 | 2.87 | 0.23 | 6 | 3.78 | 0.78 | 0.43 |
| 2007 | 6 | 17.06 | 6.25 | 0.24 | 7 | 4.02 | 2.51 | 0.06 |
| 2008 | 6 | 7.58 | 2.29 | 0.24 | 6 | 6.15 | 2.51 | 0.33 |

Table 14. Estimates of 1SW and MSW adult returns to the St. Mary's River for the years 1984 to 1996. Values are based on the reported recreational catch for the entire river and the mean catch rate for each size class observed on the LaHave River.

| Year | 1SW | MSW | Total |
| :---: | :---: | :---: | :---: |
| 1984 | 3453 | 767 | 4220 |
| 1985 | 3654 | 2853 | 6507 |
| 1986 | 7353 | 3147 | 10499 |
| 1987 | 1744 | 1070 | 2814 |
| 1988 | 6675 | 2313 | 8989 |
| 1989 | 1715 | 1540 | 3255 |
| 1990 | 6010 | 913 | 6924 |
| 1991 | 4998 | 880 | 5878 |
| 1992 | 1583 | 507 | 2090 |
| 1993 | 2697 | 1320 | 4017 |
| 1994 | 462 | 100 | 562 |
| 1995 | 2038 | 437 | 2475 |
| 1996 | 1535 | 590 | 2125 |

Table 15. Estimated egg deposition (thousands) by Atlantic salmon above Morgans Falls on the LaHave River, based on annual counts at Morgans Falls fishway, the biological characteristics of the returning population, and the length-fecundity relationship calculated for female salmon from the LaHave.

| No. of eggs ('000s) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Wild | Hatchery | Toposition | Prop. <br> wild |
| 1973 | 50 | 87 | 137 | 0.36 |
| 1974 | 25 | 372 | 397 | 0.06 |
| 1975 | 91 | 501 | 592 | 0.15 |
| 1976 | 190 | 727 | 917 | 0.21 |
| 1977 | 396 | 1,086 | 1,482 | 0.27 |
| 1978 | 452 | 1,367 | 1,819 | 0.25 |
| 1979 | 1,292 | 1,284 | 2,576 | 0.50 |
| 1980 | 2,698 | 1,680 | 4,378 | 0.62 |
| 1981 | 3,263 | 1,641 | 4,904 | 0.67 |
| 1982 | 1,683 | 1,779 | 3,462 | 0.49 |
| 1983 | 1,968 | 335 | 2,303 | 0.85 |
| 1984 | 3,059 | 248 | 3,307 | 0.93 |
| 1985 | 3,421 | 413 | 3,834 | 0.89 |
| 1986 | 4,079 | 499 | 4,578 | 0.89 |
| 1987 | 4,899 | 720 | 5,619 | 0.87 |
| 1988 | 4,381 | 958 | 5,339 | 0.82 |
| 1989 | 4,315 | 1,024 | 5,339 | 0.81 |
| 1990 | 3,414 | 652 | 4,066 | 0.84 |
| 1991 | 1,354 | 376 | 1,730 | 0.78 |
| 1992 | 2,867 | 508 | 3,375 | 0.85 |
| 1993 | 1,140 | 522 | 1,662 | 0.69 |
| 1994 | 1,177 | 455 | 1,632 | 0.72 |
| 1995 | 926 | 446 | 1,372 | 0.67 |
| 1996 | 1,085 | 519 | 1,604 | 0.68 |
| 1997 | 507 | 440 | 946 | 0.54 |
| 1998 | 903 | 431 | 1,334 | 0.68 |
| 1999 | 717 | 359 | 1,076 | 0.67 |
| 2000 | 926 | 499 | 1,425 | 0.65 |
| 2001 | 829 | 785 | 1,614 | 0.51 |
| 2002 | 870 | 972 | 1,842 | 0.47 |
| 2003 | 877 | 1068 | 1,945 | 0.45 |
| 2004 | 1,027 | 926 | 1,953 | 0.53 |
| 2005 | 628 | 515 | 1,143 | 0.55 |
| 2006 | 915 | 216 | 1,131 | 0.81 |
| 2007 | 540 | 20 | 561 | 0.96 |
| 2008 | 1,078 | 0 | 1,078 | 1 |
|  |  |  |  |  |

Table 16. Estimates of freshwater production parameters from the life-history model for the St. Mary's River for the three time periods of data presented.

|  | 1997-2008 |  | 1993-2008 |  | 1984-2008 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Value | Standard <br> dev. | Value | Standard <br> dev. | Value | Standard <br> dev. |
| $M_{\text {Egg }}$ | 0.880 | 0.022 | 0.872 | 0.020 | 0.904 | 0.013 |
| $\alpha$ | 0.723 | 0.241 | 0.606 | 0.145 | 0.499 | 0.123 |
| $R_{\text {asy }}$ | 10.787 | 7.266 | 17.243 | 9.851 | 47.088 | 90.703 |
| $M_{\text {Parr }}$ | 0.504 | 0.107 | 0.460 | 0.113 | 0.402 | 0.122 |
| $j_{2}$ | 0.796 | 0.038 | 0.800 | 0.037 | 0.810 | 0.035 |
|  |  |  |  |  |  |  |
| $\widetilde{\alpha}$ | 0.044 |  | 0.039 |  | 0.030 |  |
| $\widetilde{R}_{\text {asy }}$ | 61,360 |  | 123,863 |  | 234,038 |  |

Table 17. Minimum, mean and maximum return rates for the St. Mary's River and LaHave River Atlantic salmon populations and the calculated probabilities of mortality at sea ( $M_{\text {Sea }}$ ) and maturation after 1SW $\left(m_{1}\right)$ used in the equilibrium analysis.

|  |  | Return rate (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Population | 1 SW | $2 S W$ |  | $M_{\text {Sea }}$ | $m_{1}$ |
| St. Mary's River | Minimum | 0.61 | 0.004 | 0.990 | 0.596 |
|  | Mean | 1.44 | 0.09 | 0.962 | 0.373 |
|  | Maximum | 2.25 | 0.26 | 0.936 | 0.353 |
|  | Minimum | 1.13 | 0.11 | 0.961 | 0.564 |
| LaHave River | Mean | 2.78 | 0.40 | 0.921 | 0.351 |
|  | Maximum | 7.95 | 0.86 | 0.859 | 0.287 |

### 9.0 FIGURES



Figure 1. Map showing the location of the Southern Upland region relative to the three other Atlantic salmon management regions in the Maritimes Region.


Number of sites and year
Figure 2. Mean density for the three age classes of juvenile salmon (age-0, age-1, and age-2+) in the St. Mary's River during 1985-1986 and 1990-2008. The number of sampling sites on which the mean is based is listed immediately below the $x$-axis.


Figure 3. Observed fry density as a function of estimated adult returns to the St. Mary's River for the return years 1992 to 2007. The linear equation for the predicted relationship (thick line), as well as the associated $R^{2}$ value for the regression are given.


Figure 4. Counts of Atlantic salmon at Morgans Falls fishway on the LaHave River, NS, from 1974 to 2008, divided into the proportions of wild-origin and hatchery-origin 1SW and MSW adults.


Number of sites and year
Figure 5. Total mean parr density (age-1 and age-2) per $100 \mathrm{~m}^{2}$ as determined by electrofishing in the LaHave River for the years 1979-1984, 1987, 1988, 1990-1995, and 1997-2008. The number of sampling sites each year is listed immediately below the $x$-axis.


Figure 6. Estimated egg deposition (1000s) relative to the conservation requirement by wild and hatchery Atlantic salmon above Morgans Falls from 1973-2008. No adults of hatchery origin contributed to egg deposition in 2008.


Figure 7. Percent of the conservation requirement (1.96x10 eggs) attained annually above Morgans Falls on the LaHave River from 1990-2008.


Figure 8. Survival index (Ln Recruits/spawning salmon) of Atlantic salmon above Morgans Falls on the LaHave River for the spawning escapements in 1970-2003 and the associated returns of adults from 1974-2007.


Figure 9. Distribution of the sites visited during the electrofishing survey for Atlantic salmon juveniles throughout the Southern Upland in 2008.


Figure 10. Box plots showing the density of Atlantic salmon in Southern Upland rivers based on electrofishing during 2008. The black dot shows the median density and the box shows the inter-quartile spread. Hollow dots indicate zero salmon density at sites in that watershed. The whiskers are drawn to the minimum and maximum. " $N$ " is the number of sites that were electrofished in each river.


Figure 11. Distribution of the sites visited during the electrofishing survey for Atlantic salmon juveniles throughout the Southern Upland in 2000.
juvenile Atlantic salmon 2000


Figure 12. Box plots showing the density of Atlantic salmon in Southern Upland rivers based on electrofishing during 2000. The black dot shows the median density and the box shows the inter-quartile spread. Hollow dots indicate zero density at sites on that watershed. The whiskers are drawn to the minimum and maximum. " $N$ " is the number of sites that were electrofished in each river.


Figure 13. A comparison of the mean juvenile densities (all age classes combined) in watersheds throughout the Southern Upland in 2000 and 2008. Watersheds in which no salmon were captured are shown in black.


Figure 14. Comparison of densities of juvenile salmon (all age categories) at sites that were electrofished in 2000 and again in 2008. The line is the one-to-one line, above which densities have increased from 2000 to 2008 and below which they have decreased.


Figure 15. Comparison of densities of juvenile salmon fry at sites that were electrofished in 2000 and again in 2008. The line is the one-to-one line, above which densities have increased from 2000 to 2008 and below which they have decreased.


Figure 16. Comparison of densities of juvenile salmon parr at sites that were electrofished in 2000 and again in 2008. The line is the one-to-one line, above which densities have increased from 2000 to 2008 and below which they have decreased.


Figure 17. The percent change in reported recreational catch for all rivers in SFA 20, where the mean catch in three time periods was compared with the mean during 1983-1987. Rivers in which the decline in catch was $>95 \%$ or the increase was $>200 \%$ are labeled with the actual value. Missing points in the most recent time period represent rivers that have been closed to angling for the full 5 -year period.


Figure 18. The percent change in reported recreational fishing effort for all rivers in SFA 20, where mean effort in three time periods was compared with mean effort during 1983-1987. Rivers in which the decline in catch was $>95 \%$ or the increase was > $200 \%$ are labeled with the actual value. Missing points in the most recent time period represent rivers that have been closed to angling.


Figure 19. The percent change in reported recreational catch for all rivers in SFA 21, where the mean catch in three time periods was compared with the mean during 1983-1987. Rivers in which the decline in catch was $>95 \%$ or the increase was $>200 \%$ are labeled with the actual value. Missing points in the most recent time period represent rivers that have been closed to angling.


Figure 20. The percent change in reported recreational fishing effort for all rivers in SFA 21, where mean effort in three time periods was compared with mean effort during 1983-1987. Rivers in which the decline in catch was $>95 \%$ or the increase was > $200 \%$ are labeled with the actual value. Missing points in the most recent time period represent rivers that have been closed to angling.


Figure 21. Observed adult escapement from 1979 to 1999 (points) and predicted 10-year population declines from the ratio method (horizontal dashed lines) and log-linear model (solid line) for the Liscomb River.

## East River, Sheet Harbour Esc.



Figure 22. Observed adult escapement from 1970 to 2008 (points) and predicted 15-year population declines from the ratio method (horizontal dashed lines) and log-linear model (solid line) for East River, Sheet Harbour.


Figure 23. Observed adult escapement from 1970 to 2008 (points) and the predicted 15-year population declines from the ratio method (horizontal dashed lines) and log-linear model (solid line) for the LaHave River above Morgans Falls.

## St. Mary's Esc.



Figure 24. Observed adult escapement from 1974 to 2008 (points) and the predicted 15-year population declines from the ratio method (horizontal dashed lines) and log-linear model (solid line) for the St. Mary's River.


Figure 25. Observed (points) and predicted (lines) densities of juvenile Atlantic salmon from the freshwater production model for the West Branch of the St. Mary's River. Models were fit to data spanning three time periods (as labelled for each column).


Figure 26. Functional relationships between three ages of Atlantic salmon from the freshwater production model for the St. Mary's River. Models were fit to data spanning three time periods (as labelled for each column). The points are the observed data while the lines show the fitted relationship.


Figure 27. Observed (points) and predicted (lines) abundance of Atlantic salmon smolt and egg deposition from the freshwater production model for the St. Mary's River. Models were fit to data spanning three time periods (as labelled for each column).


Figure 28. Comparison of the smolt-to-1SW return rates predicted from the life-history model for the St. Mary's River for three time periods: 1984-2008, 1993-2008 and 1997-2008. Return rates observed on the LaHave River from 1996-2007 are plotted for comparison.


Figure 29. Equilibrium analysis of the salmon population dynamics in the St. Mary's River. The points are the observed egg depositions and smolt production for the 2003 to 2005 cohort years. The curved, dashed lines represent freshwater production as calculated from data spanning 1984-2008 (A), 19932008 (B) and 1997-2008 (C). The straight, dashed lines represent marine production as calculated at the minimum observed return rates (a), the mean observed return rates (b) and the maximum observed return rates (c) for 1SW and 2SW adults. Dark shading indicates egg depositions above the conservation egg requirement, medium shading is between $50 \%$ and $100 \%$ of the egg requirement, and the light shading is below $50 \%$ of the requirement.


Figure 30. Equilibrium analysis of the salmon population dynamics in the LaHave River above Morgans Falls. The points are the observed egg depositions and smolt production for the 1994 to 2005 cohort years. The curved, solid line (B) represents current freshwater production, as calculated from smolt abundance and egg deposition data from 1996-2008. The curved, solid line (A) represents current freshwater production if the 2005 smolt cohort is not included in the analysis. The straight, dashed lines represent marine production as calculated at the minimum observed return rates (a), the mean observed return rates (b), and the maximum observed return rates (c) for 1SW and 2SW adults. Dark shading indicates egg depositions above the conservation egg requirement, medium shading is between $50 \%$ and $100 \%$ of the egg requirement, and the light shading is below $50 \%$ of the requirement.

### 10.0 APPENDICES

Appendix 1. Atlantic salmon parr (age-1 and age-2) densities from the annual electrofishing survey on the LaHave River, for sites above and below Morgans Falls during 1979 to 2008.

| Site | Year $\mathrm{N}$ | $\begin{array}{r} 1979 \\ 10 \end{array}$ | $\begin{array}{r} 1980 \\ 30 \end{array}$ | $\begin{array}{r} 1981 \\ 25 \end{array}$ | $\begin{array}{r} 1982 \\ 7 \end{array}$ | $\begin{array}{r} 1983 \\ 12 \end{array}$ | $\begin{array}{r} 1984 \\ 10 \end{array}$ | $\begin{array}{r} 1987 \\ 10 \end{array}$ | $\begin{array}{r} 1988 \\ 10 \end{array}$ | $\begin{array}{r} 1990 \\ 11 \end{array}$ | $\begin{array}{r} 1991 \\ 7 \end{array}$ | $\begin{array}{r} 1992 \\ 17 \end{array}$ | $\begin{array}{r} 1993 \\ 14 \end{array}$ | $\begin{array}{r} 1994 \\ 14 \end{array}$ | $\begin{array}{r} 1995 \\ 19 \end{array}$ | 1996 | $\begin{array}{r} 1997 \\ 14 \end{array}$ | $\begin{array}{r} 1998 \\ 12 \end{array}$ | $\begin{array}{r} 1999 \\ 13 \end{array}$ | $\begin{array}{r} 2000 \\ 16 \end{array}$ | $\begin{array}{r} 2001 \\ 16 \end{array}$ | $\begin{array}{r} 2002 \\ 16 \end{array}$ | $\begin{array}{r} 2003 \\ 12 \end{array}$ | $\begin{array}{r} 2004 \\ 16 \end{array}$ | $\begin{array}{r} 2005 \\ 16 \end{array}$ | $\begin{array}{r} 2006 \\ 11 \end{array}$ | $\begin{array}{r} 2007 \\ 11 \end{array}$ | $\begin{array}{r} 2008 \\ 8 \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Above Morgan Falls |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Misners \#8 |  |  | 3.70 | 5.80 | 8.99 | 12.34 | 5.97 | 8.62 | 6.75 | 9.50 | 9.60 | 20.40 | 6.50 | 8.90 | 8.10 |  | 12.40 | 9.10 | 18.00 | 14.30 | 9.60 | 15.30 |  | 5.60 | 15.46 | 21.95 | 2.61 | 1.00 |
| Cherryfield above bridge |  |  |  |  |  |  |  |  | 4.60 | 12.30 | 14.90 | 18.10 | 7.70 | 5.10 | 6.10 |  |  | 10.50 | 15.10 | 13.00 | 19.30 | 22.60 | 24.80 | 11.59 | 14.56 | 32.76 | 9.94 | 8.80 |
| Cherryfield \#10 |  | 4.51 | 7.78 | 6.05 |  | 8.59 |  | 7.50 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Falkland Ridge Bridge |  |  | 8.94 | 8.09 |  | 4.18 |  |  |  |  | 7.60 | 13.00 | 7.20 | 6.90 | 2.80 |  | 12.20 | 9.20 |  | 14.40 | 8.60 | 6.40 | 18.50 | 3.02 | 8.63 |  |  |  |
| Ohio R. site \#6 |  | 0.00 | 3.13 | 1.99 |  |  | 1.74 |  |  | 5.00 | 4.10 | 19.00 |  |  |  |  | 6.30 | 10.50 | 7.60 | 18.10 | 20.60 | 8.80 | 9.10 | 1.81 | 21.63 | 9.97 | 13.31 | 3.60 |
| \#7 |  | 0.00 | 0.00 | 0.37 |  |  |  |  |  |  |  |  |  |  |  |  | 1.60 | 3.40 |  | 0.40 | 0.10 | 5.20 |  | 7.58 | 3.08 |  |  |  |
| \#34 |  |  | 0.00 | 3.21 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#36 |  |  | 3.35 | 3.72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#37 |  |  | 1.32 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Ohio R. behind Conrad's |  |  |  |  |  |  |  |  |  |  | 6.30 | 11.00 |  |  |  |  | 6.40 |  |  | 11.60 | 10.10 | 4.40 |  | 4.13 | 13.66 | 9.13 | 0.00 |  |
| North R. \#14 |  | 2.78 | 7.04 | 4.91 | 4.37 | 7.25 | 4.40 | 9.24 | 10.63 | 24.70 | 29.70 | 31.20 | 15.40 | 13.10 | 17.00 |  |  | 22.30 | 27.40 | 18.80 | 26.10 | 17.10 |  | 16.30 | 18.00 | 42.49 | 11.12 |  |
| \#21 |  | 0.45 | 0.48 | 0.95 | 6.16 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#40 |  |  | 3.39 | 4.29 |  | 7.52 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North River (below Moosehorn) |  |  |  |  |  |  |  |  |  |  | 0.90 |  | 5.60 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Above Armstrong Lake |  |  |  |  |  |  |  |  | 3.22 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sixty Bk. \#26 |  | 0.28 | 2.46 | 7.04 | 20.54 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#41 |  |  | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#42 |  |  | 0.00 | 0.00 | 9.28 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#47 |  |  |  | 5.82 | 11.76 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mason Med Bk.\#38 |  |  | 0.00 | 9.46 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#39 |  |  | 0.00 | 0.59 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Yearly Averages |  | 1.34 | 2.60 | 3.66 | 10.18 | 7.98 | 4.04 | 8.45 | 6.30 | 12.88 | 10.44 | 18.78 | 8.48 | 8.50 | 8.50 |  | 7.78 | 10.83 | 17.03 | 12.94 | 13.49 | 11.40 | 17.47 | 7.15 | 13.57 | 23.26 | 7.40 | 4.47 |
| Below Morgan Falls |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Main river \#30 |  |  | 2.63 | 2.44 |  | 1.15 | 5.58 | 3.61 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#31 |  |  | 3.17 | 2.17 |  | 7.78 |  | 7.12 | 2.58 | 7.50 |  | 5.40 | 6.70 | 3.30 | 3.00 |  | 1.00 |  | 0.20 | 0.10 | 0.20 | 0.50 | 0.40 | 0.24 | 0.00 |  |  |  |
| \#32 |  |  | 7.45 | 1.18 |  | 1.23 | 4.07 | 15.91 | 1.58 | 0.80 |  | 10.10 | 5.10 | 1.60 | 2.30 |  | 4.10 |  |  | 1.80 | 0.10 | 0.00 | 0.10 | 0.23 | 0.00 |  |  |  |
| \#3 |  |  | 2.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| West Br. \#1 |  | 5.10 | 4.08 | 4.01 |  |  | 1.62 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#22 |  | 5.09 | 2.03 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#23 |  | 3.96 | 0.54 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| West Br. Wentzell Rd. |  |  |  |  |  |  |  |  |  | 8.20 |  | 12.70 | 16.00 | 10.90 | 7.40 |  | 9.00 | 14.40 | 12.80 | 11.50 | 13.80 | 7.00 | 5.90 | 5.42 | 10.64 | 14.91 | 0.00 | 10.60 |
| West Br. Frauzel Rd. |  |  |  |  |  |  |  |  |  | 15.00 |  | 13.60 | 12.20 | 8.70 | 9.70 |  | 6.60 | 13.90 | 19.20 | 4.70 | 9.40 | 5.70 | 16.80 | 7.43 | 10.03 | 7.49 | 13.28 | 26.30 |
| \#24 |  | 4.87 | 3.29 |  |  | 3.65 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| West Br. Fire Brook |  |  |  |  |  |  |  |  | 5.27 |  |  | 5.90 | 16.50 | 21.50 | 16.00 |  |  |  | 32.50 | 22.80 | 20.00 | 25.60 | 18.10 | 20.67 | 32.61 | 8.24 | 19.38 | 9.20 |
| North Br. \#16 |  |  | 4.75 | 6.00 |  |  | 3.93 | 4.99 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#19 |  |  | 0.88 |  |  | 3.28 |  | 11.86 | 2.65 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#20 |  |  |  |  |  |  | 1.30 |  |  |  |  |  | 1.60 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#43 |  |  | 3.61 |  |  |  |  | 2.43 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#44 |  |  | 7.66 |  |  | 3.17 | 2.46 | 4.62 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#45 |  |  | 0.95 | 1.28 | 1.60 | 22.73 | 1.03 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#46 |  |  | 2.49 | 5.93 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Br. above Fancy Pool (upper) |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 3.20 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Br. above Fancy Pool |  |  |  |  |  |  |  |  | 2.69 | 3.10 |  | 7.20 | 5.80 |  | 6.20 |  | 5.90 |  |  |  |  |  |  |  |  |  |  |  |
| North Br. above MacKay's Bridge |  |  |  |  |  |  |  |  | 3.15 | 6.40 |  | 12.40 |  | 11.90 | 3.90 |  | 12.30 | 10.10 | 7.90 | 5.20 | 4.10 | 1.50 | 2.40 | 3.88 | 8.56 |  |  |  |
| North Br. Holland's Cabin |  |  |  |  |  |  |  |  |  |  |  | 20.10 |  | 18.40 | 7.90 |  | 13.00 | 12.70 | 4.00 | 0.30 | 1.00 | 0.00 | 1.30 | 2.02 | 0.00 | 3.71 | 0.00 |  |
| North Br. Veinot's Campground |  |  |  |  |  |  |  |  |  |  |  | 17.50 | 11.40 | 31.80 | 14.30 |  | 16.40 | 12.50 | 43.30 | 10.60 | 11.80 | 16.20 | 19.10 | 13.51 | 18.40 | 23.93 | 23.93 | 9.60 |
| North Br. above Texas Lake |  |  |  |  |  |  |  |  |  | 13.30 |  | 11.20 | 7.50 | 13.30 | 12.80 |  | 20.60 | 10.40 | 10.30 | 11.20 | 5.10 | 3.90 | 8.80 | 13.68 | 18.52 | 9.85 | 0.00 |  |
| North Br., Lake Paul Brook |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5.30 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Br.,above County Line Lake |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4.60 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Br., Hardwood Lake Brook |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 6.50 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| North Br. below Sherbrooke Lake |  |  |  |  |  |  |  |  |  |  |  | 6.20 |  | 4.50 | 5.60 |  |  |  | 0.50 |  |  |  |  |  |  |  |  |  |
| Yearly Averages |  | 4.76 | 3.25 | 2.88 | 1.60 | 6.14 | 2.86 | 7.22 | 2.99 | 7.76 |  | 11.12 | 9.20 | 12.59 | 7.25 |  | 9.88 | 12.33 | 14.52 | 7.58 | 7.28 | 6.71 | 8.10 | 7.45 | 10.97 | 11.35 | 9.43 | 13.93 |

Appendix 2. Summary of the electrofishing survey in the Southern Upland during 2008. Site IDs correspond with the Diadromous Fish Division (DFD) electrofishing database. Catch is the number of fish captured on the first pass of the survey. Other species include yellow perch, mummichog, banded killifish, brown bullhead, chain pickerel, sea lamprey, gaspereau and alosa spp. Organizations were the DFO Diadromous Fish Division (DFD BIO), the Nova Scotia Department of Agriculture and Fisheries (NSDoAF) and the Bluenose Coastal Action Foundation (BCAF).

| River | Site ID | UTM |  | Organiz. | $\begin{gathered} \text { Area } \\ \left(\mathrm{m}^{2}\right) \\ \hline \end{gathered}$ | Mo. | Day | ShockingTime (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Atlantic salmon | American eel | Brook trout | Brown trout | White sucker | SMouth- bass | Chub spp. | Other Cyprinids | Stickleback spp. | Others |
| Annapolis | SU105 | 321913 | 4968332 | DFD BIO | 197 | 8 | 5 | 768 | 0 | 16 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| Annapolis | SU104 | 344908 | 4976408 | DFD BIO | 574 | 8 | 5 | 971 | 0 | 5 | 0 | 0 | 6 | 0 | 13 | 7 | 0 | 0 |
| Annapolis | SU109 | 309804 | 4960452 | DFD BIO | 381 | 8 | 7 | 773 | 1 | 12 | 0 | 0 | 0 | 0 | 5 | 2 | 0 | 0 |
| Annapolis | SU108 | 317421 | 4965759 | DFD BIO | 604 | 8 | 7 | 1011 | 1 | 15 | 2 | 0 | 7 | 0 | 6 | 0 | 0 | 1 |
| Annapolis | SU110 | 356468 | 4984811 | DFD BIO | 530 | 8 | 8 | 549 | 0 | 10 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 |
| Annapolis | SU111 | 359218 | 4983996 | DFD BIO | 666 | 8 | 8 | 952 | 0 | 12 | 1 | 0 | 3 | 0 | 4 | 0 | 0 | 0 |
| Annapolis | SU107 | 349917 | 4980092 | DFD BIO | 795 | 8 | 7 | 1087 | 4 | 15 | 0 | 0 | 3 | 0 | 21 | 1 | 0 | 0 |
| Annis | SU9C | 259496 | 4870901 | DFD BIO | 305 | 7 | 30 | 793 | 0 | 7 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 |
| Annis | SU9B | 259748 | 4870901 | DFD BIO | 603 | 7 | 31 | 1984 | 0 | 35 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 8 |
| Annis | SU9A | 259670 | 4869618 | DFD BIO | 772 | 7 | 31 | 1263 | 0 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bear | SU3B | 290609 | 4938330 | DFD BIO | 414 | 8 | 1 | 820 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Belliveau | SU5A | 256260 | 4917970 | DFD BIO | 174 | 7 | 29 | 494 | 0 | 10 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blacks Brk | SU102 | 313067 | 4849444 | DFD BIO | 934 | 8 | 1 | 743 | 0 | 8 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Chegoggin | SU106 | 247041 | 4862117 | DFD BIO | 234 | 8 | 6 | 525 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 8 |
| Clyde <br> East | SU13B | 296352 | 4850891 | DFD BIO | 733 | 7 | 30 | 836 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| (St Margarets) | SU31A | 431035 | 4948451 | DFD BIO | 4077 | 7 | 14 | 955 | 0 | 11 | 11 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| East (Chester) | SU27B | 409385 | 4944546 | DFD BIO | 301 | 7 | 23 | 1002 | 1 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| East (Chester) | SU27C | 408453 | 4942713 | DFD BIO | 320 | 7 | 15 | 545 | 0 | 1 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 1 |
| East (Chester) | SU27A | 407560 | 4938806 | DFD BIO | 498 | 7 | 15 | 1110 | 0 | 6 | 0 | 0 | 0 | 0 | 5 | 2 | 0 | 0 |
| East (Lockport) East Brook | SU16A | 327403 | 4845846 | DFD BIO | 515 | 7 | 30 | 767 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| (Porter's Lake) East Brook | SU38A | 470352 | 4963085 | DFD BIO | 560 | 7 | 9 | 703 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| (Porter's Lake) | SU38B | 470485 | 4963226 | DFD BIO | 405 | 7 | 9 | 556 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ecum Secum | SU54A | 559416 | 4992152 | DFD BIO | 267 | 8 | 26 | 1327 | 11 | 15 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ecum Secum | SU54B | 565491 | 4984766 | DFD BIO | 222 | 9 | 18 | 812 | 0 | 8 | 1 | 0 | 0 | 0 | 0 | 2 |  | 0 |
| Ecum Secum | SU54C | 565005 | 4984243 | DFD BIO | 201 | 9 | 18 | 821 | 0 | 4 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Ecum Secum | SU54D | 561250 | 4994977 | DFD BIO | 236 | 9 | 18 | 959 | 0 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gaspereau Brk | Gasb003 | 578938 | 4986969 | DFD BIO | 533 | 9 | 19 | 923 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |


|  |  | UTM |  | Organiz. | Area$\left(\mathrm{m}^{2}\right)$ | Mo. | Day | Shocking <br> Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Site ID | Easting | Northing |  |  |  |  |  | Atlantic salmon | American eel | $\begin{gathered} \text { Brook } \\ \text { trout } \end{gathered}$ | $\begin{gathered} \text { Brown } \\ \text { trout } \end{gathered}$ | White sucker | $\begin{aligned} & \text { SMouth- } \\ & \text { bass } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Others |
| Gegogan Brk | SU57A | 578807 | 4992603 | DFD BIO | 390 | 8 | 25 | 910 | 0 | 34 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Gold | Gold002 | 385312 | 4954982 | BCAF | 440 | 9 | 22 | 484 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gold | Gold003 | 385934 | 4956026 | BCAF | 670 | 8 | 22 | 905 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gold | Gold005 | 384532 | 4948356 | BCAF |  | 9 | 18 | 997 | 33 | 3 | 0 | 0 | 5 | 0 | 0 | 11 | 0 | 1 |
| Gold | Gold015 | 387104 | 4939677 | BCAF | 636 | 9 | 8 | 859 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Gold | Gold016 | 382814 | 4967805 | BCAF | 523 | 9 | 2 | 716 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 |
| Gold | Gold017 | 390284 | 4962345 | BCAF | 610 | 9 | 5 | 351 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |  | 0 | 2 |
| Gold <br> Granite Village | Gold018 | 385236 | 4954914 | BCAF | 711 | 8 | 29 | 696 | 11 | 1 | 7 | 0 | 0 | 0 | 0 | 4 | 0 | 0 |
| Brk | SU103 | 341129 | 4859242 | DFD BIO | 788 | 8 | 1 | 804 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Halfway Brk | Hafb003 | 543164 | 4972466 | DFD BIO | 186 | 9 | 19 | 749 | 4 | 13 |  | 0 | 0 | 0 | 3 | 0 | 0 | 0 |
| Indian | SU30A | 429462 | 4949492 | DFD BIO | 752 | 7 | 14 | 490 | 0 | 2 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ingram | SU29B | 423642 | 4948781 | DFD BIO | 441 | 7 | 15 | 1138 | 0 | 17 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ingram | SU29A | 422930 | 4949962 | DFD BIO | 643 | 7 | 15 | 1047 | 0 | 34 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ingram | SU29C | 422584 | 4952851 | DFD BIO | 303 | 9 | 5 | 946 | 0 | 13 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ingram | SU29D | 422798 | 4956021 | DFD BIO | 357 | 9 | 5 | 940 | 0 | 8 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Jordan | SU15B | 320531 | 4861372 | DFD BIO | 334 | 7 | 31 | 528 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Jordan | SU15C | 320441 | 4864508 | DFD BIO | 254 | 8 | 29 | 798 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| LaHave | LHav008 | 359406 | 4940137 | DFD BIO | 2975 | 9 | 2 | 2654 | 67 | 1 | 0 | 0 | 2 | 27 | 0 | 0 | 0 | 0 |
| LaHave | LHav101 | 356321 | 4943231 | DFD BIO | 1887 | 8 | 19 | 2013 | 69 | 0 | 0 | 0 | 3 | 0 | 11 | 0 | 0 | 0 |
| LaHave | LHav105 | 371358 | 4920720 | DFD BIO | 2529 | 8 | 22 | n/a | 37 | 1 | 0 | 0 | 2 | 0 | 3 | 0 | 0 | 0 |
| LaHave | LHav106 | 363843 | 4932713 | DFD BIO | n/a | 8 | 28 | 1996 | 17 | 3 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| LaHave | LHav107 | 366437 | 4919392 | DFD BIO | n/a | 8 | 28 | n/a | 46 | 2 | 0 | 0 | 1 | 0 | 3 | 0 | 0 | 0 |
| LaHave | LHav108 | 372650 | 4940027 | DFD BIO | 3195 | 8 | 21 | 2874 | 33 | 5 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 |
| LaHave | LHav114 | 358860 | 4930747 | DFD BIO | 637 | 8 | 22 | 1756 | 7 | 1 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 |
| LaHave | LHav001 | 369489 | 4920500 | DFD BIO | n/a | 8 | 28 | n/a | 25 | 3 | 1 | 0 | 2 | 0 | 3 | 0 | 0 | 0 |
| LaHave | LHav016 | 362400 | 4932490 | DFD BIO | 575 | 9 | 9 | n/a | 0 | 0 | 17 | 0 | 6 | 0 | 5 | 0 | 0 | 0 |
| LaHave | LHav106 | 363843 | 4932713 | DFD BIO | 440 | 8 | 28 | 1996 | 17 | 3 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| Little West | SU46B | 535940 | 4972807 | DFD BIO | 552 | 7 | 18 | 835 | 0 | 35 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Little West | SU46A | 534694 | 4972090 | DFD BIO | 312 | 7 | 18 | 405 | 0 | 13 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Martin's | SU24A | 393365 | 4927080 | DFD BIO | 988 | 8 | 8 | 1112 | 0 | 12 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| Martin's | SU24B | 392275 | 4927203 | DFD BIO | 571 | 9 | 4 | 1469 | 0 | 29 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Medway | Medw108 | 341858 | 4918886 | DFD BIO | 1227 | 9 | 10 |  | 22 | 6 | 3 | 0 | 0 | 0 | 1 | 0 | 0 | 23 |
| Medway | Medw109 | 367838 | 4892343 | DFD BIO | 716 | 9 | 11 | 1387 | 36 | 22 | 8 | 0 | 3 | 0 | 0 | 0 | 0 | 7 |
| Medway | SU20C | 351993 | 4902512 | DFD BIO | 1200 | 8 | 25 |  | 3 | 5 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| River | Site ID | UTM |  | Organiz. | $\begin{gathered} \text { Area } \\ \left(\mathrm{m}^{2}\right) \\ \hline \end{gathered}$ | Mo. | Day | Shocking Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Atlantic salmon | American $\qquad$ | $\begin{gathered} \text { Brook } \\ \text { trout } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Brown } \\ \text { trout } \\ \hline \end{gathered}$ | White sucker | $\begin{gathered} \text { SMouth- } \\ \text { bass } \end{gathered}$ | $\begin{aligned} & \text { Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Others |
| Medway | Medw101 | 332773 | 4922777 | DFD BIO | 300 | 9 | 11 |  | 0 | 2 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Mersey | SU19D | 352217 | 4883899 | DFD BIO | 838 | 7 | 24 | 1081 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mersey | SU19C | 355212 | 4882299 | DFD BIO | 849 | 7 | 24 | 780 | 0 | 35 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| Mersey | SU19E | 347046 | 4887243 | DFD BIO | 132 | 7 | 24 | 344 | 0 | 0 | 0 | 0 | 2 | 0 |  | 0 | 0 | 0 |
| Middle (Chester) | Midd001 | 398211 | 4936046 | BCAF | 926 | 9 | 10 | 933 | 8 | 0 | 10 | 0 | 2 | 0 |  | 5 | 0 | 1 |
| Middle (Chester) | Midd002 | 398676 | 4936162 | BCAF | 722 | 9 | 12 | 533 | 7 | 0 | 14 | 0 | 3 | 0 | 0 | 6 | 0 | 0 |
| Mosher | SU52C | 556590 | 4980877 | DFD BIO | 177 | 8 | 26 | 500 | 0 | 10 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mosher | SU52A | 556245 | 4982251 | DFD BIO | 519 | 8 | 26 | 1318 | 2 | 28 | 5 | 0 | 1 | 0 |  | 0 | 0 | 0 |
| Mosher | SU52D | 556552 | 4986000 | DFD BIO | 328 | 9 | 17 | 955 | 7 | 4 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| Mushamush | SU23A | 385698 | 4925430 | DFD BIO | 986 | 7 | 18 | 1514 | 5 | 17 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Mushamush | SU23B | 378077 | 4929624 | DFD BIO | 517 | 7 | 18 | 915 | 1 | 4 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Mushamush | SU23D | 381032 | 4931963 | DFD BIO | 310 | 9 | 12 | 892 | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mushamush | SU23C | 377311 | 4930183 | DFD BIO | 250 | 9 | 12 | 978 | 0 | 10 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 |
| Musquodoboit | SU40A | 491778 | 4990608 | DFD BIO | 145 | 8 | 11 | 703 | 17 | 1 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Musquodoboit | SU40B | 497596 | 4994368 | DFD BIO | 242 | 8 | 11 | 705 | 35 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Musquodoboit | SU40D | 493631 | 4993652 | DFD BIO | 214 | 9 | 10 | 820 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Musquodoboit | SU40C | 497566 | 4994541 | DFD BIO | 269 | 9 | 10 | 1023 | 11 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nine Mile | SU32A | 441988 | 4944810 | DFD BIO | 697 | 7 | 14 | 947 | 0 | 6 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 |
| Petite | SU21A | 383595 | 4899418 | DFD BIO | 665 | 7 | 25 | 1183 | 0 | 14 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Petite | SU21B | 378331 | 4907214 | DFD BIO | 574 | 7 | 25 | 978 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Petite | SU21C | 382175 | 4900622 | DFD BIO | 1399 | 7 | 25 | 797 | 3 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Purney Brk | SU100 | 318343 | 4850131 | DFD BIO | 541 | 7 | 31 | 679 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Quoddy | SU51B | 551817 | 4978810 | DFD BIO | 264 | 9 | 16 | 915 | 0 | 10 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Quoddy | SU51A | 551808 | 4980531 | DFD BIO | 337 | 9 | 16 | 982 | 1 | 25 | 0 | 0 | 1 | 0 | 8 | 0 | 0 | 0 |
| Quoddy | SU51D | 550845 | 4977212 | DFD BIO | 65 | 9 | 16 | 377 | 0 | 9 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Quoddy | SU51C | 551508 | 4975684 | DFD BIO | 203 | 9 | 17 | 877 | 1 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rodney Brk | SU101 | 318306 | 4847940 | DFD BIO | 521 | 7 | 31 | 916 | 0 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Roseway | SU112A | 310499 | 4858650 | DFD BIO | 243 | 8 | 28 | 417 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| Roseway | SU112B | 304917 | 4872963 | DFD BIO | 315 | 8 | 28 | 736 | 0 | 9 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| Roseway | SU112C | 302865 | 4878314 | DFD BIO | 125 | 8 | 28 | 562 | 0 | 0 | 19 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| Round Hill | SU2B | 309503 | 4956659 | DFD BIO | 542 | 7 | 29 | 1148 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Sable | SU17A | 333537 | 4856649 | DFD BIO | 1055 | 7 | 30 | 1060 | 0 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sable | SU17B | 333960 | 4856577 | DFD BIO | 295 | 7 | 31 |  | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Lake Major) | SU35A | 464095 | 4947750 | DFD BIO | 511 | 7 | 4 | 817 | 0 | 9 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| River | Site ID | UTM |  | Organiz. | $\begin{aligned} & \text { Area } \\ & \left(\mathrm{m}^{2}\right) \\ & \hline \end{aligned}$ | Mo. | Day | Shocking <br> Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Atlantic salmon | American eel | Brook trout | $\begin{aligned} & \text { Brown } \\ & \text { trout } \end{aligned}$ | White sucker | SMouthbass | $\begin{aligned} & \text { Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Others |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lake Major) | SU35B | 463950 | 4949263 | DFD BIO | 1912 | 7 | 10 | 1473 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lake Major) | SU35C | 463513 | 4949944 | DFD BIO | 2089 | 7 | 4 | 1364 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Halifax Co.) | SU41A | 496311 | 4964468 | DFD BIO | 233 | 7 | 16 | 363 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon <br> (Halifax Co.) | SU41B | 492615 | 4967838 | DFD BIO | 265 | 7 | 16 | 360 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Halifax Co.) | SU41C | 491486 | 4968638 | DFD BIO | 507 | 7 | 16 | 490 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36A | 469956 | 4948862 | DFD BIO | 722 | 7 | 8 | 957 | 0 | 27 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 6 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36B | 469669 | 4949135 | DFD BIO | 728 | 7 | 8 | 1311 | 0 | 30 | 0 | 0 | - 1 | 0 | 0 | 0 | 0 | 14 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36C | 469465 | 4954326 | DFD BIO | 1082 | 7 | 10 | 1131 | 0 | 40 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Digby) | SU8C | 254166 | 4887363 | DFD BIO | 213 | 7 | 31 | 607 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Digby) | SU8A | 248621 | 4883096 | DFD BIO | 465 | 8 | 6 | 538 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Digby) | SU8B | 252041 | 4882194 | DFD BIO | 1420 | 8 | 6 | 1553 | 12 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Port Dufferin) | SU50B | 548235 | 4979707 | DFD BIO | 263 | 7 | 23 | 545 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon <br> (Port Dufferin) |  | 547465 | 4977140 | DFD BIO | 833 | 7 | 28 | 1020 | 10 | 2 | 0 | 0 | 3 | 0 | 0 | 0 |  | 0 |
| Ship Harbour | SU42B | 504608 | 4967832 | DFD BIO | 448 | 7 | 16 | 1149 | 8 | 24 | 0 | 0 | 6 | 0 | 8 | 0 | 0 | 0 |
| Smith Brk | SU53A | 562330 | 4979886 | DFD BIO | 535 | 7 | 28 | 1016 | 12 | 12 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's | STMR854.2 | 577086 | 5013497 | DFD BIO | 718 | 9 | 16 | 1485 | 14 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's | STMR854.4 | 577040 | 5013648 | DFD BIO | 908 | 9 | 16 | 1809 | 49 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's | STMR855.1 | 561110 | 5013537 | DFD BIO | 626 | 8 | 26 | 1515 | 12 | 5 | 17 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| St. Mary's | STMR858.1 | 549950 | 5013416 | DFD BIO | 242 | 9 | 15 | 592 | 0 | 5 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's | STMR859.4 | 552876 | 5012910 | DFD BIO | 3172 | 9 | 17 | 4641 | 35 | 28 | 0 | 0 | 17 | 0 | 16 | 0 | 0 | - 1 |
| St. Mary's | STMR8510.8 | 553790 | 5030955 | DFD BIO | 914 | 8 | 26 | 1552 | 31 |  | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 |
| St. Mary's | STMR863.1 | 569912 | 5021222 | DFD BIO | 1145 | 9 | 26 | 1336 | 46 | 12 | 1 | 0 | 2 | 0 | 0 | 2 | 1 | 0 |
| St. Mary's | STMR867.1 | 552930 | 5032085 | DFD BIO | 648 | 8 | 26 | 1414 | 37 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - 2 |
| St. Mary's | STMR923 | 571938 | 5019086 | DFD BIO | 1121 | 9 | 26 | 1294 | 25 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| St. Mary's | STMR924 | 546607 | 5014243 | DFD BIO | 4600 | 9 | 25 | 5096 | 71 | 26 | 3 | 0 | 16 | 0 | 27 | 2 | 1 | 1 |
| St. Mary's | STMR925.1+2 | 555837 | 5014230 | DFD BIO | 574 | 9 | 15 | 1267 | 37 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| St. Mary's | STMR928 | 526196 | 5016130 | DFD BIO | 1247 | 9 | 15 | 2922 | 28 | 8 | 0 | 0 | 5 | 0 | 0 | 0 |  | 0 |
| Tangier | SU43C | 522667 | 4962362 | DFD BIO | 801 | 7 | 17 | 783 | 0 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tangier | SU43A | 514284 | 4978523 | DFD BIO | 509 | 7 | 17 | 809 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  |  | UTM |  | Organiz. | Area$\left(\mathrm{m}^{2}\right)$ | Mo. | Day | Shocking Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Site ID | Easting | Northing |  |  |  |  |  | Atlantic salmon | American $\qquad$ eel | $\begin{gathered} \text { Brook } \\ \text { trout } \end{gathered}$ | $\begin{gathered} \text { Brown } \\ \text { trout } \end{gathered}$ | White sucker | $\begin{aligned} & \text { SMouth- } \\ & \text { bass } \end{aligned}$ | $\begin{aligned} & \text { Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Others |
| Tangier | SU43B | 516961 | 4977079 | DFD BIO | 847 | 7 | 17 | 1211 | 0 | 37 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Tidney | SU18A | 337120 | 4859709 | DFD BIO | 902 | 7 | 30 | 727 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tusket | SU10A | 265735 | 4895220 | DFD BIO | 670 | 7 | 30 | 1627 | 1 | 32 | 0 | 0 | 4 | 3 | 0 | 0 | 0 | 3 |
| Tusket | SU10C | 262640 | 4867840 | DFD BIO | 259 | 8 | 5 | 665 | 0 | 11 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tusket | SU10B | 266239 | 4890674 | DFD BIO | 479 | 7 | 30 | 1243 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tusket | SU10E | 266646 | 4888596 | DFD BIO | 1224 | 8 | 6 | 1086 | 0 | 13 | 0 | 0 | 1 | 4 | 0 | 0 | 0 | 19 |
| Tusket | SU10D | 265850 | 4890632 | DFD BIO | 457 | 8 | 6 | 757 | 0 | 4 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 36 |
| Tusket | SU10F | 272318 | 4888521 | DFD BIO | 855 | 8 | 7 | 1199 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 |
| Tusket | SU10G | 287151 | 4888210 | DFD BIO | 306 | 8 | 7 | 957 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Tusket | SU10H | 274190 | 4884688 | DFD BIO | 1380 | 8 | 7 | 1046 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| West Brook <br> (Porter's Lake) | SU37A | 469518 | 4961793 | DFD BIO | 539 | 7 | 9 | 928 | 0 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| West Sheet <br> Harbour <br> West Sheet | WRSH001 | 515810 | 4992742 | NSDoAF |  | 7 | 30 | 1098 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Harbour West Sheet | WRSH002 | 523366 | 4980562 | NSDoAF | 611 | 7 | 30 | 528 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Harbour <br> West Sheet | WRSH003 | 529134 | 4979288 | NSDoAF | 784 | 8 | 19 | 1568 | 1 | 6 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 |
| Harbour <br> West Sheet | WRSH004 | 530096 | 4978469 | NSDoAF | 910 | 8 | 19 | 2371 | 0 | 7 | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 3 |
| Harbour West Sheet | WRSH005 | 523224 | 4990034 | NSDoAF |  | 8 | 19 | 1115 | 1 | 1 | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 0 |
| Harbour West Sheet | WRSH006 | 521871 | 4983915 | NSDoAF |  | 8 | 19 | 1386 | 0 | 3 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 |
| Harbour | WRSH007 | 518005 | 4986360 | NSDoAF | 826 | 8 | 20 | 1012 | 0 | 3 | 1 | 0 | 2 | 0 | 12 | 0 | 0 | 0 |
| West Taylor Bay | SU45A | 531239 | 4965654 | DFD BIO | 111 | 7 | 17 | 438 | 0 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| West Taylor Bay | SU45B | 529810 | 4965986 | DFD BIO | 390 | 7 | 17 | 738 | 0 | 9 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix 3. Summary of the Southern Upland electrofishing survey in 2000. Site IDs correspond with the Diadromous Fish Division (DFD) electrofishing database. Catch is the number of fish captured on the first pass of the survey. Other species include yellow perch, striped bass, banded killifish, sea lamprey and Alosa spp. Organizations were the DFO Diadromous Fish Division (DFD BIO), the Nova Scotia Department of Agriculture and Fisheries (NSDoAF) and the Bluenose Coastal Action Foundation (BCAF).

|  |  | UTM |  | Area$\left(\mathrm{m}^{2}\right)$ |  | Day | Shocking <br> Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Site ID | Easting | Northing Organiz. |  |  |  |  | Atlantic salmon | American eel | Brook trout | Brown Trout | White sucker | SMouth bass | Chub spp. | Other Cyprinids | Stickleback spp. | Other species |
| Annis | SU9A | 259678 | 4869480 DFD BIO | 900 | 9 | 20 | 320 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Annis | Su9B | 259744 | 4870867 DFD BIO | 900 | 9 | 20 | 650 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Argyle | SU11A | n/a | n/a DFD BIO | 420 | 10 | 5 | 350 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Belliveau | SU5A | 256445 | 4918260 DFD BIO | 100 | 9 | 12 | 382 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Chezzetcook | SU39A | 479941 | 4959980 DFD BIO | 1,500 | 9 | 7 | 509 | 0 | 17 | 19 | 8 | 2 | 0 | 0 | 0 | 0 | 0 |
| Clyde | SU13A | n/a | n/a DFD BIO | 750 | 9 | 22 | 320 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| Clyde | SU13B | 296283 | 4850920 DFD BIO | 600 |  |  | 581 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Country Harbour | SU61A | 586952 | 5013180 DFD BIO | 300 | 9 | 20 | 457 | 42 | 31 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Country Harbour | SU61B | 585771 | 5017097 DFD BIO | 200 | 9 | 20 | 370 | 17 | 23 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| East (Chester) | SU27A | 407490 | 4938745 DFD BIO | 4,125 | 9 | 14 | 2180 | 4 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| East (Chester) | SU27B | 409428 | 4944517 DFD BIO | 1,200 | 9 | 18 | 1619 | 65 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| East (Lockport) East (St | SU16A | 327515 | 4845781 DFD BIO | 300 | 9 | 22 | 380 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Margarets) <br> East Brk | SU31A | 430946 | 4948417 DFD BIO | 300 | 9 | 28 | 494 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| (Porter's Lake) <br> East Brk | SU38A | 470234 | 4962381 DFD BIO | 2,550 | 9 | 7 | 762 | 0 | 23 | 3 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| (Porter's Lake) | SU38B | 470465 | 4963168 DFD BIO | 240 | 9 | 7 | 265 | 0 | 14 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| East Taylor Bay | SU44A | 530138 | 4966196 DFD BIO | 150 | 9 | 13 | 442 | 0 | 27 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ecum Secum | SU54A |  | DFD BIO |  |  |  |  | 50 | 28 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Gaspereau Brk | SU56A | 578936 | 4986937 DFD BIO | 450 | 9 | 19 | 178 | 0 | 56 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Gaspereau Brk | SU56B | 575100 | 4991275 DFD BIO | 1,680 | 9 | 19 | 2724 | 41 | 136 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gegogan Brk | SU57A | 578885 | 4992601 DFD BIO | 275 | 9 | 20 | 299 | 0 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| Gold | SU25A | 384757 | 4955032 DFD BIO |  | 9 | 8 | 2135 | 74 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gold | SU25B | 383447 | 4956600 DFD BIO |  | 9 | 8 | 1345 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Indian Indian Harbour | SU30A | 428156 | 4949125 DFD BIO | 600 | 9 | 27 | 675 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lakes <br> Indian Harbour | SU59A | 558974 | 4991323 DFD BIO | 245 |  | 20 | 352 | 0 | 147 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 1 |
| Lakes Indian Harbour | SU59B | 587529 | 4999337 DFD BIO | 210 |  |  | 191 | 0 | 44 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lakes | SU59C | 588069 | 4998726 DFD BIO | 338 | 9 | 20 | 200 | 2 | 0 | 15 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |


|  |  | UTM |  | Area <br> $\left(\mathrm{m}^{2}\right)$ | Mo. | Day | Shocking$\qquad$ | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Site ID | Easting | Northing Organiz. |  |  |  |  | Atlantic salmon | American eel | $\begin{aligned} & \text { Brook } \\ & \text { trout } \end{aligned}$ | $\begin{array}{r} \text { Brown } \\ \text { Trout } \\ \hline \end{array}$ | White sucker | $\begin{gathered} \text { SMouth } \\ \text { bass } \\ \hline \end{gathered}$ | $\begin{aligned} & \text { Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Other species |
| Ingram | SU29A | 422975 | 4949929 DFD BIO | 4,500 | 9 | 26 | 2292 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ingram | SU29B | 423579 | 4948756 DFD BIO | 910 | 9 | 26 | 674 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Issac's Harbour | SU62A | 605129 | 5011544 DFD BIO | 600 | 9 | 21 | 479 | 0 | 2 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Issac's Harbour | SU62B | 604190 | 5006040 DFD BIO | 750 | 9 | 21 | 508 | 4 | 300 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Jordan | SU15A | 319911 | 4856886 DFD BIO | 1,500 | 9 | 7 | 410 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Jordan | SU15B | 320548 | 4861404 DFD BIO | 300 | 9 | 7 | 693 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Kirby | SU49A | 543165 | 4972464 DFD BIO | 105 | 9 | 13 | 442 | 16 | 17 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Kirby | SU49B | 543120 | 4972339 DFD BIO | 90 | 9 | 13 | 322 | 14 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Lahave | LHav002 | 366869 | 4930705 DFD BIO | 1806 | 8 | 25 |  | 2 | 0 | 0 | 0 | 0 | 47 | 0 | 0 | 0 | 0 |
| Lahave | LHav008 | 359406 | 4940137 DFD BIO | 1456 | 7 | 21 | 2843 | 123 | 0 | 0 | 0 | 8 | 0 | 3 | 0 | 0 | 5 |
| Lahave | LHav013 | n/a | n/a DFD BIO | n/a | 8 | 29 | 2115 | 16 | 4 | 71 | 0 | 45 | 0 | 55 | 9 | 0 | 0 |
| Lahave | LHav031 | n/a | n/a DFD BIO | n/a | 8 | 4 | 480 | 0 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lahave | LHav101 | 356321 | 4943231 DFD BIO | 1061 | 7 | 13 | 2603 | 60 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 |
| Lahave | LHav102 | n/a | n/a DFD BIO | 753 | 8 | 25 | 520 | 3 | 0 | 7 | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
| Lahave | LHav103 | n/a | n/a DFD BIO | 1081 | 8 | 4 | 1334 | 79 | 0 | 1 | 0 | 21 | 0 | 10 | 0 | 0 | 0 |
| Lahave | LHav104 | 366639 | 4931248 DFD BIO | 1728 | 8 | 29 | 1374 | 22 | 36 | 0 | 0 | 0 | 102 | 0 | 0 | 0 | 2 |
| Lahave | LHav105 | 371358 | 4920720 DFD BIO | 774 | 7 | 19 | 2566 | 69 | 2 | 2 | 0 | 0 | 2 | 7 | 0 | 0 | 0 |
| Lahave | LHav106 | 363843 | 4932713 DFD BIO | 752 | 7 | 26 | 1385 | 106 | 0 | 1 | 0 | 1 | 0 | 3 | 0 | 0 | 2 |
| Lahave | LHav107 | 366437 | 4919392 DFD BIO | 768 | 8 | 21 | 654 | 76 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 0 | 0 |
| Lahave | LHav108 | 372650 | 4940027 DFD BIO | 1051 | 7 | 14 | 2740 | 97 | 1 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lahave | LHav109 | 373431 | 4941739 DFD BIO | 1018 | 8 | 23 | 1457 | 79 | 14 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 0 |
| Lahave | LHav110 | $\mathrm{n} / \mathrm{a}$ | n/a DFD BIO | 1051 | 8 | 22 | 1266 | 30 | 47 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 0 |
| Lahave | LHav111 | 370359 | 4933069 DFD BIO | 605 | 8 | 21 | 778 | 6 | 0 | 0 | 0 | 3 | 8 | 0 | 0 | 0 | 0 |
| Lahave | LHav112 | 359758 | 4940617 DFD BIO | 607 | 7 | 26 | 1741 | 73 | 0 | 0 | 0 | 5 | 0 | 6 | 0 | 0 | 5 |
| Lahave | LHav113 | 360310 | 4933221 DFD BIO | 1290 | 8 | 25 | 2078 | 117 | 0 | 1 | 0 | 3 | 0 | 2 | 0 | 0 | 3 |
| Lahave | LHav114 | 358860 | 4930747 DFD BIO | 900 | 7 | 21 | 1415 | 90 | 0 | 0 | 0 | 10 | 0 | 1 | 0 | 0 | 1 |
| Liscombe | SU55A | n/a | n/a DFD BIO | 1,400 |  |  |  | 50 | 12 | 1 | 0 | 5 | 0 | 3 | 0 | 3 | 0 |
| Little West | SU46A | 534624 | 4972076 DFD BIO | 2,100 | 9 | 13 | 522 | 0 | 39 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Little West | SU46B | 535951 | 4972813 DFD BIO | 910 | 9 | 13 | 405 | 0 | 59 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Martin's | SU24A | 392951 | 4927524 DFD BIO | n/a | 8 | 31 | 343 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Martin's | SU24B | 392289 | 4927380 DFD BIO | 900 | 8 | 31 | 492 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Medway | SU20A | 341672 | 4919524 DFD BIO | n/a | 9 | 5 | 2097 | 74 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Medway | SU20B | 343918 | 4910857 DFD BIO | n/a | 9 | 6 | 1608 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Medway | SU20C | 351993 | 4902512 DFD BIO | n/a | 9 | 6 | 671 | 16 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  |  | UTM |  |  | $\begin{aligned} & \text { Area } \\ & \left(\mathrm{m}^{2}\right) \\ & \hline \end{aligned}$ | Mo. | Day | Shocking Time (s) |  | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Site ID | Easting | Northing | Organiz. |  |  |  |  |  | Atlantic salmon | American eel | Brook trout | Brown Trout | White sucker | $\begin{gathered} \text { SMouth } \\ \text { bass } \\ \hline \end{gathered}$ | Chub spp. | Other Cyprinids | Stickleback spp. | Other species |
| Mersey | SU19A | 351349 | 4884623 | DFD BIO | 675 | 9 | 27 |  | 770 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mersey | SU19B | 356688 | 4882683 | DFD BIO | 120 | 9 | 27 |  | 340 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mersey | SU19C | 355231 | 4882314 | DFD BIO | 700 | 9 |  | 7 | 526 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Middle | SU26A | 398230 | 4936017 | DFD BIO | 3,000 | 9 | 1 |  | 2256 | 74 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Middle | SU26B | 397579 | 4938060 | DFD BIO | 1,800 | 9 | 14 |  | 2165 | 13 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mosher | SU52A | 556233 | 4982256 | DFD BIO | 630 | 9 | 15 | 5 | 1006 | 60 | 36 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Mosher | SU52B | 556285 | 4985530 | DFD BIO | 1,750 | 9 | 15 |  | 590 | 17 | 32 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 40 |
| Mushamush | SU23A | 385626 | 4925671 | DFD BIO | 900 | 8 | 3 |  | 1964 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Mushamush | SU23B | 377794 | 4929736 | DFD BIO | n/a | 8 | 3 |  | 2192 | 191 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Musquodoboit | SU40A | 491834 | 4990569 | DFD BIO | 106 | 9 |  | 7 | 651 | 45 | 0 | 43 | 0 | 2 | 0 | 0 | 0 | 1 | 0 |
| Musquodoboit | SU40B | 497514 | 4994320 | DFD BIO | 105 | 9 |  | 8 | 560 | 14 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| New Harbour | SU63A | 606411 | 5015844 | DFD BIO | 350 | 9 | 2 |  | 420 | 0 | 21 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| New Harbour | SU63B | 609935 | 5014855 | DFD BIO | 495 | 9 | 2 |  | 452 | 0 | 21 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| New Harbour | SU63C | 615608 | 5009214 | DFD BIO | 600 | 9 | 2 |  | 395 | 1 | 23 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nine Mile | SU32A | 441921 | 4944765 | DFD BIO | 2,160 | 9 | 2 |  | 1266 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Petite | SU21A | 383425 | 4899966 | DFD BIO | 2,500 | 9 |  | 1 | 537 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Petite | SU21B | 377846 | 4907845 | DFD BIO | n/a | 9 |  | 1 | 625 | 227 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Petite | SU21C | 381953 | 4900703 | DFD BIO | 2,000 | 9 | 13 |  | 1299 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Petite | SU21D | 381594 | 4901577 | DFD BIO | 1,800 | 9 | 13 |  | 875 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Petite | SU21E | 381613 | 4901787 | DFD BIO | 2,500 | 9 | 13 |  | 1393 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Quoddy | SU51A | 551780 | 4980480 | DFD BIO | 1,725 | 9 | 1 |  | 939 | 9 | 0 | 2 | 0 | 8 | 0 | 4 | 0 | 0 | 1 |
| Quoddy | SU51B | 551818 | 4978808 | DFD BIO | n/a | 9 | 14 |  | 1168 | 11 | 211 | 10 | 0 | 2 | 0 | 4 | 0 | 0 | 2 |
| Quoddy | SU51C | 551510 | 4975720 | DFD BIO | 320 | 9 | 1 |  | 366 | 1 | 46 | 2 | 0 | 2 | 0 | 2 | 0 | 0 | 0 |
| Round Hill | SU2A | 363125 | 4932829 | DFD BIO | 2,500 | 9 | 1 |  | 1215 | 28 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Round Hill | SU2B | 309183 | 4957206 | DFD BIO | 4,000 | 9 |  | 1 | 2820 | 39 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sable | SU17A | 333288 | 4857493 | DFD BIO | 2,250 | 9 |  | 7 | 373 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Digby) | SU8A | 252045 | 4882174 | DFD BIO | 750 | 9 |  | 9 | 2800 | 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Digby) | SU8B | 248584 | 4883140 | DFD BIO | 1,050 | 9 |  |  | 350 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Salmon (Halifax) | SU41A | 496319 | 4964487 | DFD BIO | 360 | 9 | 1 | 1 | 303 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Halifax) | SU41B | 492615 | 4967838 | DFD BIO | 320 | 9 | 1 | 1 | 304 | 0 | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| Salmon (Halifax) <br> Salmon (Lake | SU41C | 491579 | 4968797 | DFD BIO | 1,365 | 9 |  |  | 708 | 0 | 12 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Major) | SU35A | 464089 | 4947787 | DFD BIO | 900 | 9 |  | 6 | 600 | 0 | 25 | 6 | 0 | 4 | 0 | 0 | 0 | 0 | 2 |
| Major) | SU35B | 464285 | 4948899 | DFD BIO | 80 | 9 |  | 6 | 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| River | Site ID | UTM |  | Area $\left(\mathrm{m}^{2}\right)$ | Mo. |  | Shocking <br> Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing Organiz. |  |  |  |  | Atlantic salmon | American | Brook trout | Brown Trout | White sucker | SMouth | $\begin{aligned} & \text { h Chub } \\ & \text { spp. } \end{aligned}$ | Other Cyprinids | Stickleback spp. | Other species |
| Salmon (Lake |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Major) | SU35C | 463504 | 4949986 DFD BIO | 455 | 9 | 6 | 389 | 0 | 53 | 0 | 0 | 10 | 0 | 0 | 0 | 0 | 4 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36A | 470064 | 4948869 DFD BIO | 1,800 | 9 | 6 | 431 | 0 | 90 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 26 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36B | 469587 | 4949205 DFD BIO | 4,100 | 9 | 6 | 1150 | 2 | 115 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 9 |
| Salmon |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lawrencetown) | SU36C | 469158 | 4955033 DFD BIO | 480 | 9 | 7 | 340 | 0 | 11 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmon (Port |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dufferin) | SU50A | 547463 | 4977058 DFD BIO | 2,320 | 9 | 14 | 1505 | 20 | 17 | 1 | 0 | 4 | 0 | 2 | 0 | 0 | 0 |
| Salmon (Port |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dufferin) | SU50B | 548523 | 4979891 DFD BIO | 225 | 9 | 14 | 388 | 0 | 0 | 15 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Ship Harbour | SU42A | 501468 | 4974180 DFD BIO | 960 | 9 | 11 | 460 | 0 | 17 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| Ship Harbour | SU42B | 504637 | 4967824 DFD BIO | 1,365 | 9 | 12 | 1690 | 53 | 51 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Sissibo | SU4A | 264248 | 4922044 DFD BIO | 1,500 | 9 | 12 | 393 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Smith Brk | SU53A | 562351 | 4979865 DFD BIO | 1,120 | 9 | 19 | 1219 | 60 | 16 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's River | STMR8510.2 | 554478 | 5030442 DFD BIO | 1,109 | 8 | 9 |  | 12 | 17 | 0 | 0 | 2 | 0 | 0 | 3 | 0 | 0 |
| St. Mary's River | STMR8510.8 | 553790 | 5030955 DFD BIO | 681 | 8 | 9 |  | 252 | 13 | 0 | 0 | 11 | 0 | 3 | 14 | 0 | 0 |
| St. Mary's River | STMR853.1 | n/a | n/a DFD BIO | 298 | 8 | 15 |  | 10 | 6 | 4 | 4 | 1 | 0 | 2 | 0 | 0 | 0 |
| St. Mary's River | STMR853.2 | 570795 | 5013550 DFD BIO | 678 | 8 | 15 |  | 14 | 3 | 12 | 12 | 3 | 0 | 1 | 0 | 0 | 0 |
| St. Mary's River | STMR854.2 | 577086 | 5013497 DFD BIO | 783 | 8 | 1 |  | 31 | 15 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 236 |
| St. Mary's River | STMR854.4 | 577040 | 5013648 DFD BIO | 673 | 8 | 1 |  | 29 | 17 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's River | STMR855.1 | 561110 | 5013537 DFD BIO | 485 | 7 | 11 |  | 41 | 4 | 16 | 16 | 8 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's River | STMR858.1 | 549950 | 5013416 DFD BIO | 380 | 7 | 11 |  | 0 | 6 | 33 | 33 | 0 | 0 | 0 | 0 | 0 | 0 |
| St. Mary's River | STMR859.4 | 552876 | 5012910 DFD BIO | 3,104 | 8 | 15 |  | 27 | 15 | 0 | 0 | 26 | 0 | 15 | 7 | 0 | 4 |
| St. Mary's River | STMR867.1 | 552930 | 5032085 DFD BIO | 865 | 8 | 8 |  | 164 | 11 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 1 |
| St. Mary's River | STMR867.2 | 552850 | 5032138 DFD BIO | 808 | 8 | 8 |  | 45 | 22 | 1 | 1 | 8 | 0 | 3 | 1 | 0 | 0 |
| St. Mary's River | STMR924 | 546607 | 5014243 DFD BIO | 4,389 | 8 | 25 |  | 91 | 25 | 1 | 1 | 19 | 0 | 8 | 10 | 0 | 0 |
| St. Mary's River | STMR925.1+2 | n/a | n/a DFD BIO | 521 | 8 | 1 |  | 71 | 9 | 1 | 0 | 5 | 0 | 9 | 1 | 0 | 2 |
| St. Mary's River | STMR928 | 526196 | 5016130 DFD BIO | 1,363 | 8 | 11 |  | 54 | 27 | 2 | 2 | 39 | 0 | 20 | 2 | 0 | 0 |
| Tangier | SU43A | 514331 | 4978572 DFD BIO | 805 | 9 | 12 | 708 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Tangier | SU43B | 517069 | 4977081 DFD BIO | 1,190 | 9 | 12 | 345 | 0 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| Tangier | SU43C | n/a | n/a DFD BIO |  | 9 | 12 | 250 | 0 | 125 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tidney | SU18A | 336728 | 4860030 DFD BIO | 1,725 | 9 | 7 | 472 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tusket | SU10A | 265831 | 4895172 DFD BIO | 2,200 | 9 | 19 | 1820 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 30 |
| Tusket | SU10B | 266254 | 4890619 DFD BIO | 3,000 | 9 | 19 | 1727 | 33 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| West Brk (Porter's Lake) | SU37A | 469564 | 4961764 DFD BIO | 1,350 | 9 | 7 | 830 | 0 | 17 | 19 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |


| River | Site ID | UTM |  |  | Area$\left(\mathrm{m}^{2}\right)$ | Mo. | Day | Shocking <br> Time (s) | Catch |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing | Organiz. |  |  |  |  | Atlantic salmon | American eel | Brook trout | Brown Trout | White sucker | SMouth bass | Chub spp. | Other Cyprinids | Stickleback spp. | Other species |
| West River Sheet |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Harbour | SU47A | n/a | n/a | DFD BIO | $\mathrm{n} / \mathrm{a}$ |  |  |  | 59 | 5 | 0 | 0 | 4 | 0 | 1 | 0 | 0 | 13 |
| West River Sheet |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Harbour | SU47B | n/a | n/a | DFD BIO | $\mathrm{n} / \mathrm{a}$ |  |  |  | 1 | 3 | 1 | 0 | 9 | 0 | 16 | 0 | 0 | 0 |
| West River Sheet |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Harbour | SU47C | n/a | n/a | DFD BIO | n/a |  |  |  | 34 | 4 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 |
| West Taylor Bay | SU45A | 529632 | 4966094 | DFD BIO | 1,200 | 9 | 13 | 429 | 0 | 26 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix 4. Recreational catch and effort data from the license-stub return program in SFAs 20 and 21 for 2007.

| River name | Year | Status | Number of anglers | One-sea-winter |  |  | Two-sea-winter |  |  | Total catch | $\begin{gathered} \text { Effort } \\ \text { (rod-days) } \end{gathered}$ | Confidence Interval |  | CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Retained | Released | Total | Retained | Released | Total |  |  | 5\% | 95\% |  |
| Salmon Fishing Area 20 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BARRINGTON | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| CHEZZETCOOK | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| CLAM HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| COLE HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| COUNTRY HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| EAST: SHEET HARBOUR | 2007 |  | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 7 | 7 | 8 | 0.20 |
| ECUM SECUM | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| GASPEREAUX BROOK | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| GEGOGAN | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| GUYSBOROUGH | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| HALFWAY BROOK | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| ISAAC'S HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| KIRBY | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| LARRY'S | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| LAWRENCETOWN LAKE (SALMON RIVER) | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| LISCOMB | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| LITTLE SALMON (LAKE MAJOR) | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| MOSER | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| MUSQUODOBOIT | 2007 |  | 18 | 0 | 27 | 27 | 0 | 8 | 8 | 35 | 126 | 121 | 131 | 0.32 |
| NECUM TEUCH (SMITH BROOK) | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| NEW HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| PORT DUFFERIN | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| PORTERS LAKE (EAST BROOK) | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| QUODDY | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| ROCKY RUN PORTERS L. | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| SAINT FRANCIS | 2007 |  | 1 | 0 | 6 | 6 | 0 | 1 | 1 | 7 | 3 | 3 | 3 | 2.50 |
| SAINT MARY'S | 2007 |  | 69 | 3 | 205 | 208 | 0 | 89 | 89 | 297 | 597 | 573 | 621 | 0.49 |
| SALMON: GUYSBOROUGH CO. | 2007 |  | 7 | 2 | 10 | 11 | 0 | 3 | 3 | 14 | 55 | 53 | 58 | 0.24 |
| SALMON: HALIFAX CO. | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| SHIP HBR. L. CHARLOTTE | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| TANGIER | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| TAYLOR BAY BROOK | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| THREE FATHOM HARBOUR BROOK | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| WEST: SHEET HARBOUR | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| Salmon Fishing Area 21 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BARRINGTON | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| BROAD | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| CLYDE | 2007 |  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 0.00 |
| EAST: LUNENBURG CO. | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| GOLD | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| INGRAM | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| JORDAN | 2007 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |
| LAHAVE | 2007 |  | 46 | 0 | 94 | 94 | 0 | 23 | 23 | 117 | 497 | 477 | 517 | 0.23 |
| MARTINS | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| MEDWAY | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| MERSEY | 2007 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |
| METEGHAN | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| MIDDLE: LUNENBURG CO. | 2007 |  | 1 | 0 | 3 | 3 | 0 | 0 | 0 | 3 | 3 | 3 | 3 | 1.00 |
| MUSHAMUSH | 2007 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |
| NINE MILE | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| PETITE RIVIERE | 2007 |  | 2 | 0 | 10 | 10 | 0 | 3 | 3 | 13 | 33 | 32 | 34 | 0.41 |
| ROSEWAY | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| SACKVILLE | 2007 |  | 7 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 49 | 47 | 51 | 0.03 |
| SALMON: DIGBY CO. | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| SISSIBOO | 2007 | River Closed |  |  |  |  |  |  |  |  |  |  |  |  |
| TUSKET | 2007 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |

Appendix 5. Reported recreational catch and effort in SFA 20 from 1983 to 2007. Missing values indicate a river has been closed to recreational angling.

| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| BARRINGTON | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| CHEZZETCOOK | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| CLAM HARBOUR | Catch Small | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  |  |  |  |
|  | Retained Small | 2 | 0 | 1 | 0 | 0 | 1 | 0 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  |  |  |  |
|  | Effort (rod days) | 4 | 0 | 1 | 0 | 0 | 1 | 0 | 20 | 5 | 2 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  |  |  |  |  |
| COLE HARBOUR | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 9 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 6 | 34 | 23 | 0 | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
| COUNTRY HARBOUR | Catch Small | 25 | 34 | 32 | 43 | 46 | 22 | 23 | 38 | 34 | 4 | 15 | 1 | 24 | 3 | 0 |  | 0 |  |  |  |  |  |  | 0 |  |
|  | Catch Large | 11 | 5 | 7 | 15 | 7 | 13 | 7 | 5 | 11 | 1 | 2 | 0 | 4 | 5 | 0 |  | 0 |  |  |  |  |  |  | 0 |  |
|  | Retained Small | 21 | 24 | 24 | 20 | 33 | 19 | 21 | 17 | 20 | 9 | 16 | 1 | 13 | 1 | 0 |  | 0 |  |  |  |  |  |  | 1 |  |
|  | Retained Large | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  | 0 |  |
|  | Effort (rod days) | 187 | 187 | 200 | 259 | 197 | 198 | 116 | 156 | 150 | 66 | 76 | 7 | 159 | 12 | 0 |  | 0 |  |  |  |  |  |  | 2 |  |
| EAST: SHEET HARBOUR | Catch Small | 0 |  | 110 | 2 | 91 | 148 | 66 | 26 | 14 | 24 | 37 | 0 | 1 | 31 | 3 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Catch Large | 0 |  | 16 | 2 | 19 | 17 | 12 | 0 | 1 | 6 | 5 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Retained Small | 2 |  | 38 | 15 | 38 | 53 | 57 | 34 | 17 | 32 | 47 | 4 | 2 | 23 | 7 | 3 | 1 | 0 | 1 | 3 | 1 | 1 | 0 | 1 | 1 |
|  | Retained Large | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 2 |  | 466 | 63 | 278 | 686 | 437 | 176 | 83 | 205 | 319 | 34 | 13 | 231 | 158 | 7 | 6 | 0 | 3 | 5 | 3 | 5 | 0 | 8 | 7 |
| ECUM SECUM | Catch Small | 88 | 108 | 103 | 74 | 91 | 101 | 69 | 89 | 16 | 56 | 41 | 12 | 17 | 25 | 3 |  | 2 |  |  |  | 1 |  |  |  |  |
|  | Catch Large | 6 | 10 | 13 | 9 | 21 | 10 | 14 | 5 | 1 | 9 | 5 | 1 | 2 | 5 | 0 |  | 0 |  |  |  | 0 |  |  |  |  |
|  | Retained Small | 78 | 69 | 56 | 61 | 59 | 66 | 53 | 53 | 27 | 40 | 42 | 18 | 22 | 11 | 4 |  | 1 |  |  |  | 1 |  |  |  |  |
|  | Retained Large | 3 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  | 0 |  |  |  |  |
|  | Effort (rod days) | 664 | 819 | 914 | 846 | 661 | 603 | 420 | 828 | 359 | 459 | 465 | 151 | 284 | 94 | 167 |  | 2 |  |  |  | 9 |  |  |  |  |
| GASPEREAUX BROOK | Catch Small | 11 | 3 | 4 | 1 | 6 | 2 | 11 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 2 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 11 | 9 | 5 | 6 | 6 | 4 | 7 | 7 | 1 | 3 | 2 | 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 106 | 31 | 18 | 20 | 40 | 21 | 47 | 48 | 2 | 28 | 13 | 7 | 2 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |


| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| GEGOGAN | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
| GUYSBOROUGH | Catch Small | 1 | 2 | 1 | 1 | 0 | 5 | 3 | 2 | 1 | 0 | 3 | 0 | 1 | 0 | 0 |  |  |  |  |  |  |  | 1 |  |  |
|  | Catch Large | 2 | 0 | 0 | 0 | 0 | 3 | 0 | 5 | 3 | 0 | 3 | 2 | 2 | 0 | 0 |  |  |  |  |  |  |  | 0 |  |  |
|  | Retained Small | 15 | 8 | 2 | 4 | 8 | 7 | 2 | 3 | 2 | 1 | 4 | 2 | 2 | 0 | 0 |  |  |  |  |  |  |  | 1 |  |  |
|  | Retained Large | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  | 0 |  |  |
|  | Effort (rod days) | 110 | 43 | 11 | 32 | 35 | 40 | 8 | 19 | 11 | 2 | 17 | 5 | 12 | 0 | 0 |  |  |  |  |  |  |  | 1 |  |  |
| HALFWAY BROOK | Catch Small | 3 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 5 | 2 | 0 | 0 | 0 | 2 | 2 | 0 | 2 | 1 | 0 | 1 | 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 48 | 14 | 0 | 0 | 0 | 7 | 9 | 0 | 5 | 2 | 0 | 11 | 11 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| ISAAC'S HARBOUR | Catch Small | 15 | 10 | 20 | 33 | 32 | 25 | 18 | 30 | 3 | 7 | 6 | 0 | 4 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 12 | 1 | 5 | 1 | 0 | 5 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 21 | 10 | 16 | 12 | 11 | 14 | 14 | 10 | 5 | 5 | 5 | 2 | 5 | 2 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 194 | 93 | 141 | 137 | 116 | 145 | 94 | 123 | 58 | 60 | 36 | 14 | 21 | 10 | 0 |  |  |  |  |  |  |  |  |  |  |
| KIRBY | Catch Small | 17 | 12 | 28 | 18 | 0 | 6 | 9 | 4 | 1 | 3 | 2 | 1 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 4 | 4 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 7 | 5 | 18 | 10 | 3 | 3 | 6 | 5 | 1 | 4 | 3 | 1 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 95 | 53 | 105 | 112 | 25 | 13 | 39 | 23 | 25 | 26 | 28 | 1 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| LARRY'S | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 5 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| LAWRENCETOWN LAKE (SALMON RIVER) | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 5 | 0 | 0 | 1 | 0 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 2 | 0 | 0 | 0 | 1 | 0 | 2 | 5 | 7 | 1 | 1 | 2 | 1 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 8 | 0 | 0 | 0 | 1 | 0 | 11 | 34 | 42 | 5 | 2 | 8 | 3 | 5 | 0 |  |  |  |  |  |  |  |  |  |  |
| LISCOMB | Catch Small | 71 | 65 | 81 | 235 | 289 | 115 | 52 | 154 | 54 | 17 | 14 | 18 | 23 | 1 | 3 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
|  | Catch Large | 10 | 1 | 8 | 28 | 27 | 19 | 7 | 5 | 4 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
|  | Retained Small | 121 | 66 | 72 | 178 | 191 | 157 | 119 | 100 | 80 | 98 | 77 | 37 | 33 | 14 | 5 | 1 | 0 | 0 | 1 | 1 |  |  |  |  |  |
|  | Retained Large | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
|  | Effort (rod days) | 859 | 522 | 318 | 899 | 1103 | 780 | 653 | 643 | 441 | 549 | 349 | 275 | 212 | 49 | 20 | 1 | 0 | 0 | 3 | 1 |  |  |  |  |  |
| LITTLE SALMON (LAKE MAJOR) | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 1 | 0 | 0 | 7 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 1 | 0 | 4 | 0 | 0 | 0 | 31 | 6 | 0 | 0 | 5 | 0 | 0 | 59 | 0 |  |  |  |  |  |  |  |  |  |  |


| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| MOSER | Catch Small | 195 | 256 | 254 | 184 | 183 | 225 | 152 | 221 | 110 | 71 | 120 | 36 | 58 | 32 | 1 |  |  |  |  |  | 0 | 2 |  | 0 |  |
|  | Catch Large | 48 | 17 | 29 | 53 | 12 | 20 | 5 | 18 | 10 | 4 | 10 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 |  | 0 |  |
|  | Retained Small | 192 | 150 | 151 | 145 | 113 | 147 | 111 | 115 | 69 | 81 | 93 | 36 | 46 | 17 | 4 |  |  |  |  |  | 1 | 1 |  | 1 |  |
|  | Retained Large | 42 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 | 0 |  | 0 |  |
|  | Effort (rod days) | 1651 | 1309 | 1750 | 1258 | 765 | 1161 | 1005 | 1087 | 544 | 959 | 1107 | 379 | 459 | 75 | 12 |  |  |  |  |  | 1 | 2 |  | 5 |  |
| MUSQUODOBOIT | Catch Small | 97 | 153 | 400 | 535 | 298 | 313 | 242 | 281 | 110 | 54 | 175 | 60 | 107 | 194 | 20 | 14 | 19 | 1 | 0 | 14 | 25 | 15 | 1 | 27 | 27 |
|  | Catch Large | 111 | 58 | 388 | 372 | 188 | 226 | 131 | 113 | 120 | 20 | 105 | 41 | 94 | 107 | 16 | 10 | 10 | 1 | 3 | 3 | 0 | 2 | 4 | 3 | 8 |
|  | Retained Small | 330 | 197 | 340 | 545 | 394 | 470 | 353 | 365 | 250 | 113 | 359 | 124 | 183 | 76 | 27 | 5 | 8 | 3 | 6 | 7 | 6 | 7 | 6 | 14 | 18 |
|  | Retained Large | 106 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 2925 | 1530 | 3217 | 5201 | 2793 | 3564 | 2641 | 3010 | 1841 | 617 | 3057 | 806 | 1588 | 558 | 168 | 49 | 47 | 4 | 20 | 104 | 62 | 38 | 25 | 115 | 126 |
| NECUM TEUCH (SMITH BROOK) | Catch Small | 4 | 3 | 1 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 2 | 5 | 2 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 5 | 14 | 5 | 12 | 8 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| NEW HARBOUR | Catch Small | 18 | 22 | 41 | 25 | 18 | 12 | 25 | 90 | 19 | 26 | 13 | 17 | 21 | 1 | 0 |  |  |  |  |  | 4 |  |  |  |  |
|  | Catch Large | 1 | 2 | 3 | 5 | 0 | 0 | 1 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 |  |  |  |  |
|  | Retained Small | 37 | 40 | 31 | 29 | 23 | 15 | 24 | 45 | 28 | 34 | 21 | 21 | 18 | 3 | 3 |  |  |  |  |  | 1 |  |  |  |  |
|  | Retained Large | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 |  |  |  |  |
|  | Effort (rod days) | 360 | 347 | 286 | 328 | 220 | 201 | 271 | 552 | 288 | 392 | 226 | 123 | 120 | 20 | 11 |  |  |  |  |  | 3 |  |  |  |  |
| PORT DUFFERIN | Catch Small | 42 | 37 | 79 | 20 | 41 | 31 | 4 | 26 | 4 | 2 | 19 | 5 | 9 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 4 | 2 | 8 | 0 | 7 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 37 | 30 | 42 | 24 | 21 | 27 | 6 | 20 | 14 | 10 | 18 | 12 | 5 | 2 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 294 | 275 | 354 | 200 | 173 | 211 | 72 | 168 | 102 | 101 | 195 | 111 | 38 | 7 | 0 |  |  |  |  |  |  |  |  |  |  |
| PORTERS LAKE <br> (EAST BROOK) | Catch Small |  |  |  |  |  |  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large |  |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small |  |  |  |  |  |  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large |  |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) |  |  |  |  |  |  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| QUODDY | Catch Small | 4 | 5 | 8 | 0 | 0 | 8 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 10 | 5 | 7 | 6 | 5 | 4 | 3 | 4 | 3 | 1 | 2 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 53 | 33 | 30 | 23 | 19 | 28 | 9 | 23 | 15 | 8 | 15 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| ROCKY RUN PORTER'S L. | Catch Small | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 2 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 13 | 0 | 9 | 0 | 0 | 0 | 26 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| SAINT FRANCIS | Catch Small | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  | 6 |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  | 1 |
|  | Retained Small | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  | 1 |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  | 0 |
|  | Effort (rod days) | 4 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  | 3 |


| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| SAINT MARY'S | Catch Small | 746 | 919 | 1453 | 1416 | 612 | 1197 | 517 | 1794 | 816 | 281 | 905 | 33 | 439 | 553 | 98 | 18 | 4 |  | 95 | 38 |  | 39 | 13 | 219 | 208 |
|  | Catch Large | 239 | 231 | 856 | 945 | 321 | 578 | 365 | 238 | 221 | 134 | 395 | 23 | 106 | 164 | 35 | 2 | 1 |  | 75 | 13 |  | 21 | 0 | 69 | 89 |
|  | Retained Small | 1067 | 726 | 846 | 999 | 693 | 908 | 782 | 789 | 650 | 603 | 840 | 274 | 473 | 168 | 76 | 4 | 7 |  | 43 | 33 |  | 13 | 32 | 57 | 69 |
|  | Retained Large | 178 | 66 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 8703 | 5571 | 6896 | 7714 | 4241 | 6810 | 5334 | 5706 | 4725 | 3763 | 6197 | 1268 | 3072 | 976 | 425 | 40 | 19 |  | 244 | 194 |  | 105 | 119 | 476 | 597 |
| SALMON: <br> GUYSBOROUGH CO. | Catch Small | 44 | 271 | 176 | 71 | 173 | 197 | 231 | 247 | 178 | 278 | 215 | 164 | 195 | 147 | 148 |  | 0 |  | 18 | 14 | 1 | 19 | 43 | 10 | 11 |
|  | Catch Large | 19 | 40 | 346 | 152 | 52 | 101 | 166 | 197 | 123 | 174 | 105 | 49 | 134 | 65 | 53 |  | 1 |  | 10 | 5 | 0 | 12 | 14 | 8 | 3 |
|  | Retained Small | 98 | 132 | 125 | 109 | 106 | 131 | 177 | 157 | 127 | 157 | 137 | 103 | 129 | 34 | 30 |  | 8 |  | 8 | 8 | 1 | 4 | 9 | 5 | 7 |
|  | Retained Large | 16 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 1164 | 1569 | 1129 | 1131 | 1015 | 1236 | 1468 | 1560 | 1555 | 1663 | 1454 | 761 | 1404 | 257 | 297 |  | 15 |  | 43 | 54 | 3 | 25 | 87 | 43 | 55 |
| SALMON: HALIFAX CO. | Catch Small | 25 | 8 | 6 | 8 | 4 | 7 | 13 | 19 | 2 | 0 | 3 | 0 | 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 3 | 2 | 2 | 1 | 0 | 2 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 25 | 10 | 16 | 13 | 5 | 8 | 19 | 11 | 5 | 6 | 4 | 0 | 4 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 257 | 100 | 129 | 85 | 50 | 73 | 110 | 109 | 26 | 11 | 32 | 0 | 10 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| SHIP HARBOUR LAKE CHARLOTTE | Catch Small | 36 | 56 | 54 | 28 | 23 | 43 | 39 | 21 | 4 | 0 | 24 | 1 | 15 | 1 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Catch Large | 7 | 12 | 4 | 4 | 4 | 6 | 3 | 2 | 0 | 1 | 4 | 0 | 4 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Small | 26 | 33 | 32 | 43 | 29 | 38 | 36 | 22 | 16 | 14 | 27 | 12 | 17 | 1 | 1 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Large | 7 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 394 | 531 | 398 | 432 | 244 | 424 | 383 | 255 | 227 | 204 | 419 | 192 | 207 | 1 | 4 |  | 0 |  |  |  |  |  |  |  |  |
| TANGIER | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 1 | 1 | 0 | 5 | 2 | 5 | 5 | 4 | 3 | 1 | 2 | 1 | 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 4 | 3 | 0 | 17 | 2 | 13 | 11 | 13 | 15 | 2 | 2 | 2 | 2 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| TAYLOR BAY BROOK | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 9 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| THREE FATHOM HARBOUR BROOK | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| WEST: SHEET HARBOUR | Catch Small | 107 | 189 | 150 | 139 | 170 | 79 | 247 | 138 | 63 | 52 | 75 | 0 | 4 | 18 | 1 | 1 |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 17 | 31 | 24 | 30 | 25 | 18 | 21 | 8 | 6 | 3 | 7 | 0 | 0 | 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 134 | 119 | 102 | 101 | 94 | 79 | 107 | 74 | 52 | 79 | 80 | 3 | 2 | 4 | 2 | 1 |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 16 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 1426 | 1377 | 1197 | 1412 | 1242 | 813 | 1726 | 962 | 742 | 926 | 1195 | 14 | 4 | 62 | 11 | 1 |  |  |  |  |  |  |  |  |  |

Appendix 6. Reported recreational catch and effort in SFA 21 from 1983 to 2007. Missing values indicate rivers that are closed to fishing.

| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| BARRINGTON | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| BROAD | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| CLYDE | Catch Small | 0 | 11 | 33 | 4 | 121 | 116 | 11 | 26 | 13 | 101 | 56 | 22 | 29 | 43 | 17 | 3 | 8 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Catch Large | 0 | 0 | 3 | 0 | 3 | 38 | 9 | 0 | 0 | 13 | 9 | 5 | 5 | 13 | 2 | 1 | 4 | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Small | 0 | 11 | 24 | 4 | 87 | 96 | 10 | 25 | 13 | 93 | 53 | 11 | 28 | 35 | 17 | 3 | 7 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 1 | 83 | 126 | 66 | 509 | 711 | 359 | 414 | 305 | 640 | 654 | 542 | 463 | 674 | 257 | 327 | 213 | 215 | 181 | 4 | 0 | 0 | 3 | 0 | 1 |
| EAST: <br> LUNENBURG CO. | Catch Small | 0 | 3 | 1 | 2 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 30 | 33 | 10 | 12 | 4 | 8 | 5 | 7 | 0 | 7 | 9 | 1 | 3 | 40 | 4 |  |  |  |  |  |  |  |  |  |  |
| GOLD | Catch Small | 59 | 280 | 423 | 314 | 305 | 185 | 422 | 340 | 46 | 179 | 261 | 28 | 148 | 174 | 119 |  |  |  |  |  | 13 |  | 0 |  |  |
|  | Catch Large | 33 | 93 | 175 | 171 | 83 | 25 | 99 | 79 | 21 | 27 | 29 | 9 | 26 | 66 | 25 |  |  |  |  |  | 1 |  | 0 |  |  |
|  | Retained Small | 47 | 245 | 382 | 294 | 284 | 165 | 379 | 277 | 43 | 171 | 223 | 23 | 126 | 156 | 101 |  |  |  |  |  | 0 |  | 0 |  |  |
|  | Retained Large | 30 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  | 0 |  | 0 |  |  |
|  | Effort (rod days) | 2033 | 1934 | 2442 | 2524 | 2100 | 1747 | 2557 | 2690 | 935 | 1685 | 2802 | 886 | 1388 | 1197 | 769 |  |  |  |  |  | 23 |  | 1 |  |  |
| INGRAM | Catch Small | 1 | 1 | 7 | 4 | 4 | 4 | 9 | 3 | 3 | 7 | 14 | 5 | 6 | 7 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 6 | 4 | 12 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 1 | 1 | 7 | 0 | 0 | 4 | 5 | 2 | 2 | 6 | 6 | 3 | 3 | 5 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 53 | 48 | 63 | 51 | 33 | 32 | 45 | 44 | 58 | 65 | 125 | 84 | 203 | 98 | 40 |  |  |  |  |  |  |  |  |  |  |
| JORDAN | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 2 | 5 | 0 | 1 | 1 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| LAHAVE | Catch Small | 300 | 1507 | 1871 | 2118 | 2948 | 1374 | 2255 | 2071 | 232 | 992 | 1118 | 123 | 513 | 1407 | 427 |  | 59 | 0 | 88 | 201 | 154 | 121 | 165 | 211 | 94 |
|  | Catch Large | 221 | 295 | 994 | 952 | 465 | 258 | 529 | 505 | 119 | 160 | 241 | 91 | 194 | 304 | 185 |  | 40 | 0 | 61 | 43 | 113 | 34 | 61 | 65 | 23 |
|  | Retained Small | 271 | 1362 | 1686 | 1847 | 2567 | 1263 | 1932 | 1747 | 195 | 902 | 917 | 105 | 450 | 1011 | 385 |  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large | 209 | 125 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 11056 | 10403 | 10110 | 13026 | 12456 | 9800 | 11127 | 12272 | 3747 | 7746 | 10289 | 3926 | 5377 | 7880 | 4081 |  | 448 | 0 | 288 | 389 | 573 | 325 | 599 | 476 | 497 |


| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| MARTINS | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 4 | 0 | 0 | 3 | 13 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 2 |  |  |  |  |  |  |  |  |
| MEDWAY | Catch Small | 145 | 395 | 475 | 635 | 870 | 515 | 480 | 536 | 42 | 452 | 265 | 56 | 91 | 455 | 101 |  | 0 |  |  |  |  |  | 1 | 0 |  |
|  | Catch Large | 72 | 104 | 299 | 311 | 138 | 162 | 163 | 125 | 40 | 46 | 64 | 31 | 26 | 82 | 55 |  | 0 |  |  |  |  |  | 0 | 0 |  |
|  | Retained Small | 140 | 375 | 454 | 604 | 826 | 474 | 454 | 513 | 40 | 418 | 236 | 52 | 86 | 415 | 90 |  | 0 |  |  |  |  |  | 0 | 0 |  |
|  | Retained Large | 69 | 61 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  | 0 | 0 |  |
|  | Effort (rod days) | 5905 | 4207 | 4140 | 4606 | 5017 | 4710 | 4428 | 4776 | 2267 | 4127 | 4414 | 2369 | 1810 | 3218 | 1690 |  | 6 |  |  |  |  |  | 1 | 3 |  |
| MERSEY | Catch Small | 5 | 9 | 5 | 47 | 77 | 102 | 98 | 134 | 61 | 41 | 15 | 4 | 0 | 5 | 3 | 5 | 5 | 11 | 1 | 17 | 4 | 17 | 4 | 0 | 0 |
|  | Catch Large | 0 | 1 | 1 | 12 | 6 | 23 | 45 | 8 | 17 | 1 | 2 | 5 | 1 | 0 | 3 | 0 | 2 | 0 | 3 | 1 | 0 | 5 | 4 | 0 | 0 |
|  | Retained Small | 5 | 9 | 5 | 46 | 65 | 88 | 95 | 124 | 55 | 39 | 15 | 4 | 0 | 5 | 3 | 5 | 5 | 11 | 0 | 10 | 4 | 17 | 1 | 0 | 0 |
|  | Retained Large | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 201 | 254 | 159 | 661 | 1207 | 1511 | 1668 | 1931 | 1936 | 1422 | 1378 | 975 | 358 | 396 | 148 | 486 | 60 | 135 | 141 | 189 | 75 | 444 | 62 | 0 | 0 |
| METEGHAN | Catch Small |  |  |  |  |  |  |  |  |  |  | 1 | 3 | 9 | 12 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large |  |  |  |  |  |  |  |  |  |  | 1 | 1 | 5 | 5 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small |  |  |  |  |  |  |  |  |  |  | 1 | 0 | 2 | 2 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) |  |  |  |  |  |  |  |  |  |  | 2 | 27 | 75 | 92 | 0 |  |  |  |  |  |  |  |  |  |  |
| MIDDLE: <br> LUNENBURG CO. | Catch Small | 0 | 10 | 12 | 5 | 3 | 4 | 6 | 3 | 1 | 0 | 2 | 4 | 10 | 13 | 1 |  |  |  |  |  |  |  |  | 0 | 3 |
|  | Catch Large | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |  |  |  |  |  |  |  |  | 0 | 0 |
|  | Retained Small | 0 | 4 | 10 | 2 | 1 | 0 | 3 | 1 | 1 | 0 | 0 | 1 | 6 | 10 | 1 |  |  |  |  |  |  |  |  | 0 | 0 |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  | 0 | 0 |
|  | Effort (rod days) | 46 | 28 | 43 | 21 | 61 | 7 | 16 | 19 | 5 | 2 | 17 | 3 | 35 | 39 | 9 |  |  |  |  |  |  |  |  | 3 | 3 |
| MUSHAMUSH | Catch Small | 0 | 3 | 56 | 56 | 63 | 20 | 25 | 33 | 9 | 18 | 16 | 0 | 13 | 18 | 4 |  | 0 |  | 0 | 0 | 1 | 0 | 1 | 3 | 0 |
|  | Catch Large | 0 | 0 | 8 | 11 | 10 | 2 | 6 | 7 | 0 | 3 | 5 | 0 | 1 | 2 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Small | 0 | 3 | 49 | 54 | 58 | 20 | 24 | 32 | 9 | 16 | 12 | 0 | 11 | 16 | 4 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 2 | 33 | 190 | 380 | 317 | 279 | 252 | 289 | 32 | 76 | 138 | 18 | 67 | 171 | 47 |  | 0 |  | 0 | 0 | 1 | 0 | 5 | 6 | 0 |
| NINE MILE | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 25 | 0 | 2 | 8 | 5 | 4 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| PETITE RIVIERE | Catch Small | 43 | 110 | 123 | 90 | 94 | 104 | 157 | 144 | 30 | 178 | 80 | 11 | 26 | 117 | 43 |  | 4 |  |  |  |  |  |  |  | 10 |
|  | Catch Large | 2 | 13 | 46 | 36 | 14 | 22 | 16 | 21 | 5 | 10 | 20 | 10 | 9 | 15 | 15 |  | 5 |  |  |  |  |  |  |  | 3 |
|  | Retained Small | 34 | 103 | 121 | 72 | 85 | 95 | 135 | 136 | 29 | 163 | 73 | 8 | 26 | 91 | 40 |  | 4 |  |  |  |  |  |  |  | 0 |
|  | Retained Large | 2 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  | 0 |
|  | Effort (rod days) | 1125 | 829 | 786 | 863 | 756 | 847 | 915 | 831 | 469 | 1030 | 1125 | 433 | 513 | 570 | 389 |  | 56 |  |  |  |  |  |  |  | 33 |
| ROSEWAY | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |


| River | Variable | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| SACKVILLE | Catch Small |  |  |  |  | 2 | 11 | 44 | 40 | 23 | 56 | 12 | 0 | 21 | 131 | 16 | 2 | 0 |  | 0 | 1 | 7 | 0 | 3 | 7 | 1 |
|  | Catch Large |  |  |  |  | 0 | 2 | 8 | 12 | 11 | 11 | 8 | 0 | 2 | 13 | 0 | 0 | 0 |  | 4 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | Retained Small |  |  |  |  | 2 | 9 | 30 | 30 | 16 | 21 | 11 | 0 | 18 | 47 | 12 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) |  |  |  |  | 6 | 41 | 135 | 257 | 342 | 398 | 425 | 80 | 403 | 824 | 230 | 15 | 2 |  | 15 | 34 | 54 | 31 | 30 | 28 | 49 |
| SALMON: DIGBY CO. | Catch Small | 15 | 24 | 46 | 90 | 79 | 82 | 67 | 49 | 18 | 29 | 15 | 12 | 17 | 88 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Catch Large | 11 | 15 | 12 | 33 | 10 | 24 | 7 | 19 | 6 | 2 | 5 | 0 | 9 | 40 | 3 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Small | 14 | 22 | 33 | 77 | 70 | 66 | 64 | 44 | 15 | 27 | 14 | 12 | 16 | 77 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Retained Large | 11 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 602 | 535 | 333 | 531 | 540 | 560 | 573 | 582 | 301 | 262 | 219 | 160 | 219 | 492 | 55 |  | 0 |  |  |  |  |  |  |  |  |
| SISSIBOO | Catch Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Catch Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Small | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Retained Large | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
|  | Effort (rod days) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| TUSKET | Catch Small | 40 | 72 | 30 | 123 | 394 | 154 | 106 | 120 | 57 | 133 | 56 | 19 | 43 | 124 | 28 |  | 6 |  | 3 | 1 | 11 | 5 | 0 | 0 | 0 |
|  | Catch Large | 19 | 32 | 30 | 58 | 69 | 119 | 29 | 36 | 42 | 34 | 21 | 17 | 13 | 51 | 22 |  | 1 |  | 0 | 1 | 2 | 3 | 0 | 0 | 0 |
|  | Retained Small | 29 | 68 | 27 | 115 | 358 | 139 | 91 | 111 | 48 | 113 | 54 | 14 | 34 | 97 | 22 |  | 6 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Retained Large | 15 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Effort (rod days) | 558 | 495 | 466 | 533 | 1546 | 1481 | 1175 | 1230 | 1119 | 1213 | 1334 | 613 | 624 | 876 | 480 |  | 59 |  | 60 | 24 | 75 | 62 | 12 | 0 | 0 |


[^0]:    * Due to the low number of adults captured on the recapture pass, mean catchability was used to calculate this estimate.

