Fisheries and Oceans

## CSAS

Canadian Science Advisory Secretariat
Research Document 2006/008

Not to be cited without permission of the authors *

## sccs

Secrétariat canadien de consultation scientifique
Document de recherche 2006/008

Ne pas citer sans autorisation des auteurs *

> Standardizing the southern Gulf of St. Lawrence bottom trawl survey time series: Results of the 2004-2005 comparative fishing experiments and other recommendations for the analysis of the survey data

> Normalisation de la série chronologique des relevés au chalut de fond effectués dans le sud du golfe du Saint-Laurent : résultats d'études de pêche comparatives de 2004-2005 et autres recommandations pour l'analyse des données des relevés
H.P. Benoît

Fisheries and Oceans Canada, Gulf Fisheries Centre, P.O. Box 5030, Moncton, NB, E1C 9B6

[^0]
## TABLE OF CONTENTS

ABSTRACT/ RÉSUMÉ ..... iii

1. INTRODUCTION ..... 1
1.1 Background ..... 1
2. METHODS ..... 2
2.1 Comparative fishing experiments ..... 2
2.2 Analysis of Relative Fishing Efficiency of Vessels ..... 3
2.2.1 Fixed effects model ..... 3
2.2.2 Mixed effects model ..... 5
2.2.3 Analysis of catches of other invertebrates and miscellaneous taxa ..... 6
2.2.4 Incorporating data from different comparative fishing experiments ..... 6
2.2.5 Outliers ..... 7
2.2.6 Inter-vessel differences in tow distance ..... 7
2.2.7 Type-I error ..... 7
3. RESULTS ..... 8
3.1 Vessel effects (no covariates) - Fish, Crabs and Squid ..... 8
3.1.1 Fixed-effects model ..... 8
3.1.2 Mixed-effects model ..... 9
3.2 Length-Dependent Vessel effects - Fish, Crabs and Squid ..... 9
3.2.1 Fixed-effects model ..... 9
3.2.2 Mixed-effects model ..... 9
3.3 Depth-Dependent Vessel effects - Fish, Crabs and Squid ..... 10
3.4 Diel Differences in Vessel effects - Fish, Crabs and Squid ..... 10
3.5 Vessel effects - Other invertebrates and Biological Material ..... 10
4. DISCUSSION AND RECOMMENDATIONS ..... 11
4.1 Recommendations -Comparative fishing experiment results ..... 11
4.2 Recommendations - Survey area covered and Repeat sets (2003-2005) ..... 20
4.3 Recommendations - Taxonomic Identification ..... 21
5. ACKNOWLEDGEMENTS ..... 23
6. REFERENCES ..... 24
TABLES ..... 26
FIGURES ..... 68
APPENDICES ..... 121


#### Abstract

Bottom-trawl surveys have been conducted annually in the southern Gulf of St. Lawrence during the month of September since 1971. These surveys provide a time series of information on the abundance, size-composition and distribution of more than 70 species of marine and diadromous fish and over 40 marine invertebrate taxa. However, most research activities utilizing these data are contingent on the continuity of the time series for each taxon. This means avoiding or correcting for any systematic changes in catchability of the survey, such as might occur when there is a change in sampling gear, research vessel or the time of day in which scientific fishing takes place. The research vessel CCGS Teleost replaced the CCGS Alfred Needler as the survey vessel for the southern Gulf of St. Lawrence multi-species survey in 2006. The present report contains results from comparative fishing experiments conducted with these vessels that took place in 2004 and 2005. Recommendations for the application of those results, for dealing with issues related to taxonomic identification during the surveys, and for dealing with issues related to survey coverage during the period of 2003-2005 are also included. This report is a follow-up to Benoît and Swain (2003, Can. Tech. Rep. Fish. Aquat. Sci. 2505), which documents the corrections or considerations that should be taken into account when analysing the southern Gulf of St. Lawrence survey data over the period 1971-2002.


## RÉSUMÉ

Depuis 1971, des relevés au chalut de fond sont effectués tous les mois de septembre dans le sud du golfe du Saint-Laurent. Ces relevés fournissent une série chronologique d'information sur l'abondance, la répartition géographique et la distribution des tailles de plus de soixante-dix espèces de poissons marins et diadromes ainsi que de plus de quarante taxons d'invertébrés marins. Cependant, la plupart des activités de recherche s'appuyant sur ces donnés dépendent de la continuité des séries chronologiques pour chaque taxon. On doit donc éviter ou corriger tous les cas où la capturabilité d'une espèce par le relevé a changé de façon systématique, notamment lorsqu'il y a un changement de navire ou d'engin de pêche scientifique ou lorsqu'il y a un changement dans les heures durant lesquelles l'échantillonnage a lieu. Le NGCC Teleost a remplacé le NGCC Alfred Needler comme navire de recherche scientifique pour le relevé annuel au chalut de fond effectué dans le sud du golfe du Saint-Laurent en 2006. Le présent document rapporte les résultats d'études de pêche comparatives entre ces deux navires, qui ont eu lieu en 2004 et en 2005. Des recommandations sont aussi présentées pour la mise en application de ces résultats, pour rectifier les problèmes reliés à l'identification taxonomique de certaines espèces lors des relevés et pour résoudre les problèmes reliés à la couverture géographique du relevé en 2003-2005. Le présent document fait suite à celui de Benoît et Swain (2003, Can. Tech. Rep. Fish. Aquat. Sci. 2505), qui contient des corrections ou recommandations à appliquer lors des analyses des donnés du relevé du sud du golfe du Saint-Laurent pour la période 1971-2002.

## 1. INTRODUCTION

Bottom-trawl surveys have been conducted annually in the southern Gulf of St. Lawrence (NAFO Div. 4T) during the month of September since 1971 (for details see Hurlbut and Clay, 1990). These surveys provide a time series of information on the abundance, sizecomposition and distribution of over 70 marine and diadromous fish species and over 40 marine invertebrate taxa (Benoît et al., 2003a,b). This information is the cornerstone for the majority of the stock assessments of commercially important marine fishes in the southern Gulf. It is also crucial in assessing the status of many marine fishes as part of Fisheries and Oceans Canada's (DFO) Species-at-Risk mandate (Benoît et al., 2003a; Swain et al., 2006;) and in understanding changes in the structure and function of the ecosystem as a whole. These research activities are contingent upon the continuity of the time series for each species. Survey timing (i.e., season), area sampled, time of day in which fishing takes place, and the research vessel and gear used are all known to affect the availability of organisms to the gear or their catchability (e.g., Benoît and Swain 2003a,b; Pelletier 1998; Nielsen 1994). Any change in catchability resulting from modifications in one or more of these factors could, if unaccounted for, be incorrectly be interpreted as a change in resource abundance.

With the exception of the addition of three inshore strata (401, 402 and 403) in 1984 (Fig. 1a), both survey timing and area have been kept constant since 1971 in the September survey. However, past changes in survey vessel, fishing gear and the time of day of fishing have necessitated some corrections to ensure consistency of the time series for many taxa (Benoît and Swain, 2003 a,b). Furthermore, unsampled strata and repeat fishing sets (by a single vessel) at particular locations in certain years, as well as changes in the level and accuracy of taxonomic identification during surveys are all relevant for the proper analysis of the September survey data. Recommendations for dealing with these issues over the period 19712002 are documented in Benoît and Swain (2003b).

The research vessel CCGS Teleost replaced the CCGS Alfred Needler as the survey vessel for the southern Gulf of St. Lawrence multi-species survey in 2006. This report presents results of comparative fishing (paired-trawl) experiments conducted between these vessels in 2004 and 2005. Included are recommendations for the application of those results, as well as further recommendations for dealing with issues related to taxonomic identification during the survey, and also with issues related to survey coverage during the period of 2003-2005. The recommendations contained herein are in addition to those contained in Benoît and Swain (2003b) and supersede them only where specifically indicated.

### 1.1 BACKGROUND

Fishing during the September survey was carried out by the E.E. Prince from 1971-1985 using a Yankee-36 trawl and subsequently by four vessels each using a Western IIA trawl: the Lady Hammond (1985-1991), the CCGS Alfred Needler (1992-2002, and 2004-2005), the CCGS Wilfred Templeman (2003) and the CCGS Teleost (2004-2005). Specifications of the first three vessels and both gears can be found in Nielsen (1994) or Hurlbut and Clay (1990). Details on the CCGS Wilfred Templeman and the CCGS Teleost, fifty and sixty-three meter stern trawlers respectively, can be found at: http://www.ccg-gcc.gc.ca/fleet/main e.asp.

Prior to the vessel/gear changes that occurred in 1985 and 1992, comparative fishing experiments were conducted to determine the efficiency of the new vessel relative to the one being replaced (see Benoît and Swain 2003b). The CCGS Alfred Needler was unavailable for
the 2003 survey, and was temporarily replaced by the CCGS Wilfred Templeman. Comparative fishing experiments between these vessels using the Western IIA trawl have not taken place. Their relative fishing efficiency is therefore not known but is expected to be very similar as the two vessels are of the same design and few differences in efficiency were found between them using a different trawl (Cadigan et al 2006). The CCGS Teleost replaced the CCGS Alfred Needler for the September survey in 2006. In preparation for this, comparative fishing experiments therefore took place in 2004 and 2005 to intercalibrate these vessels. Onehundred and eighty paired fishing tows were planned for each year. However, due to mechanical problems and a labour dispute, the CCGS Alfred Needler was available for only a small portion of the 2004 survey. Consequently, only eleven comparative fishing sets were completed that year and the CCGS Teleost undertook the regular survey sampling. Inclement weather in 2005 resulted in only ninety comparative fishing sets being successfully completed and prevented either of the vessels from completely sampling the survey area, though they accomplished this jointly.

Comparative fishing experiments between the CCGS Alfred Needler and the CCGS Teleost using the Western IIA trawl also took place during the February survey of George's Bank, the March survey of NAFO areas 4VsW and the July survey of the Scotian Shelf (NAFO 4VWX and 5 Ze ). The latter experiment covered a biotic community that is similar to that of the southern Gulf in many respects, during a season when the behaviour of those biota should also be comparable. The results from the July experiment were therefore combined with those from the southern Gulf, given the relatively small overall number of successful comparative fishing sets and resulting lower statistical power for the latter.

A final important background item for the southern Gulf survey is to note that fishing was restricted to daylight hours (07:00-19:00) from 1971 to 1984 but has been extended to 24 hours per day since 1985. Because it is well known that fishing efficiency can vary by time of day (e.g., Benoît and Swain, 2003a; Hjellvik et al. 2002; Casey and Myers 1998) as a result of species-specific diel behaviours such as vertical migrations, hiding and trawl avoidance, it is necessary to correct survey catches to a standard time of day in order to maintain a consistent time series prior to and after 1985. A summary of such corrections for 51 fish and 13 invertebrate taxa in the September survey up to 2002 can be found in Benoît and Swain (2003b). The present report provides similar recommendations for the surveys post-2003.

## 2. METHODS

### 2.1 COMPARATIVE FISHING EXPERIMENTS

Comparative fishing in the southern Gulf between the CCGS Teleost and the CCGS Alfred Needler using the Western IIA trawl took place during the regular 2004 and 2005 surveys, specifically from September 15-16, 2004, and September 11-26, 2005. As in the regular annual surveys, fishing locations were selected randomly within strata (Fig. 1a) and standard protocols were followed. The vessels fished side-by-side ( $\leq 1 \mathrm{n}$. mile apart) and the relative position of the vessels (port or starboard) alternated at each station. The target fishing procedure was a 30 minute tow at 3.5 knots. Paired fishing was successfully conducted at 11 and 90 stations respectively during the 2004 and 2005 comparative fishing experiments (Fig. 1b). The taxonomic and common names of species covered by this study are listed in Appendix I .

Comparative fishing between the aforementioned vessels on the Scotian Shelf also took place during the regular survey of that area, from June 27 to July 26, 2005. Note that attempts to undertake comparative fishing during the 2004 July survey were unsuccessful. Protocols were as described above, with 173 successful comparative sets completed. Only the catches for those species of fish, crabs and squid which were also captured in the September survey were analysed. As a result of problems distinguishing white hake (Urophycis tenuis) from red hake (Urophycis chuss) in certain geographic areas of the Scotian shelf, only catches from NAFO division 4V (July survey strata 440-451) were selected (J. Simon, DFO Maritimes Region, personal communication). Likewise, because of difficulties distinguishing winter skate (Leucoraja ocellata) from little skate (Leucoraja erinacea) below a length of 55 cm , only larger fish of the former species were included in the analyses (J. Simon, DFO Maritimes Region, personal communication). Neither red hake nor little skate occur in the southern Gulf. Results from the Scotian Shelf comparative fishing experiments are presented here only in combination with those from the southern Gulf experiments and therefore are not explicitly presented in detail.

### 2.2 ANALYSIS OF RELATIVE FISHING EFFICIENCY OF VESSELS

The goal of the analyses presented here was to estimate the relative fishing efficiency of the CCGS Alfred Needler (denoted $N$ ) and the CCGS Teleost ( $T$ ). Two approaches were used: a fixed effects conditional distribution model (McCullagh and Nelder, 1989) and a mixed effects model (Pinhero and Bates, 2000). Each is described in turn in more detail later in this section. Past analyses of southern Gulf of St. Lawrence comparative fishing data adopted the former approach (e.g., Benoît and Swain, 2003b). Recently, Cadigan et al. (2006) proposed using mixed random-effects models to analyze these types of data, as is increasingly being done in gear selectivity studies (e.g., Millar et al. 2004). They concluded that mixed effects models were advantageous because they produced apparently more reliable statistical inferences (i.e. confidence intervals) and could better account for comparative fishing data outliers, as compared to their fixed effects counterparts. However, they stated that simulations are required to ensure the robustness of the results. Although this work is underway (N. Cadigan, personal communication), in the absence of conclusions, I have undertaken the analyses using both approaches, allowing for the evaluation of sensitivity to outliers and the consistency of results between approaches.

Throughout this section, $C_{i v}$ and $C_{i v v}$ will denote the total number and the number at length $I$, respectively, of fish, crabs or squid caught in the $i^{\text {th }}$ paired-fishing set by vessel $v=(N$ or $T)$.

### 2.2.1 FIXED EFFECTS MODEL

Under the fixed effects approach, the conditional distribution of $\mathrm{C}_{\text {iN }}$ given the toal catch $\mathrm{C}_{i}=\mathrm{C}_{i T}$ $+\mathrm{C}_{i n}$ was used and the relative fishing efficiency of the vessels was evaluated using a generalized linear model, with a logit link and a binomial error distribution (McCullagh and Nelder, 1989). Such a model evaluates the probability, $p_{i}=\mathrm{C}_{i \mathrm{~N}} / \mathrm{C}_{i}$, that a fish captured in set pair $i$ will have been captured by the CCGS Alfred Needler, relative to the probability that the CCGS Teleost will have captured this fish, 1- $p_{i}$. If there is no difference in fishing efficiency between the vessels for the given species, then $p_{i}=1-p_{i}=0.5$. However, if a difference exists in relative fishing efficiency, 1- $p_{i}$ must be multiplied by a relative catchability term $\left(b_{\mathrm{v}}\right)$ to maintain this equality. This relative catchability term is a function of the estimated intercept parameter of a logistic regression ( $\beta$, is termed the vessel effect hereafter):

$$
\text { (1) } \ln \left(\frac{p_{\mathrm{i}}}{1-p_{\mathrm{i}}}\right)=\beta_{\mathrm{v}} ; \quad \text { where } \exp \left(\beta_{\mathrm{v}}\right)=b_{\mathrm{v}}
$$

The left side of the equation is the logit transformation of $p_{i}$.
Adjustments to (1) are required because of variations in tow distance ( $d_{i v}$ ) and subsampling of catches. While the target fishing procedure was a 30 minute tow at 1.75 knots, variation in realized speed over the ground and occasional pre-emptive (early) haul-back due to problematic trawling conditions (e.g., rough bottom) resulted in differences in $d_{i v}$. Furthermore, representative length-frequencies were obtained for all fish species, most crab species and squid during each set, and subsampling was occasionally used when catches were large ( $>200$ individuals). An offset term, $\log _{e}\left(d_{i N} f_{i N} / d_{i T} f_{i T}\right.$, ) was therefore included in the logistic regression model to account for these sources of variability (McCullagh and Nelder, 1989), where $f_{i v}$ is the fraction (by weight) of the catch for which individuals were measured and counted. An offset term is essentially a covariate that has a slope fixed at one. This is described in more detail later in Section 2.2.2. The model was estimated using the maximum likelihood approach.

Though model (1) can easily handle catches of zero fish for a given species, it is obvious that a null catch for both vessels carries no information about their relative fishing efficiency and inappropriately inflates the degrees of freedom, making nominal tests of significance more liberal. As a result, only set pairs in which a given species was captured by at least one of the vessels were included in the analysis (termed relevant set pairs hereafter). Inclusion of set pairs where one vessel caught no fish is necessary to properly evaluate differences between vessels in the probability of capturing those fish.

Model (1) assumes that individual fish are captured independently, i.e., in the absence of a difference in fishing efficiency between vessels, each fish has an equal probability of being captured by either vessel. This may not always be an appropriate assumption given that fish often aggregate spatially and are therefore not captured independently. To allow for such a departure, an extra-binomial model is generally used in which overdispersion is modelled using a scale parameter $\phi$ that increases the model variance when the data are overdispersed ( $\phi>1$ ), but does not affect the parameter estimates. However, previous experience with similar models used to estimate diel and vessel effects on relative catchability (Benoît and Swain 2003a,b; Casey and Myers 1998) has suggested that this approach does not completely account for the true variability in the data, resulting in overly liberal tests of nominal statistical significance. As a result, statistical significance was assessed using randomization tests (Manly 1991) under the null hypothesis of no difference in fishing efficiency between vessels. The Pearson Chi-Square statistic (McCullagh and Nelder, 1989) was used as the test statistic. Nine-hundred and ninety-nine permutations were used with one $C_{i v}$ from each relevant set pair being randomly assigned to the CCGS Alfred Needler, and the other to the CCGS Teleost. For species where there were fewer than 12 relevant set pairs (s) and for which nominal tests were statistically significant, randomizations were limited to all possible permutations of the data $\left(2^{s}\right)$. Statistical significance was given by $(n+1) / N$, where $N$ is the total number of permutations of the data (including the original result) and $n$ is the number of random permutations that yielded a test statistic equal to or greater than that of the original observed result.

Model (1) can easily be modified to incorporate covariates that might affect the relative catchability of species between vessels, such as fish body length (or crustacean carapace
width), fishing depth and time of day (e.g., Benoît and Swain 2003a; Pelletier 1998). The magnitude of the covariate effect is effectively estimated as the slope $\left(\beta_{1}\right)$ in the case of length and depth, from the logistic regression:
(2) $\ln \left(\frac{p_{\mathrm{i}}}{1-p_{\mathrm{i}}}\right)=\beta_{\mathrm{v}}+\beta_{1} \cdot$ covariate; $\quad$ where $\exp \left(\beta_{\mathrm{v}}+\beta_{1} \times\right.$ covariate $)=b_{\mathrm{v}}$

As with Model (1) an offset term was included in fitting model (2) and statistical significance of the covariates was assessed using 999 randomizations. In order to isolate the probability of the covariate alone, statistical significance was assessed using randomizations of the covariate effect while maintaining the original vessel effect. In other words, the allocation of an observed catch to a particular vessel was not permuted, however catches from set pairs were randomly allocated to the observed levels of the covariate (i.e., in the same proportion as originally observed). As a result of limited sample sizes, the significance of each covariate was assessed in a separate analysis.

### 2.2.2 MIXED EFFECTS MODEL

Although the vessels fished as closed together as possible during the paired-trawl experiments, it was not possible to ensure that exactly the same local stock densities were fished by both vessels. The fixed-effect model assumption that a fish captured in set pair $i$ had an equal probability under the null hypothesis of being captured by either vessel is therefore not generally insured. Failure to account for differences in local densities encountered by each vessel in gear size-selectivity studies is known to lead to underestimated parameter standard errors and overly liberal tests of significance (Millar et al, 2004), likely explaining the aforementioned results obtained by Benoît and Swain (2003a,b) and Casey and Myers (1998). Unlike the fixed effects model, a mixed effects model that assumes that both vessels encounter the same local density of fish ( $\lambda_{i l}$ ) only on average can be formulated. If $\delta_{i}=\ln \left(\lambda_{i N} /\right.$ $\lambda_{i T}$, ) and $z_{i}=\log _{\mathrm{e}}\left(d_{i N} f_{i N} / d_{i T} f_{i T}\right.$,) denotes the offset term described above, the mixed effects version of (1) can be written as

$$
\text { (3) } \ln \left(\frac{p_{\mathrm{i}}}{1-p_{\mathrm{i}}}\right)=\beta_{\mathrm{v}}+z_{i}+\delta_{i} \quad \text { where } \quad \delta_{i} \sim N\left(0, \sigma^{2}\right)
$$

In this model, the $\delta$ 's are (unobserved) random variables and $\beta_{v}$ is treated as a fixed effect parameter. Because (3) contains both fixed and random effects, it is referred to as a mixed effects model.

Cadigan et al. (2006) considered these random effects in an analysis of the length-dependent relative catchability of two vessels where $\delta_{i j}=\log _{\mathrm{e}}\left(\lambda_{i j N} / \lambda_{i I T}\right)$ and,

$$
\text { (4) } \ln \left(\frac{p_{\mathrm{il}}}{1-p_{\mathrm{il}}}\right)=\beta_{\mathrm{v}}+\beta_{1} I+z_{j l}+\delta_{i \mathrm{il}}
$$

If an identical length distribution of fish were encountered by both vessels in $i$ then $\delta_{i j}=0$. However, the authors point out that in practice the length distributions can differ, with differences possibly occurring systematically with length. Consequently in addition to the
assumption of Normally-distributed random effects, they assumed that the $\delta$ 's were autocorrelated (first-order) in terms of length but independent between sets. Both $\beta_{v}$ and $\beta_{1}$ were treated as fixed effects. Because the model can account for smooth deviation from linearity in the logit proportion of Alfred Needler (in the present case) catches caused by partly systematic differences in local stock density fished by each vessel, an additional overdispersion parameter as in the fixed-effects case is not needed.

The approach of Cadigan et al. (2006) was applied to the analysis of comparative fishing data from the southern Gulf using the SAS procedure GLIMMIX for estimation (SAS Institute Inc. 2005). This procedure fits generalized linear mixed models based on linearization. The default estimation method known as residual pseudo-likelihood with a subject-specific expansion was used.

### 2.2.3 ANaLYsis of catches of other invertebrates and Miscellaneous Taxa

The models described in sections 2.2.1 and 2.2.2 are based on discrete probability distributions. They therefore apply to the number of individuals caught but not to the weight caught (a continuous variable). For most invertebrates captured in the survey, only the total catch weight in a set is recorded. Different models were therefore required to analyse those catches. Linear models with an identity link and normal error were used. Taking $W_{i v}$ as the catch weight of the species in set $i$, by vessel $v$, the following fixed-effects model was used:

$$
\text { (5) } \ln \left(W_{i v}+0.001\right)=s_{i}+\beta_{v}
$$

where $s_{i}$ is a factor representing the set and $\exp \left(\beta_{v}\right)=b_{v}$.
A mixed-effects analog, treating the interaction between the set and vessel effects as a random-effect, was defined as:

$$
\text { (6) } \ln \left(W_{i v}+0.001\right)=s_{i}+\beta_{v}+\delta_{i} \quad \text { where } \quad \delta_{i} \sim N\left(0, \sigma^{2}\right)
$$

### 2.2.4 INCORPORATING DATA FROM DIFFERENT COMPARATIVE FISHING EXPERIMENTS

Throughout the analyses, the data from the southern Gulf in 2004 and 2005 were treated as coming from a single experiment. The small number of set pairs in 2004 prevented a rigorous assessment of a year effect on the relative efficiency of the two vessels. Furthermore, a common (though not ubiquitous) assumption in the analysis of research survey data is that catchability by a particular vessel and gear is generally constant over time, hence a constant relative catchability between vessels.

As stated previously, the experiments conducted on the Scotian Shelf in July 2005 were very comparable to those from the southern Gulf in species composition, season and, to a lesser extent, habitat. Judicious combining of the data from those surveys was considered advantageous in increasing sample sizes and possibly increasing the contrasts in fish abundance or covariate magnitude among set pairs. Where sample sizes permitted, I tested for an effect of survey, treated as a binary fixed factor, in models (1) and (3) and an interaction between survey and the covariates in models (2) and (4). For the fixed effects models, randomizations were used to test the significance of the survey factor or interaction by randomly allocating set pairs to the surveys, in their original proportion. For the mixed effects models, Type-III tests were used. Survey data were combined for further analyses where the
effect was not significant at a Type-I error rate of $5 \%$. Nonetheless, results of both the September-only and combined September-July (where appropriate) analyses are presented here. Note that in the results tables, blank entries for the test of a survey effect and the combined September-July analyses indicate that there were too few sets in the July survey that captured the species in question to include those sets.

### 2.2.5 Outliers

During each series of analyses, the standardized residuals from the model fit were examined to identify outliers and sets pairs with potentially inflated leverage in the analysis, and to ensure whether a proposed model was appropriate when covariates were included (Figs 3-49). Cases with possible outliers or unduly influential set pairs (listed in Appendix II) were removed and the analysis was repeated, along with another assessment of the residuals. Results of these analyses as well as those including all sets are presented here.

Preliminary analyses confirmed the results of Cadigan et al. (2006) that mixed effects models were generally less sensitive to outliers. The magnitude of outlying Pearson residuals and the frequency of clear outliers were considerably less than in the fixed effects models. However, to better examine the robustness to outliers, those identified in the fixed effects analyses were removed and the mixed effects analyses were then repeated.

### 2.2.6 INTER-VESSEL DIFFERENCES IN TOW DISTANCE

Adjusting the model to reflect relative catches per tow distance using the offset is likely insufficient to compensate for between-vessel differences in tow duration if these differences can be large and the relationship between catch amount and tow duration/distance is not linear. Although the differences in tow distance were generally less than $5-10 \%$, some were as high as $>50 \%$ (Table 1, Fig. 2). Analyses were therefore done including all relevant set pairs and excluding those with a large difference in towed distance. A cut-off of $\geq 20 \%$ difference (corresponding to a difference of 6 minutes or less in tow duration, given a maximum 30 min . tow) was applied to exclude pairs with relatively large differences. Seven set pairs from the September experiments and four from the July experiments were consequently eliminated. In almost all cases, the removed set pairs included only a small percentage of the total experiment-wide catch of the various species (Appendix III). However, of the seven pairs from September, five were cases where the Teleost fished a longer distance than the Needler (Table 1). It is this sort of systematic difference between vessels that could generate a spurious vessel effect for relative catchability.

Because of the somewhat arbitrary nature of the choice in cut-off level, results of analyses that include all data are also reported, permitting an evaluation of the sensitivity of results to the inclusion of these sets.

### 2.2.7 TYPE-I ERROR

A large number of statistical tests were undertaken as part of these analyses, resulting in an experiment-wise Type-I error (i.e., reject the null hypothesis when it is true) rate that was higher than the nominal level. Procedures are available to control the Type-I error rate at a specified level when multiple tests are performed (e.g., Rice 1989). However, these procedures also increase the Type-II error rate (i.e., failure to reject the null hypothesis when it is false). The power of analyses (i.e., the ability to detect a false null hypothesis) of comparative fishing data is already very low (reviewed briefly in Pelletier 1998) due to small
sample size combined with the high variability characteristic of trawl survey catch rates. Thus, no adjusted to significance levels were made to control the Type-l error rate given multiple tests. Experiment-wise Type-I error is however borne in mind on a case-by-case basis when interpreting the results of analyses that are only marginally statistically significant at the $5 \%$ level and for which the results were not corroborated by those including the July experiment data.

## 3. RESULTS

Bivariate plots of CCGS Teleost and CCGS Alfred Needler catches (\# fish•tow ${ }^{-1}$ ), relative length frequencies from the 2004 and 2005 September and 2005 July comparative fishing experiments and diagnostic plots (residuals and random effects) from the various analyses are presented in Figures 3-49 for those fish, crab and squid taxa for which sufficient catches were made. Bivariate plots of catches (kg•tow ${ }^{-1}$ ) for the remaining invertebrate taxa for which sufficient catches were made are presented in Figures 50-51.

In this section, a general description of analysis results is presented, focussing on cases where the probability of the data given the respective hypothesis was near or below $5 \%$. Species-specific recommendations for the application of vessel conversion factors in light of brief summaries of results are presented in section 4.

### 3.1 Vessel effects (no covariates) - FISh, Crabs and SQuid

### 3.1.1 FIXED-EFFECTS MODEL

Vessels effects were assessed for about 50 species of fish, five crab species and for longfinned squid (Table 2). Restricting the analysis to sets pairs where the distance towed differed by less than $20 \%$ between vessels, significant vessel effects were found for white hake, Greenland cod, daubed shanny and the toad crab Hyas araneus, as well as marginally significant results (i.e., $P \sim 0.05$ ) for American plaice, yellowtail flounder, longhorn sculpin and Arctic alligatorfish. When outlying influential set pairs were removed, significant results were found for Atlantic herring and marginally significant results for Atlantic cod and snow crab. Inclusion of the July data (where appropriate) generally corroborated these results, with a significant effect found for white hake, American plaice (with and without outliers) and yellowtail flounder. Additionally, significant results were found for winter flounder, sea raven (outliers removed), sand lance, snakeblenny, Vahl's eelpout and Northern stone crab. The pvalue for cod, increased slightly relative to the analysis of the September-only data.

Analyses including all relevant set pairs, regardless of the relative tow distance, produced similar conclusions for most species (Table 3; Fig. 52). A notable exception was Atlantic cod, where significant results were obtained in the September-only and September-July analyses once outliers were removed. Another was the Atlantic hookear sculpin, where marginally significant results were also found in the combined surveys analysis.

For most of the larger-bodied demersal fishes, $\beta$, had positive values (i.e., Alfred Needler catching relatively more fish), suggesting a possible overall vessel effect despite a lack of nominal statistical significance in many cases.

### 3.1.2 MIXED-EFFECTS MODEL

The mixed-effects model fit the data very well, with approximately normally-distributed residuals and random effects for the majority of species (Figs. 3-49). The magnitude and frequency of apparent residuals was considerably less than their fixed-effects counterparts. The estimated vessel effect and its associated standard error were also less sensitive, though not insensitive, to the removal of a small number of data points identified in the fixed-effects analysis (Table 4; e.g., cod, witch flounder, arctic staghorn sculpin).

The analysis nonetheless produced very similar conclusions to the fixed-effects model (Fig. 53). Significant vessel effects were obtained for white hake, American plaice, Atlantic herring (September-only data, outliers removed), Greenland cod, longhorn sculpin, moustache sculpin, snakeblenny, daubed shanny, Atlantic hookear sculpin and the toad crab (Table 4). Inclusion of all relevant set pairs, regardless of relative tow distance, had a similar effect as in the fixedeffects analysis. A nominally significant vessel effect was found for cod when outliers were removed (September and combined-survey analyses), though the effect was marginally significant when sets were also selected based on relative tow distance.

As in the fixed-effects analysis, $\beta_{v}$ had a positive value for most of the larger-bodied demersal fishes.

### 3.2 Length-Dependent Vessel effects - Fish, Crabs and SQuid

### 3.2.1 FIXED-EFFECTS MODEL

A statistically significant length-dependent difference in catchability between the CCGS Alfred Needler and CCGS Teleost was established for herring, Greenland cod and daubed shanny, with more marginally significant results for winter flounder (excluding outliers), capelin, longhorn sculpin (including outliers) and toad crab in the September comparative fishing experiments (Table 6). In the case of the Greenland cod, possible confusion with Atlantic cod at smaller sizes, combined with a small number of total fish caught, suggests that its result may not be reliable. In the combined September-July analysis, a significant effect of length was found for American plaice, moustache sculpin and alligator fish. A significant 'survey' effect precluded testing for a length effect in many of the species for which that effect was significant based on the September-only data. One exception was capelin, for which the length effect was not significant, unlike in the September-only analysis. These results are comparable to those obtained when all relevant set pairs are included, regardless of relative tow distance (Table 7).

### 3.2.2 MIXED-EFFECTS MODEL

A significant effect of length was found for daubed shanny using the September data, and for American plaice (excluding outliers), alligatorfish and sand lance using the combined surveys data. Marginally significant results were obtained for redfish (combined surveys), herring, gaspereau, moustache sculpin, spatulate sculpin, sea poacher, stout eelblenny and toad crabs (September-only, but not the combined analyses) (Table 8). There is therefore general concordance with the fixed effects analysis. Likewise for analyses including all relevant set pairs (Table 9).

### 3.3 Depth-Dependent Vessel effects - Fish, Crabs and Squid

There were very few instances of significant depth effects in the fixed effects model analysis of the comparative fishing data (Table 10). The four cases that were nominally significant had pvalues very close to $5 \%$. Furthermore, based on binomial probability with $\sim 100$ statistical tests, as presented in this table, and $\alpha=5 \%$, we would expect approximately $8-9$ tests to produce nominally significant results by chance alone five percent of the time. Additionally, there was no inter-species consistency in the sign of the depth effect that would suggest an overall effect. This was also true for the mixed effects analysis (Table 12) and for analysis based on all relevant set pairs (Tables 11 \& 13).

### 3.4 Diel DIFFERENCES IN VESSEL EFFECTS - FISH, CRABS AND SQUID

Benoît and Swain (2003a) found significant diel differences in the catchability of many species to the September survey, presumably related to diurnal changes in vertical position in the water column, hiding behaviours or visual net avoidance. Differences between vessels in factors such as vertical trawl opening and door stability (inasmuch as it affects sediment resuspension) could conceivably result in diel-dependent differences in relative vessel catchability. However, preliminary fixed effects analyses based on model (2), with time-of-day treated as a binary factor (day, 7:00-18:59), provided little support for such an effect (Table 14). As with the analyses of an effect of depth, significant diel effects were found for only two species, at $p$-values close to $5 \%$. Furthermore, most of the catches of these two species were made during the night, thereby preventing a proper estimation of the diel effect. Consequently diel-dependent differences between the two vessels appear to be negligible. This isconsistent with the findings of Benoitt and Swain (2003a) that conversion factors for particular species tended to be very similar between vessels and even surveys. As a result, any further analyses of the effect using mixed-effects models were not pursued.

### 3.5 Vessel effects - Other invertebrates and Biological Material

For invertebrates other than the large crabs and squid, only analyses based on weights of organisms captured were undertaken (Figs. 50-51; Table 15). Given the results for fish and preliminary analyses for the invertebrates, depth-dependent differences in fishing efficiency were not further explored.

Across most of the smaller-bodied invertebrate taxa (e.g. shrimp, hermit crabs, bristle worms, whelks, brittle stars and the sea cauliflower, Gersemia rubiformis) the majority of the set pairs where only one vessel captured the taxa had the CCGS Alfred Needler reporting the null catch (Figs $50-51$ ). This is likely the result of a lower degree of vigilance aboard this vessel when sorting the catch of these smaller invertebrates by taxon. Indeed, there were 33 set pairs in which the CCGS Alfred Needler reported a catch of "miscellaneous unidentified remains", a catch-all category used to report quantities of unsorted non-fish catch, compared to 7 pairs aboard the CCGS Teleost (Fig. 51). Clearly, this systematic lower degree of sorting vigilance on one vessel across a large number of set pairs puts into question any vessel conversion factors derived for the taxa in question as the relative catchability of those taxa may not be properly reflected. Unfortunately, there is no reliable solution to deal with the bias, short of conducting additional comparative fishing sets with a stricter sorting regimen. To get an idea of the bias that this may have caused however, analyses were conducted including all set pairs as well as excluding those where one of the vessels reported both a null catch and a catch of "unidentified remains". It should be noted though that the latter approach has the disadvantage
that it may omit data that reflect a true vessel effect on catchability and does not fully account for the bias in cases where the catch of a taxon was partially sorted.

For the majority of taxa, no differences between vessels in relative catchability were found. The following text focuses on the exceptions.

There are over a dozen species of shrimp captured by the September survey. While there have been recent efforts to identify the catches of shrimp by species post survey using collected samples, species-level catch data were not available at the time these analyses were conducted. Analyses were therefore based on aggregate shrimp biomass. Significant vessel effects were found in two of the mixed-effects model analyses (Table 15). However, when sets with unsorted invertebrate catches were removed, the effect was no longer significant.

For three other taxa, bristle worms, Iceland scallop and brittle stars, though significant differences were found, there were very few sets in which both vessels captured the species. In the case of brittle stars, when both vessels reported a capture, the Needler recorded more biomass (suggesting a higher catchability), however there were also a large number of sets in which the Teleost recorded brittle stars but the Needler reported a null catch (suggesting a lower probability of capture). Because we cannot rule out that this apparent contradiction is not a spurious result of sorting vigilance, the current comparative fishing data for this and the other two species.

A significant vessel effect, robust to the removal of set pairs reporting unsorted catch, was established in fixed and mixed-effects analyses of empty mollusk shell catch (Table 15). The Teleost tended to catch approximately 1.8 times as many shells as the Needler. Though similar results were obtained for sea cauliflower, marine plants and algae, and woody debris, probabilities under the null hypothesis were close to the type-I error rate of $5 \%$. Application of conversion factors for these taxa is therefore unadvisable. Should ensuring the continuity of time series for these taxa become a priority (currently it is not), additional comparative fishing would be required to better estimate the conversion factors.

## 4. DISCUSSION AND RECOMMENDATIONS

### 4.1 ReCommendations -Comparative fishing experiment results

The following table provides a summary of results from the various analyses for all commercially important fish species and all other taxa for which statistically significant results were obtained in at least one of the analyses. Recommendations for the application of conversion factors, where appropriate, based on these results are also provided. Where results were quantitatively similar among like analyses, conversion factors estimated in the mixed effects model analyses including only set pairs with less than a $20 \%$ difference in tow distance should be used given the stronger conceptual basis for those inferences relative to the fixed-effects model equivalents. These results are found in Table 4 for length aggregated results and Table 8 for length-dependent results.

Consistent with the recommendations of Swain et al. (1995), recommendations are presented such that catches are calibrated to the current vessel (i.e., CCGS Teleost), as doing so necessitates no annual adjustments of catches as additional data are collected. (Note that standardizing to another vessel is trivial, see eqn. 1). In light of the non-significant results of analyses including a diel effect presented in section 3.4, these are not considered in this table.

Until sufficient data are collected to reliably estimate conversion factors for diel differences in catchability for the CCGS Teleost those derived for the CCGS Alfred Needler (documented in Benoît and Swain 2003b) should be used. There are about 40 fish and 7 invertebrate taxa for which diel differences in catchability to the CCGS Alfred Needler have been established.

| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
| Atlantic cod (10) | - The probability of the September only data under the null hypothesis of no vessel effect when outliers are excluded varies around a value of $5 \%$, depending on the analysis. A 10-20\% difference in fishing efficiency between vessels is estimated. <br> - $\quad P \sim 0.1$ in the fixed and mixed effects combined-surveys analyses, excluding outliers and sets with large differences in tow distance. <br> - No significant effect of length <br> - When the outlier set is removed, there is very little difference between vessels in the total length frequency (Fig. 54). Applying the marginally insignificant length-aggregated conversion factor does little to change the similarity in catch of the two vessels. | It is most reasonable to base the recommendation on analyses that exclude sets pairs with disparate towed distances for the reasons stated earlier. In those cases, the probability of the data under the null hypothesis are close to but above the somewhat arbitrary $5 \%$ level. This result, in light of the multiple statistical tests performed and the risk of committing Type-I errors, does not provide compelling evidence for a difference in catchability between the vessels. It is therefore recommended that no correction be applied for cod. However, a possible difference in catchability should be acknowledged as a potential source of bias in future assessments for this species. With a sufficient number of years following comparative fishing, it should be possible to test if this is the case by splitting the Alfred Needler and Teleost time series in the sequential population analysis and estimating the catchability of cod to the survey independently for each. |

\(\left.$$
\begin{array}{lll}\hline \text { White hake } \\
\text { (12) }\end{array}
$$ $$
\begin{array}{l}\text { Significant vessel effect in the } \\
\text { analysis of the September } \\
\text { data and in combined } \\
\text { analyses with the July data. }\end{array}
$$ \begin{array}{l}Based on the results of the September <br>
only and combined analyses, divide <br>
Needler, Hammond and E.E. Prince <br>

catches by 1.32 to obtain Teleost\end{array}\right\}\)| equivalents. Applying this conversion |
| :--- |
| No significant effect of any |
| covariates. | | factor improves the similarity between |
| :--- |
| vessels in the total length frequency |
| from the experiment (Fig. 54). |


| Redfish (23) | $\bullet$ | All analyses non-significant | No conversion factor for this species |
| :--- | :--- | :--- | :--- |
| Atlantic <br> halibut (30) | All analyses non-significant, <br> though sample sizes are <br> small | No conversion factor for this species |  |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
| Greenland halibut (31) | - All analyses non-significant | No conversion factor for this species |
| American plaice (40) | - Significant vessel effect in the analysis of the September data and in combined analyses with the July data (with and excluding outliers). <br> - Significant length-dependent vessel effect in the combined analysis with July data (fixed and mixed effects models), but not in the analysis of the September data only. | Given the observation of no lengthdependencies in the analysis of the September data despite a relatively large number of paired sets (90), recommend using a length-aggregated conversion only. <br> $b_{v}$ (mixed) Sept=1.18 <br> $b_{V}$ (mixed) Sept-Jul=1.13 <br> Based on the results of the September only and combined analyses, divide Needler and Hammond catches by 1.15 to get Teleost equivalents. Applying this conversion factor improves the similarity between vessels in the total length frequency from the experiment (Fig. 54). <br> Note that catches by the E.E. Prince are multiplied by 1.24 to get Teleost equivalents, given an existing conversion factor of 1.43 to yield Needler equivalents (see Benoît and Swain, 2003b). |
| Witch flounder (41) | - All analyses non-significant | No conversion factor for this species |
| Yellowtail flounder (42) | - Marginally significant vessel effect in the fixed-effects model analysis. Nonsignificant effect in the mixed effects model analyses. <br> - Non-significant effect of length in all analyses | No conversion factor for this species |
| Winter flounder (43) | - Marginally significant vessel effect in the fixed-effects model combined surveys analysis when outliers are removed. Non-significant effect in the mixed effects model analyses. <br> - Significant length-effect in fixed-effects model analysis | Given the non-significant results in mixed-effects model analyses, recommend using no conversion factors for this species |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
|  | excluding outliers ( $\mathrm{P}=0.01$ ). Non-significant effect in the mixed effects model analyses ( $\mathrm{P}>0.5$ ). |  |
| Atlantic herring (60) | - Significant vessel effect in the analysis of the September data in fixed and mixed effects models when outliers are removed. <br> - Vessel effects in the combined analysis with the July data were (marginally) non-significant, where tested. <br> - Marginally significant length dependent vessel effect in both fixed and mixed model analyses using the September data. This is not the case in the combined survey analyses excluding outliers. | Because catches of herring from the July survey occurred in deeper water than in September (catch-weighted mean depth of 142 vs .48 m ) there may be a difference in fish behaviour between the surveys and therefore recommend using the parameter estimates for September only. Both length-aggregated and lengthdependent correction factors improve somewhat the correspondence in total length frequencies from the Needler and Teleost in the September experiments (Fig. 54). Because they provide a comparable fit, recommend using the length aggregated conversion factor for reasons of parsimony. Therefore based on the results of the September only and combined analyses, divide Needler, Hammond and E.E. Prince catches by 1.52 to get Teleost equivalents. |


| Gaspereau / <br> Alewife (62) | Non-significant vessel effect <br> in the analysis of the <br> September-only and <br> combined surveys data. | Given the only marginally significant <br> length effect for the combined surveys <br> data but not the September-only data, <br> do not recommend applying any <br> conversion factors for gaspereau. |
| :--- | :--- | :--- |
|  | Marginally significant <br> (P=0.02) length-dependent <br> vessel effect using the <br> combined surveys data but <br> not the September-only data <br> in the mixed-effects model <br> analysis. Non-significant <br> results in the fixed-effects <br> model analyses | Furthermore, total length frequencies <br> for this species from the July <br> experiment (which caught the most <br> gaspereau) are not consistent with a <br> length effect (Fig. 14) |
|  |  |  |
| Capelin (64) | Non-significant vessel effect <br> in the analysis of the | Given the only marginally significant <br> length effect for the September data |
|  | September-only and <br> combined surveys data. | but not the combined surveys data in <br> the fixed effects models only, do not |
|  | Marginally significant |  |
| recommend applying any conversion |  |  |
| (P=0.048) length-dependent |  |  |
| factors for capelin. |  |  |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
|  | vessel effect using the September data, but not when combined with the July data in the fixed-effects model analysis. Nonsignificant results in the mixed-effects model analyses |  |
| Greenland $\operatorname{cod}(118)$ | - Significant vessel effect in the fixed and mixed effects model analyses of the September data. Not captured in the July survey <br> - Significant length-dependent vessel effect using the September data in fixed effects model analysis. | Greenland cod were captured by both vessels in 7 sets and by only one of the vessels in 19 sets. Fewer than 4 fish were generally captured at any one time. Futhermore, there is a possibility that smaller Greenland cod may have been confused with Atlantic cod, which would have a large impact on the estimated conversion factor for the former but not the latter given their relative abundances. Overall this results in very few data with which to meaningfully estimate a vessel effect, be it length aggregated or lengthdependent. <br> Recommend applying no conversion factors for this species. However, further comparative fishing would be beneficial in testing for a vessel effect. |
| Longhorn sculpin (300) | - Marginally significant vessel effect in the fixed and mixed effects model analyses of the September data. Due to a significant survey effect, analyses including the July data were not undertaken. <br> - Vessel effect was nonsignificant in an analysis of the July-only data $\begin{aligned} & \left(\beta_{v}=-0.1528 \pm 0.1003, \mathrm{n}=79\right. \\ & P=0.1317) \end{aligned}$ | Longhorn sculpin were captured by both vessels in 9 sets and by only one of the vessels in 10 sets. A total of fewer than ten individuals were captured in 15 of those relevant set pairs. Given such small sample sizes and catch amount, marginal significance in the September experiments and lack of significance in July, recommend applying no conversion factors for this species. However, further comparative fishing would be beneficial in testing for a vessel effect. |
| Mailed sculpin (304) | - Significant vessel effect in the analysis of the September data and in combined surveys mixed-effects model | Given non-significant length effects for the September-only analyses and the fact that the observed length range for this species is rather small ( $4-16 \mathrm{~cm}$ ), |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
|  | analyses. Non significant results in the fixed-effects model analyses. <br> - Significant length- dependent vessel effect in analyses based on the combined surveys data but not the September-only data. | do not recommend applying a lengthdependent correction. <br> Recommend applying the lengthaggregated conversion factor estimated in the mixed-effects model analysis. Based on those results divide Needler, Hammond and E.E. Prince catches by 1.37 to get Teleost equivalents. Applying this conversion factor improves the similarity between vessels in the total length frequency from the experiment (Fig. 54). |
| Spatulate sculpin (314) | - Non-significant vessel effect in either the fixed or mixed effects model analyses of the September-only or combined surveys data <br> - Significant length- dependent vessel effect in the mixedeffects model analysis of the September-only data ( $P=0.0194$ ), but not in the fixed-effects equivalent | Given that the length effect is not highly significant and the fact that the observed length range for this species is rather small $(4-12 \mathrm{~cm})$, do not recommend applying a lengthdependent correction. A lengthaggregated difference in catchability is not supported by the analyses; a correction is therefore not necessary. |
| Sea raven (320) | - Sea ravens were captured in only 14 relevant set pairs in September, with only five sets where both vessels captured the species. They were captured in over 60 sets in July. <br> - Significant vessel effect in the combined surveys ( $P \sim 0.02$ ) but not the September-only analyses. | The estimated correction factor for the length-aggregated combined-surveys vessel effect ( $b_{V}=0.76$ ) does not improve the correspondence of total length frequencies of the two vessels in the September experiments (Fig. 54). Given this result and the failure to find strong evidence for a difference in catchability, do not recommend applying any conversion factors. |
| Common alligatorfish (340) | - Non-significant vessel effect in the fixed and mixed effects model analyses of the September-only and combined surveys data. <br> - Significant length- dependent vessel effect in the fixed and mixed effects model analyses of the combined surveys data, but not the September- | Given the non-significant length effect in analyses of the September-only data, an observed length effect that differs somewhat between the two surveys (Fig. 31b,c) and the fact that the observed length range for this species is rather restricted ( $4-16 \mathrm{~cm}$ ), applying a conversion factor for this species for the September survey is not recommended. |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
| only data. |  |  |
| Sea poacher (350) | - Non-significant vessel effect in the fixed and mixed effects model analyses of the September-only and combined surveys data. <br> - Weak evidence for a lengthdependent vessel effect in the mixed-effects model analyses of the Septemberonly data. Non-significant results for the fixed-effects model analyses. | Given the non-significant lengthaggregated effect, only weak statistical evidence for a length-dependent effect and total length frequencies that do not suggest a large lengthdependency (Fig. 32b,c), do not recommend applying a conversion factor for this species. |
| Sand lance (610) | - Sandlance were caught in only 12 relevant set pairs in the September experiments. Catches were more frequent in the July experiments. <br> - Marginally significant vessel effect in combined surveys fixed-effects analysis. <br> - Significant length-dependent effects in fixed-effects and mixed-effects model analyses, with the Teleost catching relatively more small and fewer large fish than the Needler. | Recommend using a length-dependent correction factor for this species. Given the small number of sets from September, recommend using the conversion factor from the combined analysis with the July data <br> To obtain Teleost equivalent catches, divide Needler, Hammond and E.E. Prince catches by: $\exp \left(-1.3274+0.1084^{*}\right.$ length $)$. Applying this correction improves the similarity between vessels in the total length frequencies from the experiment (Fig. 54). |
| Laval's eelpout (620) | Marginally significant ( $P=0.048$ ) vessel effect in the mixed-effects model analysis of the September data only when all relevant set pairs, including those with large differences in towed distance, are included. | Given that results are not significant when set pairs including disparate towed distances are removed, do not recommend applying a conversion factor for this species for the September survey. |
| Snakeblenny (622) | - Marginally significant vessel effect in the mixed-effects model analysis of the September data <br> - Significant vessel effect in the fixed and mixed-effects model analyses of the combined surveys data. <br> - No significant length- | Based on the results of the combined surveys mixed-effects model analyses, divide Needler, Hammond and E.E. Prince catches by 1.96 to get Teleost equivalents. Applying this conversion factor improves the similarity between vessels in the total length frequency from the experiment (Fig. 54). |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
| dependent effects. |  |  |
| Daubed shanny (623) | - Significant length-aggregated or length-dependent differences in catchability between vessels in all analyses. | Despite the relatively low amount of length variation in this species (8-15 cm ), patterns in length dependent catchability are remarkably similar between the September and July comparative fishing experiments (Fig 38b,c). The Teleost catches relatively more fish smaller than $10-11 \mathrm{~cm}$, and fewer fish larger than that size. <br> Recommend using a length-dependent conversion factor for this species based on the combined September and July data. Therefore divide Needler, Hammond and E.E. Prince catches by $\exp \left(-1.7769+\left(0.1519^{*}\right.\right.$ length $)$ to get Teleost equivalents. Both this correction and the one from the lengthaggregated analysis improve the similarity between vessels in the total length frequencies from the experiment, though the former may provide a marginally better fit (Fig. 54). |
| Vahl's eelpout (647) | - Non-significant vessel effect in the fixed and mixed-effects model analyses of the September data <br> - Marginally significant ( $P \sim 0.03$ ) vessel effect in the fixed effects model analysis of the combined SeptemberJuly data. Not so for the mixed effects model. | Given that the vessel effect is not significant for the September data alone and is only marginally significant in the fixed-effects analyses of the combined surveys applying a conversion factor for this species is not recommended. |
| Atlantic hookear sculpin (880) | - Significant vessel effect in the mixed-effects model analyses of the September- only and combined surveys data, though there were only three set pairs in the July survey in which both vessels captured the species. | There is a distinct possibility of confusion of this species with Arctic hookear sculpin. Prior to 2004, the two species were not identified separately during annual surveys. The confusion may explain a number of cases where one vessel apparently captured one of the hookear sculpin species, while the other did not (Figs. 27 \& 28). When the species are combined, a significant |


| Species <br> (code) | Recommendation |
| :--- | :--- |
|  | vessel effect in the mixed-effects <br> model analyses of the September only <br> data is found $(P=0.0033)$. However, <br> the estimated correction factor <br> $\left(b_{b}=0.42\right)$ worsens the correspondence <br> of total length frequencies of the two <br> vessels in the September experiments <br> (Fig. 54$)$. In light of this result and in <br> the absence of more certainty in the <br> taxonomic identification of the species <br> applying a common correction factor <br> for either Atlantic (code 880) or Arctic <br>  <br> (306) hookear sculpins is not <br> recommended. |
|  |  |


| Atlantic rock <br> crab (2513) | Significant $(P=0.0142)$ vessel <br> effect in the mixed-effects <br> model analyses of the <br> combined surveys data, but <br> not in the September-only <br> analyses. Non-significant <br> fixed-effects model analysis <br> results. |
| :--- | :--- |

Given that evidence for a vessel effect isn't overly strong and that sample size is relatively small, applying a conversion factor for this species is not recommended. However, further comparative fishing would clearly be beneficial in testing for a vessel effect.

| Northern stone crab (2523) | - Non-significant vessel effect in the analyses of the September data. <br> - Significant ( $P=0.0122$ ) vessel effect in the mixed-effects model combined-survey analyses. | Given that evidence for a vessel effect isn't overly strong and that sample size is relatively small, do not recommend applying a conversion factor for this species. However, further comparative fishing would clearly be beneficial in testing for a vessel effect. |
| :---: | :---: | :---: |


| Snow crab <br> $(2526)$ | Marginally non-significant <br> (P~0.06) vessel effects in the <br> fixed and mixed-effects | Given a non-significant vessel effect <br> despite a relatively large number of <br> paired sets in September ( $n=85$ ) and a |
| :--- | :--- | :--- |
|  | model analyses of the <br> September data when set <br> pairs with disparate distances <br> towed are removed. | contradictory direction of effect with <br> the July experiments, applying a <br> conversion factor for this species is not |
|  | recommended. |  |
|  | Marginally significant <br> (P=0.049) effect of length in <br> mixed-effects model analysis <br> including all relevant set |  |
|  | pairs. When set pairs with <br> disparate distances towed <br> are removed, the probability <br> of the data under the null |  |
|  |  |  |


| Species (code) | Observed effects | Recommendation |
| :---: | :---: | :---: |
|  | hypothesis increases ( $P=0.168$ ). <br> - Analyses of the combined September and July data were not undertaken because of a significant survey effect. Indeed the vessel effect from the July survey is in the opposite direction as in the September experiment (for July, the Needler catches more crab (Fig. 45a)). |  |
| Hyas araneus (2527) | - Significant vessel effect in the fixed and mixed-effects model analyses of the September data. <br> - Analyses of the combined September and July data were not undertaken because of a significant survey effect. <br> - Marginally significant ( $P \sim 0.02$ ) length-dependent vessel effect based on the September data. | Length-dependent and lengthaggregated conversion factors do not appear to significantly improve the correspondence between Needler and Teleost total length frequencies, across the size range for this species (Fig. 54). As a result, do not recommend applying a conversion factor for this species. |

As stated previously, comparative fishing using the Western IIA has not taken place between the CCGS Templeman and either the CCGS Alfred Needler or the CCGS Teleost. There is therefore no firm basis to gauge whether catchability of certain species to the survey was different in 2003, when the former vessel undertook the survey of the southern Gulf of St. Lawrence. Although some differences exist, the CCGS Templeman and the CCGS Alfred Needler are considered sister ships and are therefore expected to have comparable fishing efficiencies when using the same fishing gear. This will have to be assumed until comparative fishing with the CCGS Templeman is undertaken. Consequently, conversion factors used when correcting for differences in relative catchability between the CCGS Alfred Needler and the CCGS Teleost should be applied to the survey catches in 2003. This also includes CCGS Alfred Needler specific corrections for diel differences in catchability (Benoît and Swain, 2003b).

### 4.2 RECOMMENDATIONS - SURVEY AREA COVERED AND REPEAT SETS (2003-2005)

As a result of a delay in commencing the survey and lost time due to inclement weather and search-and-rescue activities, three strata (402, 425 and 436) were sampled with only one fishing set and two strata ( 438 and 439) were not sampled at all during the 2003 survey. The missed strata pose a particular problem because they are in deep waters that comprise a small overall proportion of the total survey area yet are important areas for many deep-water species. Ignoring these omissions (i.e., implicitly assume that the average catch in the missed strata was equal to the average over the remainder of the survey area) will result in
considerable bias in the estimated abundance of those species whose distribution is largely restricted to deep water in September. As a result, the best approach to use in calculating abundance indices for 2003 is to fill in the missing data cells using a multiplicative analysis with year and stratum as model terms (e.g., Swain et al. 1998). This approach assumes that there is no year $x$ stratum interaction (i.e., no change in distribution between years).

Although the CCGS Teleost completely sampled the survey area in 2004, the CCGS Alfred Needler also fished a small number of randomly-selected stations not sampled by the former vessel. Once the necessary conversion factors are applied, the sampling results from these stations can contribute to the calculation of the annual abundance index. In 2005, inclement weather prevented either of the vessels from completely sampling the survey area, though they accomplished this jointly. In this case, the combination of CCGS Teleost and corrected CCGS Alfred Needler sampling results is required to calculate the annual index.

The 2004-2005 comparative fishing experiments resulted in several paired (repeat) sets at various locations in the Gulf during the annual survey (Fig. 1). Because both sets in a pair can contribute to estimating the abundance of species, though not with the same weight as unique sets at a given location, catches from repeat paired sets (Table 1) should be averaged prior to calculating stratified means and variances. Furthermore, during the 2004 survey the two vessels both sampled a number of the same survey stations, though the fishing was not simultaneous (Table 16). Although these sets were not included in the comparative fishing analyses because of the differences between vessels in the timing of fishing, catches in each repeat set pair should be averaged.
4.3 RECOMMENDATIONS - TAXONOMIC IDENTIFICATION

Although the survey protocol since 1971 has been to sort (and record) catches of fish by species, identification at sea is often problematic for some genera or species groups. In some cases, practical guides for use at sea have been developed in recent years to aid in taxonomic identification, though past survey records may be unreliable. The purpose of this section of the report is to document these problematic species or groups and to provide recommendations for dealing with potentially unreliable species accounts in past survey records.

Genus Alosa (gaspereau): No attempts are made to differentiate Alosa pseudoharengus (alewife) and $A$. aestivalis (blueback herring), and consequently both fall under the collective name of gaspereau (species code 62).

Genus Artediellus (hookear sculpins): Two species of this genus occur in the survey: Atlantic hookear (Artediellus atlanticus, species code 880) and Arctic hookear (Artediellus uncinatus, code 306). It is unlikely that these two species were properly separated in past surveys. While efforts have increased since 2004 to separate them, it is not presently clear that it is being done reliably by all survey staff. As a result, for surveys in the past and for the foreseeable future, catches of these species should be grouped at the genus level (code 323) prior to analysis.

Genus Eumicrotremus (lumpsuckers): The lumpsucker commonly captured in the survey is the Atlantic spiny lumpsucker (Eumicrotremus spinosus, species code 502). Three records of the rare Arctic species, leatherfin lumpsucker (Eumicrotremus derjugini, species code 509), in past surveys are questionable as it is unlikely that survey staff would have been properly able to distinguish it from E. spinosus (see Scott and Scott, 1988). Consequently, these records should be grouped with $E$. spinosus.

Genus Liparis (seasnails): While attempts are made to identify these fish to the species level, it is felt that this was done inconsistently prior to 2004 and the reliability of some identifications is questionable. The majority of seasnails captured by the survey are dusky seasnails (L. gibbus, species code 512) although these have on occasion mistakenly been called striped seasnail (L. liparis, species code 504) because of variations in color patterns within the species (Scott and Scott, 1988). Records prior to 2004 of Atlantic seasnail (L. atlanticus, species code 503), gelatinous seasnail (L. fabricii, species code 505), Greenland seasnail (L. tunicatus, species code 506), Gulf seasnail (L. coheni, species code 513), and Paraliparis calidus (code 868) have not been verified and are questionable given the limited tools that survey staff had at their disposal for proper identifications. New visual guides available on the surveys as of 2004 have considerably improved the taxonomic identification of this genus. Consequently, records of these species should be grouped as Liparis sp. (code 500) at least prior to 2004.

Genus Leucoraja (skates): The only species of this genus that occurs in the southern Gulf is winter skate (Leucoraja ocellata, species code 204), therefore survey records of little skate (L. erinacea, code 203) most certainly represent a misidentification of the former species (McEachran and Martin, 1978; McEachran and Musick, 1975).

Genus Lycodes (eelpouts): While attempts are made to identify these fish to the species level, it is felt that this was done inconsistently prior to 2003 and the reliability of the identification is questionable. Although the majority of eelpouts in the southern Gulf survey are Laval's eelpout (L. lavalaei, species code 620) and Vahl's eelpout (L. vahlii, species code 647), there are records prior to 2003 of Arctic eelpout (L. reticulatus, species code 641), pale eelpout (L. pallidus, species code 627), polar eelpout (L. polaris, species code 628), Newfoundland eelpout (L. terraenova, species code 619) and Vachon's eelpout (L. esmarki, species code 643). New visual guides available on the surveys as of 2003 have considerably improved the taxonomic identification of this genus, particularly for Vahl's and Laval's eelpouts. Consequently, grouping records of these species at the genus level (code 642) is recommended, at least for records prior to 2003.

Genus Pholis (gunnels): Although gunnels are rarely captured in the survey, reports of the banded gunnel (Pholis fasciata, code 633) were likely the rock gunnel (P. gunnellus, code 621), as the former species occurs only in very shallow water, whereas the later may occasionally be captured in deeper waters (Scott and Scott, 1988).

Genus Sebastes (redfish): No attempts are made to differentiate redfishes to species in the southern Gulf surveys, although these would mainly be S. fasciatus and S. mentella. Past survey records include a small number of accounts of blackbelly rosefish (Helicolenus dactylopterus dactylopterus, species code 123), a species with a more southerly distribution (Scott and Scott, 1988). Given that this species would be unlikely to occur in the southern Gulf and that it is also unlikely that survey staff would have undertaken the meristic counts required to differentiate it from the Sebastes sp. (Scott and Scott, 1988), these accounts should probably be treated as being redfish (Sebastes sp., code 23).

Genus Urophycis (hake): The only species of this genus that occurs in the southern Gulf is white hake (Urophycis tenuis, species code 12). Survey records of red hake (U. chuss, code 13) most certainly represent a misidentification of the former species (Scott and Scott 1988; T. Hurlbut, DFO Gulf region, personal communication).

Sub-Family Anoplagoninae (alligatorfish): Two species of this genus occur in the survey: common alligatorfish (Aspidophoroides monopterygius, species code 340) and arctic alligatorfish (Ulcina olrikii, code 341). It is unlikely that these two species were properly separated in past surveys, though the former species was likely the more common of the two.

While efforts have increased since 2005 to separate them, it is not presently clear that it is done reliably by all survey staff. As a result, for surveys in the past and for the foreseeable future, catches of these species should be grouped at the sub-Family level prior to analysis.

Family Cottidae (sculpins): Some taxa in this family have posed problems when it comes to proper taxonomic identification. The following species have been recorded in the survey but are dubious because these species occupy habitats that are not sampled by the survey:

- Grubby sculpin (Myoxocephalus aenus, code 303): this estuarine species could easily have been improperly identified longhorn sculpin ( $M$. octodecemspinosus, code 300) given their general morphological similarity. It is not impossible however that grubby sculpin may have been captured in small numbers in the survey.
- Twohorn sculpin (Icelus bicornis, code 313): the 35 records of this species in the survey data probably represent a misidentification of spatulate sculpin (I. spatula, species code 314), which occurs generally in the area.
- Pallid sculpin (Cottunculus thomsoni, code 308): this species has not been confirmed in the Gulf of St. Lawrence (Scott and Scott, 1988). The single record of its occurrence in the September survey may be a misidentification of the polar sculpin (C. microps, code 307), whose occurrence in the Gulf has been confirmed. I is also captured occasionally in the survey.

Family Lumpenidae (shannies): One species in particular in this family has posed problems in the survey. Slender eelblenny (Lumpenus fabricii, species code 631) have been reported several times in the southern Gulf survey database. A rigorous examination of a large number of individuals tentatively identified as slender eelblenny in 2004 and 2005 found that all individuals were incorrectly identified. The majority of the individuals were actually the much more common daubed shanny (Leptoclinus maculatus, species code 623), followed by a smaller number of individuals being snakeblenny (Lumpenus lampretaeformis, species code 622) and less than one percent being stout eelblennies (Lumpenus medius, species code 632). Based on length frequencies, any fish over 18 cm identified as slender eelblenny in past survey records can reasonably be considered a snakeblenny (this covers about half of the 662 individuals identified as slender eelblennies from 1971-2003). Though the majority of fish below 18 cm are likely daubed shanny, they cannot be reliably attributed to that species, snakeblenny or possibly stout eelblenny. These fish ( $n=335$ ) should therefore only be treated at the family level, which does not have any large implications for the other species as they are considerably more numerous: daubed shanny ( 25,000 individuals recorded since 1971), snakeblenny (about 3,000 individuals) and stout eelblenny (about 6,000).

Family Paralepididae (barracudinas): Barracudinas should be grouped to the family level (Paralepididae, code 713) because while the vast majority of instances are of white barracudina (Arctozenus risso, species code 712), there are a few unconfirmed survey records of Paralepis coregonoides (code 674).

## 5. ACKNOWLEDGEMENTS

I am very grateful for the help on the mixed-effects model analyses provided by J. Dowden and N. Cadigan. I also thank N. Cadigan, T. Hurlbut and T. Surette for their careful reviews of the manuscript

## 6. REFERENCES

Benoît, H.P., M.-J. Abgrall, and D.P. Swain. 2003a. An assessment of the general status of marine and diadromous fish species in the southern Gulf of St. Lawrence based on annual bottom-trawl surveys (1971-2002). Can. Tech. Rep. Fish. Aquat. Sci. 2472: 183 p.

Benoît, H. P., E.D. Darbyson, and D.P. Swain. 2003b. An atlas of the geographic distribution of marine fish and invertebrates in the southern Gulf of St. Lawrence based on annual bottom trawl surveys (1971-2002). Can. Data Rep. Fish. Aquat. Sci. 1112: 185 p.

Benoît, H.P. and D.P. Swain. 2003a. Accounting for length and depth-dependent diel variation in catchability of fish and invertebrates in an annual bottom-trawl survey. ICES J. Mar. Sci. 60: 1297-1316.

Benoît, H.P. and D.P. Swain. 2003b. Standardizing the southern Gulf of St. Lawrence bottomtrawl survey time series: adjusting for changes in research vessel, gear and survey protocol. Can. Tech. Rep. Fish. Aquat. Sci. no. 2505: iv + 95 pp.

Cadigan, N., S.J. Walsh and W. Brodie. 2006. Relative efficiency of the Wilfred Templeman and Alfred Needler research vessels using a Campelen 1800 shrimp trawl in NAFO Subdivision 3Ps and Divisions 3LN. DFO Can. Sci. Adv. Sec. Res. Doc. 2006/085.

Casey, J.M., and R.A. Myers. 1998. Diel variation in trawl catchability: is it as clear as day and night? Can. J. Fish. Aquat. Sci. 55:2329-2340.

Hjellvik, V., O.R. Godø, and D. Tjøstheim, D. 2002. Diurnal variation in bottom-trawl survey catches: does it pay to adjust? Can. J. Fish. Aquat. Sci. 59:33-48.

Hurlbut, T., and D. Clay. 1990. Protocols for research vessel cruises within the Gulf Region (demersal fish) (1970-1987). Can. Manuscr. Rep. Fish. Aquat. Sci. 2082: 143 p.

Manly, B.F. 1991. Randomization and Monte Carlo methods in biology. Chapman and Hall, NY.

McCullagh, P., and J.A. Nelder. 1989. Generalized linear models, $2^{\text {nd }}$ ed. Chapman and Hall, London.

McEachran, J.D., and C.O. Martin. 1978. Interrelationships and subgeneric classification of Raja erinacea and $R$. ocellata based on claspers, neurocrania and pelvic girdles (Pisces:Rajidae). Copeia 4: 593-601.

McEachran, J.D., and J.A. Musick. 1975. Distribution and relative abundance of skates (Pisces: Rajidae) which occur between Nova Scotia and Cape Hatteras. Fish. Bull. 73: 110136.

Millar, R.B., M.K. Broadhurst, and W.G. Macbeth. 2004. Modelling betweem-haul variability in size-selectivity of trawls. Fish. Res. 67: 171-181.

Nielsen, G.A. 1994. Comparison of the fishing efficiency of research vessels used in the southern Gulf of St. Lawrence groundfish surveys from 1971 to 1992. Can. Tech. Rep. Fish. Aquat. Sci. 1952.

Pelletier, D. 1998. Intercalibration of research survey vessels in fisheries: a review and an application. Can. J. Fish. Aquat. Sci. 55: 2672-2690.

Pinhero, J.C., and D.M. Bates. 2000. Mixed-effects models in S and S-PLUS. Springer-Verlag, N.Y.

Rice, W.R. 1989. Analyzing tables of statistical tests. Evolution 43: 223-225.
SAS Institute Inc. 2005. The GLIMMIX procedure, Nov. 2005. Cary, NC: SAS Institute Inc.
Scott, W.B. and M.G. Scott. 1988. Atlantic fishes of Canada. Can. Bull. Fish. Aquat. Sci. 219. 731 p.

Swain, D. P., G. A. Nielsen, and D. E. McKay. 1995. Incorporating depth-dependent differences in fishing efficiency among vessels in the research survey time series for Atlantic cod (Gadus morhua) in the southern Gulf of St. Lawrence. Can. MS Rept. Fish. Aquat. Sci. 2317: 20 p .

Swain, D.P., G.A. Poirier, and R. Morin. 1998. Status of witch flounder in NAFO Divisions 4RST, January 1998. DFO Can. Stock Assess. Res. Doc. 1998/004.

Swain, D.P., J.E. Simon, L.E. Harris and H.P. Benoît. 2006. Recovery potential assessment of 4T and 4VW winter skate (Leucoraja ocellata): biology, current status and threats. DFO Can. Sci. Advis. Sec. Res. Doc. 2006/003.

Table 1. Timing, location, depth and average distance between vessels for each set from the 2004 and 2005 comparative fishing experiments in the southern Gulf of St. Lawrence. The second-last column is the percentage difference in towing distance between vessels (the target fishing procedure was a 30 min . tow at 3.5 knots, yielding a 1.75 n . mile tow).

| Year | Day | Time | Stratum | Set | Latitude (decimal degrees) | Longitude (decimal degrees) | Depth (m) | Distance between vessels (km) | \% difference in distance towed between vessels | Teleost fishes further 1=yes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 15 | 7:02 | 423 | 83 | 47.7842 | 61.8802 | 48 | 1.1 | 4 | 1 |
| 2004 | 15 | 9:14 | 423 | 84 | 47.8690 | 62.0620 | 53 | 0.7 | 24 | 0 |
| 2004 | 15 | 16:05 | 423 | 86 | 47.7225 | 62.5843 | 76 | 1.1 | 2 | 0 |
| 2004 | 15 | 18:34 | 423 | 87 | 47.6363 | 62.4697 | 70 | 1.0 | 1 | 0 |
| 2004 | 15 | 20:27 | 423 | 88 | 47.6368 | 62.2135 | 49 | 1.3 | 15 | 0 |
| 2004 | 15 | 22:56 | 423 | 89 | 47.4738 | 62.4065 | 64 | 0.6 | 1 | 1 |
| 2004 | 16 | 1:22 | 423 | 90 | 47.4155 | 62.6730 | 63 | 0.9 | 15 | 0 |
| 2004 | 16 | 4:20 | 423 | 91 | 47.4078 | 62.9565 | 52 | 0.9 | 2 | 0 |
| 2004 | 16 | 6:46 | 423 | 92 | 47.6390 | 62.8368 | 59 | 1.1 | 2 | 0 |
| 2004 | 16 | 8:46 | 423 | 93 | 47.8357 | 62.7535 | 68 | 0.5 | 0 | 0 |
| 2004 | 16 | 11:12 | 423 | 94 | 47.8532 | 63.0272 | 71 | 0.8 | 0 | 0 |
| 2005 | 11 | 13:57 | 403 | 30 | 45.6922 | 61.7633 | 23 | 0.7 | 5 | 0 |
| 2005 | 11 | 16:42 | 403 | 31 | 45.7913 | 61.7838 | 22 | 0.7 | 0 | 1 |
| 2005 | 11 | 19:36 | 403 | 32 | 45.7712 | 61.8607 | 22 | 0.7 | 0 | 0 |
| 2005 | 11 | 23:58 | 433 | 33 | 45.9372 | 61.6870 | 39 | 0.9 | 0 | 0 |
| 2005 | 12 | 2:37 | 433 | 34 | 46.2205 | 61.5362 | 50 | 1.0 | 2 | 0 |
| 2005 | 12 | 5:39 | 433 | 35 | 46.2175 | 61.7343 | 54 | 0.7 | 1 | 0 |
| 2005 | 12 | 8:51 | 434 | 36 | 46.5273 | 61.5057 | 60 | 1.2 | 2 | 0 |
| 2005 | 12 | 10:59 | 434 | 37 | 46.7003 | 61.3380 | 68 | 1.0 | 5 | 0 |
| 2005 | 12 | 12:58 | 434 | 38 | 46.7275 | 61.1073 | 76 | 0.9 | 1 | 1 |
| 2005 | 12 | 15:35 | 437 | 39 | 46.7487 | 61.1128 | 116 | 0.8 | 0 | 0 |
| 2005 | 13 | 1:58 | 434 | 40 | 46.8858 | 61.0723 | 87 | 0.9 | 0 | 0 |
| 2005 | 13 | 4:38 | 434 | 41 | 46.8913 | 61.3785 | 63 | 1.1 | 3 | 0 |
| 2005 | 13 | 6:42 | 434 | 42 | 46.9600 | 61.3288 | 50 | 0.9 | 2 | 0 |
| 2005 | 13 | 8:30 | 434 | 43 | 47.0372 | 61.2355 | 50 | 1.2 | 0 | 0 |
| 2005 | 13 | 22:06 | 431 | 49 | 46.9705 | 61.7802 | 45 | 0.5 | 16 | 0 |
| 2005 | 14 | 2:32 | 431 | 51 | 46.8850 | 61.9815 | 56 | 0.8 | 28 | 1 |
| 2005 | 14 | 9:15 | 423 | 54 | 47.2268 | 63.0767 | 63 | 1.0 | 39 | 1 |
| 2005 | 14 | 11:19 | 429 | 55 | 47.0642 | 63.1743 | 56 | 0.8 | 18 | 0 |
| 2005 | 14 | 13:48 | 429 | 56 | 46.9122 | 63.1337 | 57 | 0.8 | 8 | 0 |
| 2005 | 14 | 16:11 | 429 | 57 | 46.7593 | 63.3505 | 47 | 1.2 | 9 | 1 |
| 2005 | 14 | 18:13 | 401 | 58 | 46.6663 | 63.6162 | 28 | 0.7 | 15 | 0 |
| 2005 | 14 | 20:27 | 401 | 59 | 46.7142 | 63.7755 | 26 | 1.8 | 8 | 0 |
| 2005 | 14 | 23:05 | 429 | 60 | 46.9677 | 63.7357 | 45 | 0.6 | 6 | 0 |
| 2005 | 15 | 1:04 | 429 | 61 | 47.1202 | 63.6917 | 56 | 1.1 | 0 | 1 |
| 2005 | 15 | 3:20 | 422 | 62 | 47.2185 | 64.1170 | 40 | 0.7 | 5 | 1 |
| 2005 | 15 | 12:55 | 420 | 67 | 47.1863 | 64.6885 | 26 | 0.5 | 1 | 1 |
| 2005 | 15 | 14:10 | 420 | 68 | 47.2542 | 64.5577 | 35 | 0.8 | 3 | 1 |
| 2005 | 15 | 15:53 | 422 | 69 | 47.3768 | 64.3753 | 52 | 0.6 | 2 | 1 |


| Year | Day | Time | Stratum | Set | Latitude (decimal degrees) | Longitude (decimal degrees) | Depth (m) | Distance between vessels (km) | \% difference in distance towed between vessels | Teleost fishes further 1=yes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 15 | 18:35 | 423 | 70 | 47.3872 | 63.8015 | 65 | 0.9 | 8 | 0 |
| 2005 | 15 | 20:01 | 423 | 71 | 47.3077 | 63.6743 | 64 | 0.9 | 1 | 1 |
| 2005 | 16 | 0:00 | 423 | 73 | 47.2942 | 63.4278 | 57 | 0.4 | 6 | 1 |
| 2005 | 16 | 1:19 | 423 | 74 | 47.4198 | 63.5920 | 62 | 0.9 | 10 | 1 |
| 2005 | 16 | 2:35 | 423 | 75 | 47.4845 | 63.6150 | 69 | 1.0 | 9 | 1 |
| 2005 | 16 | 3:57 | 423 | 76 | 47.5080 | 63.6858 | 67 | 0.8 | 14 | 1 |
| 2005 | 16 | 5:21 | 423 | 77 | 47.5433 | 63.7958 | 67 | 0.7 | 14 | 1 |
| 2005 | 16 | 7:13 | 422 | 78 | 47.5197 | 64.0585 | 40 | 1.0 | 13 | 0 |
| 2005 | 16 | 9:24 | 422 | 79 | 47.5875 | 64.1490 | 65 | 0.8 | 0 | 0 |
| 2005 | 16 | 11:38 | 420 | 80 | 47.6517 | 64.4802 | 29 | 0.8 | 2 | 0 |
| 2005 | 16 | 14:03 | 422 | 81 | 47.6385 | 63.9907 | 50 | 0.6 | 10 | 1 |
| 2005 | 16 | 15:19 | 422 | 82 | 47.7028 | 63.9398 | 63 | 1.5 | 56 | 1 |
| 2005 | 16 | 20:25 | 423 | 84 | 47.6122 | 63.3962 | 80 | 0.9 | 0 | 1 |
| 2005 | 16 | 23:19 | 423 | 85 | 47.6970 | 62.9302 | 59 | 1.0 | 10 | 0 |
| 2005 | 17 | 00:54 | 423 | 86 | 47.7862 | 62.7927 | 64 | 0.6 | 12 | 0 |
| 2005 | 17 | 2:41 | 423 | 87 | 47.6578 | 62.7147 | 66 | 0.9 | 19 | 0 |
| 2005 | 17 | 4:32 | 423 | 88 | 47.5453 | 62.5223 | 75 | 1.0 | 9 | 0 |
| 2005 | 17 | 6:20 | 423 | 89 | 47.4128 | 62.4635 | 68 | 1.1 | 0 | 0 |
| 2005 | 17 | 9:15 | 428 | 90 | 47.5490 | 62.0545 | 34 | 0.7 | 39 | 1 |
| 2005 | 17 | 11:30 | 423 | 91 | 47.7870 | 62.4198 | 71 | 0.6 | 1 | 1 |
| 2005 | 17 | 14:37 | 424 | 92 | 48.0333 | 62.8565 | 69 | 1.1 | 29 | 1 |
| 2005 | 17 | 17:57 | 424 | 93 | 47.9322 | 63.5912 | 69 | 1.1 | 3 | 0 |
| 2005 | 17 | 19:41 | 422 | 94 | 47.9627 | 63.8447 | 90 | 0.8 | 2 | 0 |
| 2005 | 17 | 21:13 | 422 | 95 | 47.8955 | 64.0060 | 86 | 0.9 | 1 | 0 |
| 2005 | 17 | 23:05 | 420 | 96 | 48.0010 | 64.2552 | 26 | 0.9 | 10 | 0 |
| 2005 | 18 | 1:05 | 420 | 97 | 48.0432 | 64.2928 | 22 | 1.0 | 4 | 1 |
| 2005 | 18 | 2:46 | 417 | 98 | 48.1792 | 64.2690 | 49 | 0.6 | 31 | 0 |
| 2005 | 18 | 7:16 | 416 | 100 | 48.3077 | 63.9370 | 95 | 0.9 | 14 | 1 |
| 2005 | 18 | 10:17 | 416 | 101 | 48.5343 | 64.0887 | 102 | 0.6 | 3 | 0 |
| 2005 | 18 | 11:56 | 416 | 102 | 48.5252 | 63.9875 | 121 | 0.8 | 7 | 1 |
| 2005 | 18 | 14:25 | 416 | 103 | 48.4217 | 63.7667 | 123 | 0.5 | 5 | 1 |
| 2005 | 18 | 16:29 | 416 | 104 | 48.2727 | 63.5557 | 95 | 1.1 | 2 | 1 |
| 2005 | 18 | 19:41 | 417 | 105 | 48.2300 | 64.0708 | 67 | 0.8 | 1 | 0 |
| 2005 | 18 | 22:24 | 417 | 106 | 48.2642 | 64.5428 | 97 | 0.8 | 2 | 0 |
| 2005 | 19 | 21:08 | 419 | 108 | 48.0572 | 65.8047 | 30 | 0.9 | 3 | 0 |
| 2005 | 20 | 1:28 | 419 | 109 | 47.7667 | 65.6167 | 23 | 0.8 | 4 | 1 |
| 2005 | 20 | 5:44 | 419 | 110 | 47.8958 | 65.4877 | 59 | 0.8 | 2 | 0 |
| 2005 | 20 | 9:05 | 418 | 111 | 47.8940 | 65.0725 | 68 | 1.0 | 4 | 0 |
| 2005 | 20 | 10:50 | 418 | 112 | 47.9745 | 64.8737 | 66 | 0.8 | 3 | 0 |
| 2005 | 20 | 17:42 | 422 | 114 | 47.9967 | 63.9547 | 89 | 0.9 | 2 | 0 |
| 2005 | 20 | 22:31 | 416 | 116 | 48.4433 | 63.5462 | 111 | 0.9 | 1 | 0 |
| 2005 | 21 | 00:43 | 416 | 117 | 48.5782 | 63.5893 | 117 | 0.8 | 6 | 1 |
| 2005 | 21 | 6:57 | 416 | 118 | 48.6718 | 63.8083 | 148 | 0.8 | 1 | 0 |
| 2005 | 22 | 23:48 | 415 | 122 | 48.7640 | 63.2078 | 302 | 0.8 |  | 0 |
| 2005 | 23 | 2:09 | 425 | 123 | 48.5745 | 63.0438 | 285 | 0.8 | 0 | 1 |


| Year | Day | Time | Stratum | Set | Latitude <br> (decimal <br> degrees) | Longitude <br> (decimal <br> degrees) | Depth <br> $(m)$ | Distance <br> between <br> vessels <br> (km) difference <br> in distance <br> towed <br> between <br> vessels | Teleost <br> fishes <br> further <br> $1=y e s ~$ |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  | 0 |
| 2005 | 23 | $4: 18$ | 425 | 124 | 48.5545 | 63.0725 | 251 | 0.9 | 0 | 0 |
| 2005 | 23 | $9: 23$ | 424 | 126 | 48.2820 | 63.0512 | 68 | 0.9 | 2 | 0 |
| 2005 | 23 | $11: 06$ | 424 | 127 | 48.1578 | 63.1733 | 70 | 0.8 | 0 | 0 |
| 2005 | 23 | $15: 52$ | 424 | 129 | 48.1125 | 63.1155 | 66 | 0.7 | 2 | 0 |
| 2005 | 23 | $21: 20$ | 427 | 131 | 48.0987 | 62.3523 | 67 | 0.8 | 1 | 0 |
| 2005 | 24 | $3: 35$ | 425 | 134 | 48.3905 | 62.1633 | 312 | 1.0 | 9 | 0 |
| 2005 | 25 | $17: 04$ | 425 | 145 | 48.1463 | 61.2392 | 329 | 0.9 | 1 | 0 |
| 2005 | 25 | $20: 19$ | 427 | 146 | 47.9787 | 61.3633 | 61 | 0.7 | 6 | 0 |
| 2005 | 25 | $23: 22$ | 439 | 147 | 47.8710 | 60.7380 | 261 | 0.7 | 2 | 1 |
| 2005 | 26 | $1: 41$ | 438 | 148 | 47.6782 | 60.5683 | 126 | 1.2 | 1 | 0 |
| 2005 | 26 | $4: 13$ | 438 | 149 | 47.4428 | 60.4293 | 104 | 0.9 | 1 | 0 |
| 2005 | 26 | $6: 37$ | 439 | 150 | 47.3315 | 60.1920 | 294 | 0.8 | 2 | 1 |
| 2005 | 26 | $9: 32$ | 439 | 151 | 47.2453 | 60.2030 | 202 | 0.6 | 0 | 0 |
| 2005 | 26 | $12: 08$ | 437 | 152 | 47.1260 | 60.4077 | 164 | 1.0 | 6 | 0 |
| 2005 | 26 | $14: 22$ | 437 | 153 | 47.0927 | 60.4237 | 154 | 0.8 | 3 | 0 |
| 2005 | 26 | $16: 19$ | 437 | 154 | 47.1183 | 60.5427 | 170 | 0.6 | 1 | 0 |
| 2005 | 26 | $18: 27$ | 437 | 155 | 47.1397 | 60.7128 | 167 | 0.7 | 2 | 1 |

Table 2. Results of fixed effects model analyses of length-aggregated data, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the vessel and survey effect and (3) a difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded (=1) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3. September \& July data |  |  | $\chi^{2}$ | $\boldsymbol{P}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | $\chi^{2}$ | P | DF | $\beta_{\mathrm{v}}$ | SE |  |  |
| ATL. COD | 0 | 81 | -0.1390 | 0.0788 | 3.11 | 0.5960 | 12.21 | 0.2850 | 187 | -0.3094 | 0.0760 | 16.55 | 0.1790 |
| ATL. COD | 1 | 79 | 0.1171 | 0.0517 | 5.12 | 0.0640 | 0.40 | 0.6350 | 183 | 0.0994 | 0.0444 | 5.00 | 0.1000 |
| WHITE HAKE | 0 | 20 | 0.2855 | 0.0782 | 13.3 | 0.0100 | 2.11 | 0.3620 | 34 | 0.2276 | 0.0636 | 12.82 | 0.0110 |
| REDFISH <br> (SEBASTES SP.) | 0 | 23 | 0.0046 | 0.0972 | 0.00 | 0.9810 | 0.39 | 0.6400 | 141 | 0.1552 | 0.0846 | 3.37 | 0.9250 |
| REDFISH <br> (SEBASTES SP.) | 1 | 23 | 0.0046 | 0.0972 | 0.00 | 0.9820 | 0.38 | 0.6610 | 132 | 0.0925 | 0.0552 | 2.81 | 0.9200 |
| ATLANTIC HALIBUT | 0 | 10 | 0.5554 | 0.4404 | 1.59 | 0.3360 | 1.11 | 0.4200 | 50 | 0.0718 | 0.1933 | 0.14 | 0.8000 |
| GREENLAND HALIBUT | 0 | 37 | 0.0001 | 0.1045 | 0.00 | 0.9970 | 2.42 | 0.4970 | 70 | 0.1292 | 0.0827 | 2.44 | 0.3160 |
| AMERICAN PLAICE | 0 | 89 | 0.1265 | 0.0402 | 9.92 | 0.0530 | 0.71 | 0.5460 | 231 | 0.1460 | 0.0322 | 20.57 | 0.0040 |
| AMERICAN PLAICE | 1 | 89 | 0.1265 | 0.0402 | 9.92 | 0.0380 | 2.85 | 0.2310 | 230 | 0.1624 | 0.0300 | 29.26 | 0.0030 |
| WITCH FLOUNDER | 0 | 23 | 0.2902 | 0.1756 | 2.73 | 0.3990 | 1.77 | 0.4780 | 121 | 0.1404 | 0.0609 | 5.31 | 0.2960 |
| WITCH FLOUNDER | 1 | 22 | 0.0958 | 0.1451 | 0.44 | 0.7560 | 0.00 | 0.9990 | 120 | 0.0962 | 0.0569 | 2.86 | 0.4820 |
| YELLOWTAIL | 0 | 36 | 0.1828 | 0.0836 | 4.78 | 0.0580 | 0.01 | 0.9550 | 108 | 0.1888 | 0.0493 | 14.70 | 0.0120 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 18 | 0.5245 | 0.1863 | 7.92 | 0.2440 | 1.12 | 0.6210 | 48 | 0.4582 | 0.1105 | 17.18 | 0.0910 |
| WINTER FLOUNDER | 1 | 16 | 0.4582 | 0.1420 | 10.4 | 0.1150 | 1.14 | 0.5210 | 46 | 0.3963 | 0.0877 | 20.43 | 0.0160 |
| STRIPED ATL | 0 | 5 |  |  |  |  | 0.30 | 0.6300 | 49 | 0.1552 | 0.1934 | 0.64 | 0.5330 |
| WOLFFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL. HERRING | 0 | 56 | 1.0122 | 0.2543 | 15.8 | 0.4330 | 13.41 | 0.0280 |  |  |  |  |  |
| ATL. HERRING | 1 | 50 | 0.7294 | 0.0730 | 99.9 | 0.0130 | 18.62 | 0.0120 |  |  |  |  |  |
| GASPEREAU | 0 | 6 |  |  |  |  | 19.36 | 0.0380 |  |  |  |  |  |
| RAINBOW SMELT | 0 | 8 | -0.0096 | 0.2718 | 0.00 | 0.9480 |  |  |  |  |  |  |  |
| CAPELIN | 0 | 59 | 0.1578 | 0.1155 | 1.87 | 0.6280 | 0.84 | 0.5630 | 87 | 0.2839 | 0.1936 | 2.15 | 0.7460 |


|  | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3. September \& July data |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | P | $\chi^{2}$ | P | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | P |
| CAPELIN | 1 | 59 |  |  |  |  | 9.31 | 0.0290 |  |  |  |  |  |
| ATL. MACKEREL | 0 | 17 | -1.2770 | 0.5677 | 5.06 | 0.2700 | 2.42 | 0.3693 | 24 | -0.9050 | 0.4292 | 4.45 | 0.3980 |
| LONGFIN HAKE | 0 | 4 |  |  |  |  |  |  | 28 | -0.0255 | 0.2064 | 0.02 | 0.8990 |
| FOURBEARD | 0 | 17 | 0.0058 | 0.2457 | 0.00 | 0.9630 | 1.38 | 0.4570 | 41 | 0.1610 | 0.1762 | 0.83 | 0.5210 |
| ROCKLING |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GREENLAND COD | 0 | 25 | 0.8973 | 0.3387 | 7.02 | 0.0050 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 21 | 0.2385 | 0.1576 | 2.29 | 0.2450 | 0.65 | 0.5410 | 98 | 0.1340 | 0.0782 | 2.93 | 0.2090 |
| SMOOTH SKATE | 0 | 10 | -0.1103 | 0.2292 | 0.23 | 0.6190 | 0.99 | 0.3560 | 47 | 0.1166 | 0.1617 | 0.52 | 0.4990 |
| WINTER SKATE | 0 | 5 |  |  |  |  |  |  | 16 | 0.1357 | 0.3929 | 0.12 | 0.8450 |
| ATLANTIC HAGFISH | 0 | 11 | -0.0220 | 0.3310 | 0.00 | 0.9950 | 4.30 | 0.3000 | 41 | 0.6882 | 0.2310 | 8.88 | 0.1640 |
| LONGHORN | 0 | 18 | 0.9341 | 0.2301 | 16.4 | 0.0350 | 17.39 | 0.0160 |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SHORTHORN | 0 | 30 | 0.0423 | 0.2159 | 0.04 | 0.9150 | 0.03 | 0.9100 | 37 | 0.0538 | 0.2011 | 0.07 | 0.8690 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SHORTHORN | 1 | 29 | -0.2127 | 0.1973 | 1.16 | 0.3120 | 0.38 | 0.5560 | 36 | -0.1717 | 0.1883 | 0.83 | 0.4010 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 0 | 31 | 0.6103 | 0.2454 | 6.19 | 0.8120 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 30 | -0.0281 | 0.2081 | 0.02 | 0.9180 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE | 0 | 49 | 0.2707 | 0.1308 | 4.28 | 0.1370 | 1.43 | 0.3940 | 100 | 0.1509 | 0.0932 | 2.62 | 0.2220 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC HOOKEAR | 0 | 20 | 0.1759 | 0.4133 | 0.18 | 0.7840 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE | 0 | 38 | -0.2195 | 0.1936 | 1.29 | 0.4280 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 12 | 0.0876 | 0.3318 | 0.07 | 0.8790 | 0.80 | 0.4570 | 71 | -0.2926 | 0.1289 | 5.15 | 0.2540 |
| SEA RAVEN | 1 | 12 |  |  |  |  | 1.28 | 0.2710 | 69 | -0.2655 | 0.1123 | 5.59 | 0.0190 |
| ALLIGATORFISH | 0 | 38 | 0.0030 | 0.1626 | 0.00 | 0.9930 | 2.11 | 0.2970 | 91 | -0.1242 | 0.1114 | 1.24 | 0.2770 |
| ARCTIC | 0 | 26 | -0.7763 | 0.3026 | 6.58 | 0.0570 |  |  |  |  |  |  |  |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 36 | -0.0964 | 0.1506 | 0.41 | 0.5240 | 11.51 | 0.0110 |  |  |  |  |  |
| THREESPINE | 0 | 13 | 0.8892 | 0.3315 | 7.19 | 0.3700 |  |  |  |  |  |  |  |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 7 |  |  |  |  | 0.57 | 0.7060 | 22 | -0.2663 | 0.1903 | 1.96 | 0.3660 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3. September \& July data |  |  | $\chi^{2}$ | $\boldsymbol{P}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | $\chi^{2}$ | $\boldsymbol{P}$ | DF | $\beta_{\mathrm{v}}$ | SE |  |  |
| ATL SPINY | 0 | 23 | 0.0954 | 0.2421 | 0.16 | 0.6810 | 2.00 | 0.3010 | 48 | 0.2650 | 0.1839 | 2.08 | 0.1940 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 38 | -0.0348 | 0.1643 | 0.04 | 0.9330 | 0.84 | 0.7010 | 47 | -0.0144 | 0.1516 | 0.01 | 0.9610 |
| NORTHERN SAND | 0 | 10 | -2.2011 | 0.6718 | 10.73 | 0.1240 | 2.05 | 0.2740 | 53 | 1.0721 | 0.2841 | 14.24 | 0.0120 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NORTHERN SAND | 1 | 10 | $-2.2011$ | 0.6718 | 10.73 | 0.1310 | 107.2 | 0.1170 | 46 | 0.7430 | 0.1769 | 17.64 | 0.0250 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 12 | 0.7445 | 0.5062 | 2.16 | 0.3060 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 43 | 0.2742 | 0.1533 | 3.20 | 0.1370 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 20 | 0.4840 | 0.2517 | 3.70 | 0.0970 | 2.51 | 0.3140 | 43 | 0.7130 | 0.1634 | 19.04 | 0.0060 |
| DAUBED SHANNY | 0 | 65 | -0.3893 | 0.0666 | 34.16 | 0.0010 | 169.5 | 0.0010 |  |  |  |  |  |
| 4-LINE SNAKE | 0 | 24 | 0.2298 | 0.2581 | 0.79 | 0.5180 |  |  |  |  |  |  |  |
| BLENNY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 28 | 0.1971 | 0.1126 | 3.07 | 0.1060 |  |  |  |  |  |  |  |
| OCEAN | 0 | 8 |  |  |  |  | 5.52 | 0.0940 | 36 | -0.4106 | 0.1895 | 4.69 | 0.1150 |
| POUT(COMMON) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| VAHL'S EELPOUT | 0 | 17 | 0.4203 | 0.2281 | 3.40 | 0.3680 | 0.47 | 0.6660 | 45 | 0.5225 | 0.1440 | 13.17 | 0.0240 |
| VAHL'S EELPOUT | 1 | 16 | 0.0718 | 0.1556 | 0.21 | 0.7390 | 10.51 | 0.0750 | 44 | 0.3794 | 0.1300 | 8.52 | 0.0300 |
| WHITE | 0 | 6 |  |  |  |  |  |  | 18 | -0.1363 | 0.2759 | 0.24 | 0.6930 |
| BARRACUDINA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL. HOOKEAR | 0 | 32 | -0.2267 | 0.2362 | 0.92 | 0.5350 | 4.39 | 0.3270 | 55 | -0.5233 | 0.1950 | 7.20 | 0.0890 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 15 | -0.0599 | 0.1566 | 0.15 | 0.6540 | 12.38 | 0.0430 |  |  |  |  |  |
| HYAS | 0 | 67 | -0.0272 | 0.1285 | 0.04 | 0.8790 | 3.37 | 0.2920 |  |  |  |  |  |
| COARCTATUS |  |  |  |  |  |  |  |  |  |  |  |  |  |
| HYAS | 1 | 66 | -0.1314 | 0.1185 | 1.23 | 0.3650 | 74.13 | 0.0010 |  |  |  |  |  |
| COARCTATUS |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NORTHERN STONE | 0 | 10 | 0.9883 | 0.5948 | 2.76 | 0.1160 | 1.85 | 0.2930 | 41 | 0.5854 | 0.2176 | 7.24 | 0.0120 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 84 | -0.0274 | 0.0772 | 0.13 | 0.7910 | 15.89 | 0.0210 |  |  |  |  |  |
| SNOW CRAB | 1 | 82 | -0.1460 | 0.0669 | 4.76 | 0.0710 | 266.6 | 0.0010 |  |  |  |  |  |
| Hyas araneus | 0 | 43 | -0.7651 | 0.2401 | 10.15 | 0.0100 | 64.05 | 0.0200 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 23 | 0.0737 | 0.2125 | 0.12 | 0.7580 | 0.03 | 0.9280 | 128 | 0.1501 | 0.1119 | 1.80 | 0.6060 |
| SHORT-FIN SQUID | 1 | 23 | 0.0737 | 0.2125 | 0.12 | 0.7540 | 0.02 | 0.9820 | 127 | 0.0499 | 0.1034 | 0.23 | 0.8410 |

Table 3. Results of fixed effects model analyses of length-aggregated data (all relevant set pairs) testing for (1) a difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the vessel and survey effect and (3) a difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded (=1) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

| Species | outlier | 1. September survey data only |  |  |  |  | 2. Survey effect |  | 3. September \& July data |  |  | $\chi^{2}$ | P |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | P | $\chi^{2}$ | $\boldsymbol{P}$ | DF | $\beta_{\mathrm{v}}$ | SE |  |  |
| ATL. COD | 0 | 87 | -0.0837 | 0.0790 | 1.12 | 0.693 | 0.00 | 0.951 | 197 | -0.2323 | 0.0762 | 9.30 | 0.342 |
| ATL. COD | 1 | 85 | 0.1636 | 0.0550 | 8.86 | 0.013 | 9.16 | 0.322 | 193 | 0.1617 | 0.0473 | 11.67 | 0.013 |
| WHITE HAKE | 0 | 20 | 0.2855 | 0.0782 | 13.34 | 0.007 | 0.05 | 0.903 | 35 | 0.2745 | 0.0599 | 21.02 | 0.008 |
| REDFISH (SEBASTES SP.) | 0 | 24 | 0.0029 | 0.0957 | 0.00 | 0.993 | 0.44 | 0.676 | 147 | 0.1586 | 0.0826 | 3.69 | 0.922 |
| REDFISH (SEBASTES SP.) | 1 | 24 |  |  |  |  | 0.43 | 0.671 | 138 | 0.0980 | 0.0545 | 3.24 | 0.898 |
| HALIBUT(ATLANTIC) | 0 | 10 | 0.5554 | 0.4404 | 1.59 | 0.330 | 1.11 | 0.439 | 50 | 0.0718 | 0.1933 | 0.14 | 0.790 |
| GREENLAND HALIBUT | 0 | 38 | 0.0013 | 0.1033 | 0.00 | 0.997 | 0.19 | 0.829 | 74 | 0.0449 | 0.0933 | 0.23 | 0.760 |
| AMERICAN PLAICE | 0 | 95 | 0.1439 | 0.0383 | 14.16 | 0.022 | 1.98 | 0.319 | 242 | 0.1565 | 0.0311 | 25.36 | 0.005 |
| AMERICAN PLAICE | 1 | 95 |  |  |  |  | 0.34 | 0.681 | 241 | 0.1720 | 0.0290 | 35.14 | 0.001 |
| WITCH FLOUNDER | 0 | 23 | 0.2902 | 0.1756 | 2.73 | 0.359 | 1.88 | 0.458 | 125 | 0.1367 | 0.0600 | 5.18 | 0.294 |
| WITCH FLOUNDER | 1 | 22 | 0.0958 | 0.1451 | 0.44 | 0.729 | 0.00 | 0.993 | 124 | 0.0930 | 0.0561 | 2.74 | 0.506 |
| YELLOWTAIL | 0 | 40 | 0.1920 | 0.0766 | 6.29 | 0.024 | 0.00 | 0.987 | 114 | 0.1934 | 0.0473 | 16.70 | 0.013 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 19 | 0.5210 | 0.1815 | 8.24 | 0.246 | 2.70 | 0.320 | 51 | 0.4189 | 0.1063 | 15.53 | 0.077 |
| WINTER FLOUNDER | 1 | 17 | 0.4539 | 0.1381 | 10.80 | 0.108 | 2.37 | 0.436 | 49 | 0.3530 | 0.0844 | 17.48 | 0.011 |
| STRIPED ATL WOLFFISH | 0 | 5 |  |  |  |  | 0.27 | 0.681 | 50 | 0.1824 | 0.1938 | 0.89 | 0.453 |
| ATL. HERRING | 0 | 60 | 1.0097 | 0.2457 | 16.89 | 0.443 | 19.15 | 0.015 |  |  |  |  |  |
| ATL. HERRING | 1 | 54 | 0.7273 | 0.0716 | 103.1 | 0.013 | 13.92 | 0.033 |  |  |  |  |  |
| GASPEREAU | 0 | 6 |  |  |  |  | 19.69 | 0.044 |  |  |  |  |  |
| CAPELIN | 0 | 61 | 0.1062 | 0.1212 | 0.77 | 0.778 | 9.74 | 0.020 | 91 | 0.2479 | 0.1893 | 1.72 | 0.775 |
| CAPELIN | 1 | 61 |  |  |  |  | 1.11 | 0.510 | 87 | 0.2548 | 0.1034 | 6.07 | 0.499 |
| ATL. MACKEREL | 0 | 20 | -0.8947 | 0.4669 | 3.67 | 0.416 | 1.86 | 0.464 | 28 | -0.6387 | 0.3733 | 2.93 | 0.531 |
| LONGFIN HAKE | 0 | 4 |  |  |  |  | 1.06 | 0.392 | 29 | -0.0264 | 0.2025 | 0.02 | 0.907 |
| FOURBEARD ROCKLING | 0 | 18 | -0.0071 | 0.2404 | 0.00 | 0.954 | 1.89 | 0.376 | 44 | 0.1752 | 0.1715 | 1.04 | 0.487 |
| GREENLAND COD | 0 | 28 | 0.7283 | 0.2565 | 8.06 | 0.004 | - | - |  |  |  |  |  |
| THORNY SKATE | 0 | 21 | 0.2385 | 0.1576 | 2.29 | 0.239 | 0.63 | 0.563 | 100 | 0.1364 | 0.0775 | 3.10 | 0.192 |

| Species | outlier | DF | 1. September survey data only |  |  |  | 2. Survey effect |  | 3. September \& July data |  |  | $\chi^{2}$ | $\boldsymbol{P}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | P | $\chi^{2}$ | $\boldsymbol{P}$ | DF | $\beta_{\mathrm{v}}$ | SE |  |  |
| SMOOTH SKATE | 0 | 10 | -0.1103 | 0.2292 | 0.23 | 0.629 | 0.82 | 0.357 | 49 | 0.0994 | 0.1623 | 0.38 | 0.541 |
| WINTER SKATE | 0 | 5 |  |  |  |  | 0.32 | 0.644 | 16 | 0.1357 | 0.3929 | 0.12 | 0.845 |
| ATLANTIC HAGFISH | 0 | 11 | -0.0220 | 0.3310 | 0.00 | 0.995 | 4.05 | 0.312 | 42 | 0.6701 | 0.2297 | 8.51 | 0.172 |
| LONGHORN SCULPIN | 0 | 20 | 0.9203 | 0.2256 | 16.64 | 0.040 | 17.80 | 0.014 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 32 | 0.0370 | 0.2122 | 0.03 | 0.910 | 0.38 | 0.554 | 39 | 0.0488 | 0.1983 | 0.06 | 0.889 |
| SHORTHORN SCULPIN | 1 | 31 | -0.2142 | 0.1951 | 1.21 | 0.317 | 0.03 | 0.901 | 38 | -0.1737 | 0.1863 | 0.87 | 0.393 |
| ARCTIC STAGHORN | 0 | 36 | 0.3909 | 0.2391 | 2.67 | 0.899 | - | - | 36 | 0.3909 | 0.2391 | 2.67 | 0.905 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 35 | -0.1838 | 0.1981 | 0.86 | 0.554 | - | - | 35 | -0.1838 | 0.1981 | 0.86 | 0.534 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 51 | 0.2171 | 0.1253 | 3.00 | 0.164 | 0.87 | 0.499 | 103 | 0.1299 | 0.0905 | 2.06 | 0.263 |
| ARCTIC HOOKEAR | 0 | 22 | 0.0820 | 0.3836 | 0.05 | 0.892 | - | - | 23 | 0.1021 | 0.3755 | 0.07 | 0.879 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 41 | -0.3075 | 0.1945 | 2.50 | 0.249 | - | - | 44 | -0.2604 | 0.1909 | 1.86 | 0.310 |
| SEA RAVEN | 0 | 13 | 0.1295 | 0.3235 | 0.16 | 0.705 | 1.44 | 0.223 | 74 | -0.2872 | 0.1252 | 5.26 | 0.241 |
| SEA RAVEN | 1 | 13 | 0.1295 | 0.3235 | 0.16 | 0.718 | 1.79 | 0.442 | 72 | -0.2598 | 0.1091 | 5.68 | 0.017 |
| ALLIGATORFISH | 0 | 40 | -0.1174 | 0.1561 | 0.57 | 0.546 | 0.52 | 0.620 | 94 | -0.1742 | 0.1084 | 2.58 | 0.170 |
| ARCTIC ALLIGATORFISI | 0 | 28 | -0.6934 | 0.2916 | 5.65 | 0.082 | - | - | 28 | -0.6934 | 0.2916 | 5.65 | 0.078 |
| ATL SEA POACHER | 0 | 37 | -0.0903 | 0.1492 | 0.37 | 0.522 | 11.65 | 0.013 |  |  |  |  |  |
| THREESPINE | 0 | 13 | 0.8892 | 0.3315 | 7.19 | 0.363 | - | - | 13 | 0.8892 | 0.3315 | 7.19 | 0.363 |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 7 |  |  |  |  | 0.73 | 0.632 | 23 | -0.2480 | 0.1779 | 1.94 | 0.304 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 25 | 0.1057 | 0.2264 | 0.22 | 0.660 | 1.59 | 0.335 | 51 | 0.2493 | 0.1772 | 1.98 | 0.222 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 41 | -0.0598 | 0.1551 | 0.15 | 0.878 | 0.91 | 0.673 | 50 | -0.0396 | 0.1442 | 0.08 | 0.932 |
| NORTHERN SAND LANC | 0 | 11 | -2.2125 | 0.6442 | 11.80 | 0.087 | 132.4 | 0.286 | 54 | 1.0711 | 0.2815 | 14.48 | 0.012 |
| NORTHERN SAND LANC | 1 | 11 |  |  |  |  | 5.84 | 0.090 | 47 | 0.7419 | 0.1752 | 17.94 | 0.036 |
| FISH DOCTOR | 0 | 15 | 0.7099 | 0.4454 | 2.54 | 0.285 | - | - |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 48 | 0.3000 | 0.1387 | 4.68 | 0.077 | - | - |  |  |  |  |  |
| SNAKEBLENNY | 0 | 20 | 0.4840 | 0.2517 | 3.70 | 0.094 | 2.47 | 0.320 | 45 | 0.7121 | 0.1626 | 19.18 | 0.002 |
| DAUBED SHANNY | 0 | 69 | -0.4214 | 0.0700 | 36.26 | 0.001 | 195.3 | 0.001 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 27 | 0.3846 | 0.2027 | 3.60 | 0.257 | - | - |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 30 | 0.2024 | 0.1099 | 3.39 | 0.102 | - |  |  |  |  |  |  |
| OCEAN POUT(COMMON | 0 | 8 |  |  |  |  | 5.47 | 0.089 | 37 | -0.4094 | 0.1826 | 5.02 | 0.107 |

| Species | outlier | 1. September survey data only |  |  |  |  | 2. Survey effect |  | 3. September \& July data |  |  | $\chi^{2}$ | P |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | DF | $\beta_{\mathrm{v}}$ | SE | $\chi^{2}$ | P | $\chi^{2}$ | $\boldsymbol{P}$ | DF | $\beta_{\mathrm{v}}$ | SE |  |  |
| VAHL'S EELPOUT | 0 | 17 | 0.4203 | 0.2281 | 3.40 | 0.325 | 4.65 | 0.076 | 48 | 0.5284 | 0.1415 | 13.94 | 0.023 |
| VAHL'S EELPOUT | 1 | 16 | 0.0718 | 0.1556 | 0.21 | 0.718 | 1.66 | 0.632 | 47 | 0.3892 | 0.1285 | 9.17 | 0.016 |
| WHITE BARRACUDINA | 0 | 6 |  |  |  |  | 0.02 | 0.959 | 18 | -0.1363 | 0.2759 | 0.24 | 0.698 |
| ATL. HOOKEAR SCULPII | 0 | 36 | -0.3321 | 0.2272 | 2.14 | 0.314 | 20.20 | 0.203 | 60 | -0.5927 | 0.1888 | 9.86 | 0.048 |
| ATL ROCK CRAB | 0 | 16 | -0.0673 | 0.1538 | 0.19 | 0.571 | 11.49 | 0.038 |  |  |  |  |  |
| HYAS COARCTATUS | 0 | 73 | -0.0030 | 0.1205 | 0.00 | 0.973 | 19.18 | 0.002 |  |  |  |  |  |
| HYAS COARCTATUS | 1 | 72 | -0.0843 | 0.1129 | 0.56 | 0.553 | 17.16 | 0.285 | 130 | 0.2086 | 0.0858 | 5.91 | 0.119 |
| NORTHERN STONE CRA] | 0 | 10 | 0.9883 | 0.5948 | 2.76 | 0.129 | 2.08 | 0.250 | 42 | 0.5571 | 0.2094 | 7.08 | 0.017 |
| SNOW CRAB | 0 | 91 | -0.0867 | 0.0770 | 1.27 | 0.373 | 29.19 | 0.001 |  |  |  |  |  |
| SNOW CRAB | 1 | 89 | -0.2022 | 0.0674 | 9.01 | 0.018 | 201.12 | 0.012 |  |  |  |  |  |
| TOAD CRAB | 0 | 48 | -0.6583 | 0.2144 | 9.43 | 0.012 | 62.41 | 0.021 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 24 | 0.0674 | 0.2067 | 0.11 | 0.741 | 0.27 | 0.926 | 131 | 0.1573 | 0.1102 | 2.04 | 0.564 |
| SHORT-FIN SQUID | 1 | 24 |  |  |  |  | 0.00 | 0.993 | 130 | 0.0584 | 0.1019 | 0.33 | 0.786 |

Table 4. Results of mixed effects model analyses of length-aggregated data, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the vessel and survey effect and (3) a difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value $=0$ ) or excluded ( $=1$ ) from the analysis. $P$ is the probability value for the $t$-statistic.

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\mathrm{v}}$ | SE | t | P | $t$ | P | DF | $\beta_{\mathrm{v}}$ | SE | , | P |
| ATL. COD | 0 | 80 | 0.0944 | 0.0827 | 1.142 | 0.2570 | 0.729 | 0.4670 | 185 | 0.0448 | 0.0798 | 0.562 | 0.5746 |
| ATL. COD | 1 | 79 | 0.1321 | 0.0748 | 1.766 | 0.0812 | 0.361 | 0.7188 | 182 | 0.1074 | 0.0687 | 1.563 | 0.1198 |
| WHITE HAKE | 0 | 20 | 0.2792 | 0.0972 | 2.873 | 0.0094 | 0.135 | 0.8931 | 34 | 0.2708 | 0.0833 | 3.251 | 0.0026 |
| REDFISH (SEBASTES SP.) | 0 | 23 | -0.1128 | 0.1492 | -0.756 | 0.4574 | -0.894 | 0.3728 | 141 | 0.0994 | 0.1185 | 0.839 | 0.4031 |
| REDFISH (SEBASTES SP.) | 1 | 23 |  |  |  |  | -0.882 | 0.3793 | 132 | 0.0429 | 0.0968 | 0.444 | 0.6581 |
| ATLANTIC HALIBUT | 0 | 10 | 0.5554 | 0.3986 | 1.393 | 0.1938 | 0.628 | 0.5332 | 49 | 0.0901 | 0.2206 | 0.409 | 0.6847 |
| GREENLAND HALIBUT | 0 | 37 | 0.1999 | 0.1767 | 1.131 | 0.2652 | 0.826 | 0.4118 | 69 | 0.0961 | 0.1236 | 0.777 | 0.4398 |
| AMERICAN PLAICE | 0 | 89 | 0.1642 | 0.0478 | 3.436 | 0.0009 | 1.475 | 0.1417 | 228 | 0.1087 | 0.0444 | 2.450 | 0.0150 |
| AMERICAN PLAICE | 1 | 89 |  |  |  |  | 1.126 | 0.2615 | 227 | 0.1254 | 0.0421 | 2.982 | 0.0032 |
| WITCH FLOUNDER | 0 | 23 | 0.2891 | 0.2198 | 1.316 | 0.2013 | 1.618 | 0.1082 | 120 | 0.0281 | 0.0904 | 0.311 | 0.7566 |
| WITCH FLOUNDER | 1 | 22 | 0.1498 | 0.1796 | 0.834 | 0.4133 | 0.945 | 0.3466 | 119 | 0.0032 | 0.0834 | 0.039 | 0.9693 |
| YELLOWTAIL | 0 | 36 | 0.0887 | 0.1519 | 0.584 | 0.5630 | 0.202 | 0.8403 | 107 | 0.0693 | 0.0801 | 0.865 | 0.3888 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 18 | 0.2383 | 0.2536 | 0.940 | 0.3597 | 0.086 | 0.9321 | 48 | 0.2391 | 0.1508 | 1.586 | 0.1193 |
| WINTER FLOUNDER | 1 | 16 | 0.2251 | 0.2552 | 0.882 | 0.3909 | 0.034 | 0.9734 | 46 | 0.2339 | 0.1479 | 1.581 | 0.1207 |
| STRIPED ATL WOLFFISH | 0 | 5 | 0.7576 | 0.9949 | 0.761 | 0.4807 | 0.734 | 0.4667 | 49 | 0.0688 | 0.2390 | 0.288 | 0.7745 |
| ATL. HERRING | 0 | 55 | 0.5431 | 0.2722 | 1.995 | 0.0510 | 1.445 | 0.1506 | 155 | 0.2686 | 0.1438 | 1.867 | 0.0637 |
| ATL. HERRING | 1 | 49 | 0.4179 | 0.1658 | 2.520 | 0.0150 | 1.001 | 0.3186 | 149 | 0.1900 | 0.1225 | 1.551 | 0.1230 |
| GASPEREAU | 0 | 6 | -1.5132 | 0.9643 | -1.569 | 0.1676 | -1.998 | 0.0537 | 35 | -0.6099 | 0.2760 | -2.210 | 0.0337 |
| RAINBOW SMELT | 0 | 8 |  |  |  |  |  |  |  |  |  |  |  |
| CAPELIN | 0 | 58 | 0.1140 | 0.1605 | 0.710 | 0.4804 | -1.204 | 0.2318 | 86 | 0.2313 | 0.1574 | 1.469 | 0.1454 |
| CAPELIN | 1 | 58 |  |  |  |  | -2.094 | 0.0394 |  |  |  |  |  |
| ATL. MACKEREL | 0 | 17 | -0.8828 | 0.8795 | -1.004 | 0.3295 | -0.306 | 0.7622 | 23 | -0.7248 | 0.6711 | -1.080 | 0.2913 |
| ARCTIC COD | 0 | 6 |  |  |  |  |  |  |  |  |  |  |  |
| LONGFIN HAKE | 0 | 4 |  |  |  |  | -1.096 | 0.2826 | 28 | 0.4212 | 0.3610 | 1.167 | 0.2531 |
| FOURBEARD ROCKLING | 0 | 17 | 0.1573 | 0.3229 | 0.487 | 0.6325 | -0.578 | 0.5664 | 41 | 0.2715 | 0.2146 | 1.265 | 0.2129 |

| Species | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P | $t$ | $\boldsymbol{P}$ | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P |
| GREENLAND COD | 0 | 25 | 0.8985 | 0.3327 | 2.700 | 0.0122 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 21 | 0.1726 | 0.1776 | 0.972 | 0.3422 | 0.239 | 0.8115 | 98 | 0.1345 | 0.0957 | 1.406 | 0.1629 |
| SMOOTH SKATE | 0 | 10 | -0.1103 | 0.2430 | -0.454 | 0.6596 | -0.709 | 0.4822 | 47 | 0.0710 | 0.1793 | 0.396 | 0.6939 |
| WINTER SKATE | 0 | 5 |  |  |  |  | -0.187 | 0.8545 |  | -0.1013 | 0.4973 | -0.204 | 0.8412 |
|  |  |  |  |  |  |  |  |  | 16 |  |  |  |  |
| ATLANTIC HAGFISH | 0 | 11 | 0.0145 | 0.3794 | 0.038 | 0.9703 | -1.031 | 0.3085 | 41 | 0.4167 | 0.2674 | 1.558 | 0.1268 |
| LONGHORN SCULPIN | 0 | 18 | 0.8736 | 0.3081 | 2.836 | 0.0110 | 3.624 | 0.0005 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 30 | -0.0042 | 0.2887 | -0.015 | 0.9884 | -0.303 | 0.7638 | 37 | 0.0292 | 0.2613 | 0.112 | 0.9117 |
| SHORTHORN SCULPIN | 1 | 29 | -0.2186 | 0.2047 | -1.068 | 0.2944 | -0.646 | 0.5226 | 36 | -0.1734 | 0.1918 | -0.904 | 0.3719 |
| ARCTIC STAGHORN | 0 | 31 | 0.1116 | 0.2766 | 0.403 | 0.6894 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 30 | -0.0009 | 0.2593 | -0.003 | 0.9973 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 49 | 0.4433 | 0.1736 | 2.553 | 0.0138 | 1.097 | 0.2754 | 100 | 0.3159 | 0.1318 | 2.397 | 0.0184 |
| ARCTIC HOOKEAR | 0 | 20 | -0.4776 | 0.6503 | -0.734 | 0.4712 |  |  | 21 | -0.3651 | 0.6272 | -0.582 | 0.5667 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 38 | -0.1032 | 0.2559 | -0.403 | 0.6890 |  |  | 41 | 0.0061 | 0.2525 | 0.024 | 0.9808 |
| SEA RAVEN | 0 | 12 | 0.0625 | 0.3647 | 0.171 | 0.8668 | 0.728 | 0.4693 | 71 | -0.2818 | 0.1577 | -1.787 | 0.0782 |
| SEA RAVEN | 1 | 12 |  |  |  |  | 1.095 | 0.2772 | 69 | -0.2718 | 0.1126 | -2.414 | 0.0184 |
| ALLIGATORFISH | 0 | 38 | -0.0292 | 0.2052 | -0.142 | 0.8878 | 0.361 | 0.7191 | 91 | -0.0832 | 0.1405 | -0.592 | 0.5551 |
| ARCTIC | 0 | 26 | -0.5091 | 0.5032 | -1.012 | 0.3210 |  |  |  |  |  |  |  |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 36 | -0.1093 | 0.1721 | -0.635 | 0.5294 | -2.439 | 0.0178 |  |  |  |  |  |
| THREESPINE | 0 | 13 | 0.3315 | 0.6158 | 0.538 | 0.5994 |  |  |  |  |  |  |  |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 7 |  |  |  |  | -0.577 | 0.5700 | 22 | -0.3199 | 0.2307 | -1.387 | 0.1794 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 23 | 0.1898 | 0.2804 | 0.677 | 0.5051 | -0.986 | 0.3294 | 48 | 0.3701 | 0.2142 | 1.728 | 0.0904 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 38 | 0.1277 | 0.2455 | 0.520 | 0.6060 | -0.613 | 0.5428 | 47 | 0.1816 | 0.2237 | 0.812 | 0.4210 |
| NORTHERN SAND | 0 | 10 | -0.8689 | 0.6811 | -1.276 | 0.2309 | -1.502 | 0.1392 | 52 | 0.6013 | 0.4169 | 1.442 | 0.1552 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NORTHERN SAND | 1 | 10 |  |  |  |  | -1.709 | 0.0943 | 46 | 0.3517 | 0.3348 | 1.051 | 0.2990 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 12 | 0.5059 | 0.6919 | 0.731 | 0.4787 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 43 | 0.3083 | 0.1865 | 1.653 | 0.1055 |  |  |  |  |  |  |  |

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P | $t$ | P | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P |
| SNAKEBLENNY | 0 | 20 | 0.7239 | 0.3368 | 2.150 | 0.0440 | 0.120 | 0.9047 | 43 | 0.6715 | 0.2074 | 3.238 | 0.0023 |
| DAUBED SHANNY | 0 | 65 | -0.2845 | 0.1095 | -2.599 | 0.0116 | -3.416 | 0.0009 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 24 | -0.0624 | 0.3978 | -0.157 | 0.8766 |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 28 | 0.1825 | 0.2788 | 0.655 | 0.5181 |  |  |  |  |  |  |  |
| OCEAN POUT(COMMON) | 0 | 8 |  |  |  |  | 1.801 | 0.0803 | 36 | -0.1814 | 0.2355 | -0.770 | 0.4462 |
| VAHL'S EELPOUT | 0 | 17 | 0.1913 | 0.3218 | 0.594 | 0.5601 | -0.514 | 0.6101 | 45 | 0.3211 | 0.1961 | 1.637 | 0.1085 |
| VAHL'S EELPOUT | 1 | 16 | -0.0089 | 0.1817 | -0.049 | 0.9615 | -1.403 | 0.1679 | 44 | 0.2539 | 0.1740 | 1.460 | 0.1515 |
| WHITE BARRACUDINA | 0 | 6 |  |  |  |  | 0.553 | 0.5876 | 18 | -0.3547 | 0.5724 | -0.620 | 0.5432 |
| ATL. HOOKEAR | 0 | 32 | -0.6678 | 0.3449 | -1.936 | 0.0617 | 1.388 | 0.1708 | 55 | -0.9789 | 0.2979 | -3.286 | 0.0018 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL. HOOKEAR | 1 | 30 | -0.8841 | 0.3346 | -2.642 | 0.0130 | 1.049 | 0.2991 | 53 | -1.1286 | 0.2938 | -3.841 | 0.0003 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 15 | -0.0599 | 0.1384 | -0.432 | 0.6716 | 1.993 | 0.0541 | 36 | -0.6354 | 0.2466 | $-2.577$ | 0.0142 |
| HYAS COARCTATUS | 0 | 67 | -0.1829 | 0.2001 | -0.914 | 0.3642 | -2.317 | 0.0222 |  |  |  |  |  |
| HYAS COARCTATUS | 1 | 66 | -0.2608 | 0.1801 | -1.448 | 0.1524 | -3.837 | 0.0002 |  |  |  |  |  |
| NORTHERN STONE | 0 | 10 | 1.4003 | 0.8290 | 1.689 | 0.1221 | 1.177 | 0.2462 | 40 | 0.6218 | 0.2370 | 2.624 | 0.0122 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 84 | -0.1033 | 0.0840 | -1.231 | 0.2219 | -5.079 | 0.0000 |  |  |  |  |  |
| SNOW CRAB | 1 | 82 | -0.1525 | 0.0792 | -1.926 | 0.0576 | -5.517 | 0.0000 |  |  |  |  |  |
| TOAD CRAB | 0 | 43 | -1.3664 | 0.3489 | -3.917 | 0.0003 | -2.182 | 0.0319 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 23 | 0.2175 | 0.2588 | 0.840 | 0.4093 | 0.917 | 0.3607 | 128 | 0.0152 | 0.1564 | 0.097 | 0.9225 |
| SHORT-FIN SQUID | 1 | 23 |  |  |  |  | 0.986 | 0.3261 | 127 | -0.0092 | 0.1563 | -0.059 | 0.9530 |

Table 5. Results of mixed effects model analyses of length-aggregated data (all relevant set pairs) testing for (1) a difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the vessel and survey effect and (3) a difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded (=1) from the analysis. $P$ is the probability value for the $t$-statistic.

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | $\boldsymbol{P}$ | $t$ | P | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | $\boldsymbol{P}$ |
| ATL. COD | 0 | 86 | 0.1544 | 0.0867 | 1.78 | 0.079 | 0.69 | 0.491 | 195 | 0.1035 | 0.0822 | 1.26 | 0.210 |
| ATL. COD | 1 | 85 | 0.1878 | 0.0815 | 2.31 | 0.024 | 0.31 | 0.757 | 192 | 0.1640 | 0.0743 | 2.21 | 0.029 |
| WHITE HAKE | 0 | 20 | 0.2792 | 0.0972 | 2.87 | 0.009 | -0.19 | 0.851 | 35 | 0.2891 | 0.0790 | 3.66 | 0.001 |
| REDFISH (SEBASTES SP.) | 0 | 24 | -0.1283 | 0.1491 | -0.86 | 0.398 | -1.04 | 0.299 | 146 | 0.0981 | 0.1168 | 0.84 | 0.402 |
| REDFISH (SEBASTES SP.) | 1 | 24 |  |  |  |  | -1.04 | 0.298 | 137 | 0.0487 | 0.0963 | 0.51 | 0.614 |
| ATLANTIC HALIBUT | 0 | 10 | 0.5554 | 0.3986 | 1.39 | 0.194 | 0.63 | 0.533 | 49 | 0.0901 | 0.2206 | 0.41 | 0.685 |
| GREENLAND HALIBUT | 0 | 38 | 0.2095 | 0.1761 | 1.19 | 0.242 | 1.40 | 0.167 | 72 | 0.0295 | 0.1358 | 0.22 | 0.828 |
| AMERICAN PLAICE | 0 | 95 | 0.1681 | 0.0453 | 3.71 | <0.001 | 1.43 | 0.153 | 238 | 0.1166 | 0.0427 | 2.73 | 0.007 |
| AMERICAN PLAICE | 1 | 95 |  |  |  |  | 1.08 | 0.281 | 237 | 0.1326 | 0.0404 | 3.28 | 0.001 |
| WITCH FLOUNDER | 0 | 23 | 0.2891 | 0.2198 | 1.32 | 0.201 | 1.65 | 0.101 | 123 | 0.0244 | 0.0883 | 0.28 | 0.783 |
| WITCH FLOUNDER | 1 | 22 | 0.1498 | 0.1796 | 0.83 | 0.413 | 0.97 | 0.333 | 122 | -0.0002 | 0.0815 | 0.00 | 0.998 |
| YELLOWTAIL | 0 | 40 | 0.1008 | 0.1339 | 0.75 | 0.456 | 0.19 | 0.847 | 113 | 0.0808 | 0.0773 | 1.05 | 0.298 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 19 | 0.2126 | 0.2424 | 0.88 | 0.391 | 0.19 | 0.851 | 51 | 0.1993 | 0.1420 | 1.40 | 0.166 |
| WINTER FLOUNDER | 1 | 17 | 0.1973 | 0.2424 | 0.81 | 0.427 | 0.13 | 0.899 | 49 | 0.1923 | 0.1387 | 1.39 | 0.172 |
| STRIPED ATL WOLFFISH | 0 | 5 |  |  |  |  | 0.68 | 0.498 | 50 | 0.1130 | 0.2391 | 0.47 | 0.639 |
| ATL. HERRING | 0 | 59 | 0.5740 | 0.2640 | 2.17 | 0.034 | 1.58 | 0.116 | 162 | 0.2708 | 0.1418 | 1.91 | 0.058 |
| ATL. HERRING | 1 | 53 | 0.4248 | 0.1645 | 2.58 | 0.013 | 1.16 | 0.250 | 156 | 0.1929 | 0.1216 | 1.59 | 0.115 |
| GASPEREAU | 0 | 6 | -1.5132 | 0.9643 | -1.57 | 0.168 | -1.96 | 0.058 | 36 | -0.6345 | 0.2739 | -2.32 | 0.026 |
| RAINBOW SMELT | 0 | 8 | 0.4500 | 0.4178 | 1.08 | 0.313 |  |  |  |  |  |  |  |
| CAPELIN | 0 | 60 | 0.0155 | 0.1820 | 0.09 | 0.932 | -1.46 | 0.147 | 89 | 0.1693 | 0.1650 | 1.03 | 0.308 |
| CAPELIN | 1 | 60 |  |  |  |  | -2.22 | 0.029 | 90 |  |  |  |  |
| ATL. MACKEREL | 0 | 20 | -0.5150 | 0.8028 | -0.64 | 0.528 | -0.30 | 0.770 | 27 | -0.3837 | 0.6136 | -0.63 | 0.537 |
| LONGFIN HAKE | 0 | 4 |  |  |  |  | -1.08 | 0.288 | 29 | 0.3957 | 0.3492 | 1.13 | 0.266 |
| FOURBEARD ROCKLING | 0 | 18 | 0.1180 | 0.3168 | 0.37 | $0.714$ | -0.80 | 0.427 | 44 | 0.2830 | 0.2090 | 1.35 | 0.183 |
| GREENLAND COD | 0 | 28 | 0.7283 | 0.2624 | 2.78 | 0.010 |  |  |  |  |  |  |  |

| Species | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | $\boldsymbol{P}$ | $t$ | P | DF | $\beta_{\mathrm{v}}$ | SE |  | $\boldsymbol{P}$ |
| THORNY SKATE | 0 | 21 | 0.1726 | 0.1776 | 0.97 | 0.342 | 0.21 | 0.832 | 100 | 0.1391 | 0.0950 | 1.46 | 0.146 |
| SMOOTH SKATE | 0 | 10 | -0.1103 | 0.2430 | -0.45 | 0.660 | -0.58 | 0.566 | 49 | 0.0441 | 0.1813 | 0.24 | 0.809 |
| WINTER SKATE | 0 | 5 |  |  |  |  | -0.19 | 0.855 | 16 | -0.1013 | 0.4973 | -0.20 | 0.841 |
| ATLANTIC HAGFISH | 0 | 11 | 0.0145 | 0.3794 | 0.04 | 0.970 | -0.93 | 0.359 | 42 | 0.3800 | 0.2659 | 1.43 | 0.160 |
| LONGHORN SCULPIN | 0 | 20 | 0.8395 | 0.3025 | 2.78 | 0.012 | 3.68 | <0.001 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 32 | -0.0138 | 0.2830 | -0.05 | 0.962 | -0.32 | 0.753 | 39 | 0.0202 | 0.2570 | 0.08 | 0.938 |
| SHORTHORN SCULPIN | 1 | 31 | -0.2200 | 0.2024 | -1.09 | 0.285 | -0.65 | 0.520 | 38 | -0.1755 | 0.1899 | -0.92 | 0.361 |
| ARCTIC STAGHORN | 0 | 36 | 0.0017 | 0.2719 | 0.01 | 0.995 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 35 | -0.0994 | 0.2635 | -0.38 | 0.708 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 51 | 0.3863 | 0.1658 | 2.33 | 0.024 | 0.91 | 0.367 | 103 | 0.2900 | 0.1273 | 2.28 | 0.025 |
| ARCTIC HOOKEAR | 0 | 22 | -0.4379 | 0.5880 | -0.74 | 0.464 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 41 | -0.2700 | 0.2644 | -1.02 | 0.313 |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 13 | 0.1143 | 0.3558 | 0.32 | 0.753 | 0.89 | 0.377 | 74 | -0.2682 | 0.1525 | -1.76 | 0.083 |
| SEA RAVEN | 1 | 13 | 0.1143 | 0.3558 | 0.32 | 0.753 | 1.24 | 0.218 | 72 | -0.2631 | 0.1077 | -2.44 | 0.017 |
| ALLIGATORFISH | 0 | 40 | -0.1114 | 0.1971 | -0.57 | 0.575 | -0.02 | 0.983 | 94 | -0.1099 | 0.1370 | -0.80 | 0.424 |
| ARCTIC | 0 | 28 | -0.4239 | 0.5006 | -0.85 | 0.404 |  |  |  |  |  |  |  |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 37 | -0.0979 | 0.1711 | -0.57 | 0.571 | -2.51 | 0.015 |  |  |  |  |  |
| THREESPINE | 0 | 13 | 0.3315 | 0.6158 | 0.54 | 0.599 |  |  |  |  |  |  |  |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 7 |  |  |  |  | -0.67 | 0.509 | 23 | -0.2883 | 0.2072 | -1.39 | 0.177 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 25 | 0.1947 | 0.2604 | 0.75 | 0.462 | -0.83 | 0.410 | 51 | 0.3389 | 0.2039 | 1.66 | 0.103 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 41 | 0.0726 | 0.2229 | 0.33 | 0.746 | -0.70 | 0.489 | 50 | 0.1235 | 0.2058 | 0.60 | 0.551 |
| NORTHERN SAND | 0 | 11 | -0.9839 | 0.6481 | -1.52 | 0.157 | -1.69 | 0.097 | 53 | 0.5561 | 0.4132 | 1.35 | 0.184 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NORTHERN SAND | 1 | 11 |  |  |  |  | -1.90 | 0.064 | 47 | 0.3140 | 0.3323 | 0.95 | 0.349 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 15 | 0.4586 | 0.5894 | 0.78 | 0.449 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 48 | 0.3520 | 0.1738 | 2.03 | 0.048 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 20 | 0.7239 | 0.3368 | 2.15 | 0.044 | 0.10 | 0.919 | 44 | 0.6692 | 0.2022 | 3.31 | 0.002 |

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P | $t$ | P | DF | $\beta_{\mathrm{v}}$ | SE | $t$ | P |
| DAUBED SHANNY | 0 | 69 | -0.3154 | 0.1173 | -2.69 | 0.009 | -3.60 | 0.001 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 27 | 0.0478 | 0.3456 | 0.14 | 0.891 |  |  | 27 | 0.0478 | 0.3456 | 0.14 | 0.891 |
| STOUT EELBLENNY | 0 | 30 | 0.2621 | 0.2770 | 0.95 | 0.352 |  |  |  |  |  |  |  |
| OCEAN POUT(COMMON) | 0 | 8 |  |  |  |  | 1.86 | 0.071 | 37 | -0.2074 | 0.2218 | -0.94 | 0.356 |
| VAHL'S EELPOUT | 0 | 17 | 0.1913 | 0.3218 | 0.59 | 0.560 | -0.61 | 0.543 | 47 | 0.3466 | 0.1887 | 1.84 | 0.073 |
| VAHL'S EELPOUT | 1 | 16 | -0.0089 | 0.1817 | -0.05 | 0.962 | -1.53 | 0.133 | 46 | 0.2811 | 0.1679 | 1.67 | 0.101 |
| WHITE BARRACUDINA | 0 | 6 |  |  |  |  | 0.55 | 0.588 | 18 | -0.3547 | 0.5724 | -0.62 | 0.543 |
| ATL. HOOKEAR | 0 | 36 | -0.8419 | 0.3269 | $-2.58$ | 0.014 | 1.22 | 0.229 | 60 | -1.0963 | 0.2841 | -3.86 | <0.001 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL. HOOKEAR | 1 | 34 | -1.0437 | 0.3125 | -3.34 | 0.002 | 0.88 | 0.383 | 58 | -1.2382 | 0.2781 | -4.45 | <0.001 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 16 | -0.0732 | 0.1430 | -0.51 | 0.616 | 1.83 | 0.075 | 38 | -0.6175 | 0.2373 | $-2.60$ | 0.013 |
| Hyas coarctatus | 0 | 73 | -0.2419 | 0.1942 | -1.25 | 0.217 | -2.50 | 0.014 |  |  |  |  |  |
| Hyas coarctatus | 1 | 72 | -0.3093 | 0.1776 | -1.74 | 0.086 | -3.95 | <0.001 |  |  |  |  |  |
| NORTHERN STONE | 0 | 10 | 1.4003 | 0.8290 | 1.69 | 0.122 | 1.25 | 0.220 | 41 | 0.5913 | 0.2258 | 2.62 | 0.012 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 91 | -0.1521 | 0.0857 | -1.78 | 0.079 | -5.24 | <0.001 |  |  |  |  |  |
| SNOW CRAB | 1 | 89 | -0.1976 | 0.0815 | -2.42 | 0.017 | -5.66 | <0.001 |  |  |  |  |  |
| Hyas araneus | 0 | 48 | -1.1424 | 0.3017 | -3.79 | <0.001 | -2.01 | 0.047 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 24 | 0.1957 | 0.2495 | 0.78 | 0.440 | 0.81 | 0.420 | 131 | 0.0349 | 0.1533 | 0.23 | 0.820 |
| SHORT-FIN SQUID | 1 | 24 |  |  |  |  | 0.88 | 0.382 | 130 | 0.0111 | 0.1532 | 0.07 | 0.943 |

Table 6. Results of fixed effects model analyses, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a length-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a length-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value $=0$ ) or excluded ( $=1$ ) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

|  |  | length effect FIXED - September |  |  |  |  | Survey effect |  |  | length effect FIXED - September \& July |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P | Deviance | P | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P |
| ATL. COD | 0 | 1224 | 0.0051 | 0.0027 | 3.64 | 0.3160 | 4001.4 | 0.0010 |  |  |  |  |  |
| ATL. COD | 1 | 1154 | 0.0022 | 0.0026 | 0.73 | 0.3890 | 3493.5 | 0.0010 |  |  |  |  |  |
| WHITE HAKE | 0 | 306 | -0.0045 | 0.0079 | 0.33 | 0.6120 | 581.3 | 0.1690 | 424 | -0.0063 | 0.0062 | 1.03 | 0.5000 |
| REDFISH (SEBASTES SP.) | 0 | 213 | -0.0043 | 0.0076 | 0.32 | 0.7970 | 7007.5 | 0.2970 | 1443 | 0.0035 | 0.0034 | 1.06 | 0.3480 |
| REDFISH (SEBASTES SP.) | 1 | 213 |  |  |  |  | 5836.4 | 0.5630 | 1394 | -0.0009 | 0.0031 | 0.08 | 0.8270 |
| HALIBUT(ATLANTIC) | 0 | 24 |  |  |  |  |  |  | 145 | -0.0025 | 0.0064 | 0.15 | 0.3100 |
| GREENLAND HALIBUT | 0 | 280 | -0.0094 | 0.0057 | 2.74 | 0.3540 | 1160.0 | 0.0090 |  |  |  |  |  |
| GREENLAND HALIBUT | 1 | 280 |  |  |  |  | 1160.0 | 0.0140 |  |  |  |  |  |
| AMERICAN PLAICE | 0 | 1977 | -0.0058 | 0.0030 | 3.73 | 0.1240 | 6596.0 | 0.2520 | 3832 | -0.0119 | 0.0022 | 29.37 | 0.0010 |
| AMERICAN PLAICE | 1 | 1977 |  |  |  |  | 6310.8 | 0.0030 |  |  |  |  |  |
| WITCH FLOUNDER | 0 | 253 | -0.0037 | 0.0150 | 0.06 | 0.7970 | 1636.0 | 0.0460 |  |  |  |  |  |
| WITCH FLOUNDER | 1 | 215 | -0.0096 | 0.0144 | 0.44 | 0.5030 | 1474.3 | 0.0050 |  |  |  |  |  |
| YELLOWTAIL | 0 | 347 | -0.0261 | 0.0100 | 6.75 | 0.2780 | 1913.1 | 0.3780 | 1067 | -0.0089 | 0.0049 | 3.21 | 0.6800 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| YELLOWTAIL | 1 | 302 | -0.0054 | 0.0109 | 0.24 | 0.7910 | 1637.9 | 0.3330 | 980 | -0.0038 | 0.0051 | 0.54 | 0.8670 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 254 | -0.0101 | 0.0107 | 0.89 | 0.6910 | 1038.9 | 0.2100 | 487 | -0.0182 | 0.0073 | 6.18 | 0.4440 |
| WINTER FLOUNDER | 1 | 182 | 0.0393 | 0.0097 | 16.49 | 0.0100 | 675.2 | 0.0160 |  |  |  |  |  |
| STRIPED ATL WOLFFISH | 0 | 4 |  |  |  |  |  |  | 138 | 0.0083 | 0.0079 | 1.12 | 0.3560 |
| ATL. HERRING | 0 | 592 | 0.0442 | 0.0232 | 3.64 | 0.1250 | 6759.9 | 0.0010 |  |  |  |  |  |
| ATL. HERRING | 1 | 493 | 0.0282 | 0.0098 | 8.28 | 0.0020 | 3853.9 | 0.0010 |  |  |  |  |  |
| GASPEREAU | 0 | 27 | 0.2495 | 0.2339 | 1.14 | 0.0520 | 297.1 | 0.0020 |  |  |  |  |  |
| RAINBOW SMELT | 0 | 66 | -0.1147 | 0.0773 | 2.20 | 0.2610 |  |  |  |  |  |  |  |
| CAPELIN | 0 | 311 | 0.0954 | 0.0260 | 13.47 | 0.0480 | 3306.3 | 0.7780 | 429 | -0.0510 | 0.0417 | 1.50 | 0.4640 |
| CAPELIN | 1 | 311 |  |  |  |  | 2234.2 | 0.3260 | 411 | 0.0566 | 0.0356 | 2.52 | 0.2040 |


| Species | outlier | length effect FIXED - September |  |  |  |  | Survey effect |  |  | length effect FIXED - September \& July |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P | Deviance | $P$ | DF | July $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P |
| ATL. MACKEREL | 0 | 42 | -0.3207 | 0.0953 | 11.32 | 0.6130 | 97.8 | 0.0040 |  |  |  |  |  |
| LONGFIN HAKE | 0 | 29 |  |  |  |  |  |  | 200 | 0.0025 | 0.0211 | 0.01 | 0.9460 |
| FOURBEARD ROCKLING | 0 | 77 | 0.0022 | 0.0351 | 0.00 | 0.9580 | 191.3 | 0.2240 | 140 | -0.0089 | 0.0298 | 0.09 | 0.7430 |
| FOURBEARD ROCKLING | 1 | 60 | 0.0260 | 0.0418 | 0.39 | 0.4770 | 167.1 | 0.9470 | 123 | -0.0024 | 0.0339 | 0.00 | 0.9330 |
| GREENLAND COD | 0 | 42 | -0.1001 | 0.0429 | 5.45 | 0.0060 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 156 | 0.0059 | 0.0088 | 0.45 | 0.5390 | 777.1 | 0.3440 | 561 | 0.0064 | 0.0045 | 2.07 | 0.8370 |
| SMOOTH SKATE | 0 | 57 | 0.0292 | 0.0191 | 2.33 | 0.1290 | 237.5 | 0.9630 | 165 | 0.0095 | 0.0108 | 0.78 | 0.4430 |
| SMOOTH SKATE | 1 | 57 |  |  |  |  | 206.3 | 0.4860 | 146 | 0.0078 | 0.0109 | 0.51 | 0.4710 |
| WINTER SKATE | 0 | 22 |  |  |  |  |  |  | 39 | -0.0091 | 0.0164 | 0.31 | 0.9120 |
| ATLANTIC HAGFISH | 0 | 38 | -0.0450 | 0.0588 | 0.58 | 0.8030 | 183.7 | 0.0160 |  |  |  |  |  |
| LONGHORN SCULPIN | 0 | 88 | -0.0733 | 0.0374 | 3.84 | 0.0640 | 963.9 | 0.0010 |  |  |  |  |  |
| LONGHORN SCULPIN | 1 | 61 | -0.0394 | 0.0442 | 0.79 | 0.4000 | 914.4 | 0.0010 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 95 | 0.0134 | 0.0258 | 0.27 | 0.7440 | 153.8 | 0.9240 | 108 | 0.0111 | 0.0244 | 0.21 | 0.8180 |
| SHORTHORN SCULPIN | 1 | 83 | 0.0345 | 0.0290 | 1.41 | 0.1430 | 133.5 | 0.6650 | 96 | 0.0294 | 0.0270 | 1.19 | 0.2380 |
| ARCTIC STAGHORN | 0 | 79 | -0.0656 | 0.0423 | 2.40 | 0.5600 |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 219 | 0.0591 | 0.0359 | 2.70 | 0.1150 | 726.2 | 0.2750 | 399 | 0.0873 | 0.0289 | 9.13 | 0.0240 |
| ARCTIC HOOKEAR | 0 | 32 | 0.9451 | 0.5784 | 2.67 | 0.8720 | 99.5 | 0.5550 | 33 | 0.8515 | 0.5428 | 2.46 | 0.8880 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 120 | 0.0667 | 0.0588 | 1.29 | 0.1880 | 200.1 | 0.0420 |  |  |  |  |  |
| SEA RAVEN | 0 | 34 | 0.0296 | 0.0420 | 0.50 | 0.4410 | 498.0 | 0.2760 | 325 | -0.0331 | 0.0098 | 11.29 | 0.3050 |
| SEA RAVEN | 1 | 34 |  |  |  |  | 424.0 | 0.2420 | 291 | -0.0041 | 0.0117 | 0.12 | 0.8810 |
| ALLIGATORFISH | 0 | 122 | 0.0594 | 0.0612 | 0.94 | 0.4170 | 372.4 | 0.7400 | 258 | 0.1760 | 0.0416 | 17.89 | 0.0020 |
| ARCTIC | 0 | 69 | -0.1939 | 0.2394 | 0.66 | 0.7650 |  |  | 69 | -0.1939 | 0.2394 | 0.66 | 0.7810 |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 142 | 0.0610 | 0.0281 | 4.71 | 0.1880 | 360.8 | 0.0030 |  |  |  |  |  |
| MARLIN-SPIKE | - | 60 | 0.0538 | 0.0346 | 2.42 | 0.1350 | 190.7 | 0.6830 | 123 | 0.0275 | 0.0219 | 1.58 | 0.1750 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 69 | -0.0168 | 0.1137 | 0.02 | 0.8570 | 172.5 | 0.1890 | 117 | -0.0320 | 0.0890 | 0.13 | 0.6660 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 135 | 0.0079 | 0.0243 | 0.10 | 0.7310 | 235.8 | 0.2340 | 146 | 0.0181 | 0.0230 | 0.62 | 0.3100 |
| FISH DOCTOR | 0 | 36 | 0.1383 | 0.1694 | 0.67 | 0.3690 | 83.8 | 0.7050 | 37 | 0.1146 | 0.1638 | 0.49 | 0.5020 |
| LAVAL`S EELPOUT & 0 & 243 & 0.0000 & 0.0076 & 0.00 & 0.9930 & 340.8 & 0.1060 & 247 & 0.0022 & 0.0075 & 0.09 & 0.7400 \\ \hline LAVAL`S EELPOUT | 1 | 229 | 0.0041 | 0.0076 | 0.29 | 0.6010 | 317.3 | 0.0930 | 233 | 0.0064 | 0.0075 | 0.71 | 0.5570 |


| Species | outlier | length effect FIXED - September |  |  |  |  | Survey effect |  | length effect FIXED - September \& July |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P | Deviance | P | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | P |
| SNAKEBLENNY | 0 | 96 | -0.0057 | 0.0269 | 0.04 | 0.8120 | 345.9 | 0.2580 | 241 | 0.0268 | 0.0155 | 3.01 | 0.1100 |
| DAUBED SHANNY | 0 | 298 | 0.1868 | 0.0317 | 34.84 | 0.0010 | 1083.0 | 0.0010 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 151 | -0.1150 | 0.0426 | 7.28 | 0.1810 |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 137 | 0.0641 | 0.0376 | 2.91 | 0.1580 |  |  |  |  |  |  |  |
| OCEAN POUT(COMMON) | 1 | 20 | -0.0360 | 0.0411 | 0.77 | 0.2670 | 172.9 | 0.0730 | 127 | -0.0251 | 0.0141 | 3.18 | 0.5020 |
| VAHL'S EELPOUT | 0 | 137 | 0.0319 | 0.0136 | 5.47 | 0.2900 | 534.4 | 0.8150 | 348 | 0.0064 | 0.0092 | 0.48 | 0.5370 |
| VAHL'S EELPOUT | 1 | 94 | 0.0105 | 0.0170 | 0.38 | 0.5990 | 444.5 | 0.1080 | 305 | -0.0022 | 0.0097 | 0.05 | 0.7900 |
| WHITE BARRACUDINA | 0 | 17 | 0.1914 | 0.1977 | 0.94 | 0.3710 | 242.5 | 0.9230 | 72 | -0.0827 | 0.0586 | 1.99 | 0.5250 |
| ATL. HOOKEAR | 0 | 76 | -0.0019 | 0.1776 | 0.00 | 0.9900 | 220.3 | 0.0850 | 122 | -0.1083 | 0.1290 | 0.70 | 0.4620 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 144 | 0.0048 | 0.0058 | 0.68 | 0.5080 | 311.9 | 0.0020 |  |  |  |  |  |
| Hyas coarctatus | 0 | 770 | 0.0025 | 0.0035 | 0.52 | 0.3970 | 2015.9 | 0.0060 |  |  |  |  |  |
| Hyas coarctatus | 1 | 740 | 0.0022 | 0.0035 | 0.39 | 0.4480 | 1934.1 | 0.0010 |  |  |  |  |  |
| NORTHERN STONE | 0 | 37 | 0.0208 | 0.0158 | 1.73 | 0.3900 | 165.5 | 0.1180 | 130 | 0.0114 | 0.0074 | 2.34 | 0.1160 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 3241 | -0.0010 | 0.0011 | 0.81 | 0.8420 | 9210.4 | 0.0010 |  |  |  |  |  |
| Hyas araneus | 0 | 152 | 0.0246 | 0.0083 | 8.71 | 0.0280 | 524.0 | 0.0010 |  |  |  |  |  |
| Hyas araneus | 1 | 138 | 0.0313 | 0.0098 | 10.10 | 0.0470 | 501.3 | 0.0010 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 61 | 0.1698 | 0.1163 | 2.13 | 0.2070 | 1476.7 | 0.5190 | 457 | -0.0611 | 0.0232 | 6.96 | 0.1920 |

Table 7. Results of fixed effects model analyses (all relevant set pairs) testing for (1) a length-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a length-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded ( $=1$ ) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

|  |  |  | 1. September survey data only |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | P | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| ATL. COD | 0 | 1324 | 0.0031 | 0.0026 | 1.46 | 0.452 | 4445.3 | 0.010 |  |  |  |  |  |
| ATL. COD | 1 | 1254 | 0.0001 | 0.0025 | 0.00 | 0.947 | 3911.3 | 0.001 |  |  |  |  |  |
| WHITE HAKE | 0 | 306 | -0.0045 | 0.0079 | 0.33 | 0.586 | 643.2 | 1.000 | 460 | -0.0083 | 0.0058 | 2.09 | 0.258 |
| REDFISH (SEBASTES SP.) | 0 | 214 | -0.0041 | 0.0076 | 0.30 | 0.789 | 7170.8 | 0.240 | 1515 | 0.0023 | 0.0033 | 0.46 | 0.547 |
| REDFISH (SEBASTES SP.) | 1 | 214 | -0.0041 | 0.0076 | 0.30 | 0.819 | 5999.0 | 0.543 | 1466 | -0.0021 | 0.0030 | 0.49 | 0.578 |
| ATLANTIC HALIBUT | 0 | 24 |  |  |  |  | 197.5 | 0.380 | 145 | -0.0025 | 0.0064 | 0.15 | 0.313 |
| GREENLAND HALIBUT | 0 | 281 | -0.0096 | 0.0057 | 2.88 | 0.385 | 1339.9 | 0.360 | 669 | -0.0059 | 0.0042 | 1.97 | 0.187 |
| GREENLAND HALIBUT | 1 | 281 |  |  |  |  | 1165.0 | 0.009 |  |  |  |  |  |
| AMERICAN PLAICE | 0 | 2100 | -0.0057 | 0.0029 | 3.79 | 0.115 | 6888.3 | 0.420 | 4000 | -0.0123 | 0.0022 | 32.4 | 0.001 |
| AMERICAN PLAICE | 1 | 2100 |  |  |  |  | 6605.0 | 0.028 |  |  |  |  |  |
| WITCH FLOUNDER | 0 | 253 | -0.0037 | 0.0150 | 0.06 | 0.802 | 1668.1 | 0.040 |  |  |  |  |  |
| WITCH FLOUNDER | 1 | 215 | -0.0096 | 0.0144 | 0.44 | 0.507 | 1506.5 | 0.010 |  |  |  |  |  |
| YELLOWTAIL FLOUNDER | 0 | 377 | -0.0255 | 0.0097 | 7.00 | 0.242 | 1960.2 | 0.400 | 1103 | -0.0087 | 0.0049 | 3.24 | 0.682 |
| YELLOWTAIL FLOUNDER | 1 | 332 | -0.0065 | 0.0104 | 0.39 | 0.740 | 1684.8 | 0.378 | 1016 | -0.0038 | 0.0050 | 0.56 | 0.873 |
| WINTER FLOUNDER | 0 | 262 | -0.0103 | 0.0106 | 0.95 | 0.691 | 1102.2 | 0.061 | 524 | -0.0181 | 0.0071 | 6.61 | 0.497 |
| WINTER FLOUNDER | 1 | 190 | 0.0389 | 0.0096 | 16.2 | 0.008 | 729.5 | 0.002 |  |  |  |  |  |
| STRIPED ATL WOLFFISH | 0 | 4 |  |  |  |  | 198.5 | 0.580 | 139 | 0.0064 | 0.0078 | 0.68 | 0.330 |
| ATL. HERRING | 0 | 606 | 0.0439 | 0.0229 | 3.66 | 0.104 | 6808.6 | 0.010 |  |  |  |  |  |
| ATL. HERRING | 1 | 507 | 0.0279 | 0.0098 | 8.12 | 0.001 | 3900.6 | 0.001 |  |  |  |  |  |
| CAPELIN | 0 | 322 | 0.1222 | 0.0253 | 23.4 | 0.027 | 3471.8 | 0.710 | 445 | -0.0133 | 0.0394 | 0.11 | 0.816 |
| CAPELIN | 1 | 322 |  |  |  |  | 2351.0 | 0.354 | 427 | 0.0861 | 0.0339 | 6.46 | 0.043 |
| ATL. MACKEREL | 0 | 47 | -0.3066 | 0.0825 | 13.8 | 0.283 | 115.4 | 0.010 |  |  |  |  |  |
| LONGFIN HAKE | 0 | 29 |  |  |  |  | 414.4 | 0.120 | 202 | 0.0023 | 0.0211 | 0.01 | 0.942 |
| FOURBEARD ROCKLING | 0 | 78 | 0.0023 | 0.0350 | 0.00 | 0.956 | 198.2 | 0.200 | 147 | -0.0120 | 0.0293 | 0.17 | 0.673 |
| FOURBEARD ROCKLING | 1 | 61 | 0.0264 | 0.0416 | 0.40 | 0.446 | 174.2 | 0.904 | 130 | -0.0065 | 0.0333 | 0.04 | 0.784 |


|  |  |  | 1. September survey data only |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | P | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| GREENLAND COD | 0 | 58 | -0.0219 | 0.0294 | 0.55 | 0.376 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 156 | 0.0059 | 0.0088 | 0.45 | 0.557 | 778.0 | 0.320 | 563 | 0.0065 | 0.0044 | 2.16 | 0.852 |
| SMOOTH SKATE | 0 | 57 | 0.0292 | 0.0191 | 2.33 | 0.112 | 242.8 | 1.000 | 168 | 0.0087 | 0.0108 | 0.65 | 0.472 |
| SMOOTH SKATE | 1 | 57 |  |  |  |  | 211.1 | 0.442 | 149 | 0.0070 | 0.0109 | 0.42 | 0.444 |
| WINTER SKATE | 0 | 22 |  |  |  |  | 53.7 | 0.950 | 39 | -0.0091 | 0.0164 | 0.31 | 0.905 |
| ATLANTIC HAGFISH | 0 | 38 | -0.0450 | 0.0588 | 0.58 | 0.810 | 186.7 | 0.040 |  |  |  |  |  |
| LONGHORN SCULPIN | 0 | 93 | -0.0681 | 0.0358 | 3.62 | 0.064 | 994.3 | 0.010 |  |  |  |  |  |
| LONGHORN SCULPIN | 1 | 66 | -0.0339 | 0.0420 | 0.65 | 0.454 | 944.8 | 0.001 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 97 | 0.0162 | 0.0257 | 0.40 | 0.603 | 156.4 | 0.890 | 110 | 0.0137 | 0.0244 | 0.32 | 0.763 |
| SHORTHORN SCULPIN | 1 | 85 | 0.0378 | 0.0290 | 1.70 | 0.078 | 136.0 | 0.674 | 98 | 0.0324 | 0.0270 | 1.44 | 0.156 |
| ARCTIC STAGHORN | 0 | 104 | -0.0761 | 0.0381 | 4.00 | 0.537 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 93 | -0.0757 | 0.0396 | 3.65 | 0.523 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 237 | 0.0491 | 0.0339 | 2.10 | 0.122 | 754.1 | 0.470 | 418 | 0.0807 | 0.0278 | 8.40 | 0.017 |
| ARCTIC HOOKEAR | 0 | 36 | 1.0586 | 0.5350 | 3.92 | 0.611 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 128 | 0.0811 | 0.0578 | 1.97 | 0.051 |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 35 | 0.0295 | 0.0418 | 0.50 | 0.468 | 510.0 | 0.300 | 334 | -0.0334 | 0.0098 | 11.7 | 0.267 |
| SEA RAVEN | 1 | 35 |  |  |  |  | 436.2 | 0.198 | 300 | -0.0051 | 0.0115 | 0.20 | 0.819 |
| ALLIGATORFISH | 0 | 135 | 0.0525 | 0.0589 | 0.79 | 0.375 | 393.6 | 0.410 | 273 | 0.1620 | 0.0402 | 16.2 | 0.002 |
| ARCTIC ALLIGATORFISH | 0 | 73 | -0.2108 | 0.2334 | 0.82 | 0.798 |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 143 | 0.0615 | 0.0280 | 4.81 | 0.190 | 367.2 | 0.010 |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 60 | 0.0538 | 0.0346 | 2.42 | 0.128 | 202.8 | 0.480 | 133 | 0.0303 | 0.0200 | 2.28 | 0.118 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY LUMPSUCKER | 0 | 75 | -0.0098 | 0.1091 | 0.01 | 0.914 | 185.8 | 0.300 | 124 | -0.0334 | 0.0866 | 0.15 | 0.616 |
| DUSKY SEASNAIL | 0 | 148 | 0.0085 | 0.0234 | 0.13 | 0.725 | 250.1 | 0.280 | 159 | 0.0183 | 0.0222 | 0.68 | 0.275 |
| NORTHERN SAND LANCE | 0 | 23 |  |  |  |  | 2806.8 | 0.010 |  |  |  |  |  |
| FISH DOCTOR | 0 | 40 | 0.0744 | 0.1512 | 0.24 | 0.702 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT & 0 & 278 & 0.0007 & 0.0072 & 0.01 & 0.928 & & & & & & & \\ \hline LAVAL`S EELPOUT | 1 | 264 | 0.0046 | 0.0072 | 0.40 | 0.698 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 96 | -0.0057 | 0.0269 | 0.04 | 0.813 | 350.7 | 0.290 | 245 | 0.0255 | 0.0154 | 2.74 | 0.124 |
| DAUBED SHANNY | 0 | 315 | 0.1986 | 0.0318 | 39.0 | 0.001 | 1163.5 | 0.010 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 180 | -0.0767 | 0.0348 | 4.85 | 0.188 |  |  |  |  |  |  |  |


| Species | outlier | DF | 1. September survey data only |  |  |  | 2. Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | $\boldsymbol{P}$ | DF | $\beta_{\text {length }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| STOUT EELBLENNY | 0 | 141 | 0.0650 | 0.0373 | 3.04 | 0.192 |  |  |  |  |  |  |  |
| VAHL'S EELPOUT | 0 | 137 | 0.0319 | 0.0136 | 5.47 | 0.287 | 550.8 | 0.790 | 363 | 0.0069 | 0.0090 | 0.58 | 0.471 |
| VAHL'S EELPOUT | 1 | 94 | 0.0105 | 0.0170 | 0.38 | 0.577 | 461.0 | 0.093 | 320 | -0.0010 | 0.0094 | 0.01 | 0.886 |
| WHITE BARRACUDINA | 0 | 17 |  |  |  |  | 242.5 | 0.910 | 72 | -0.0827 | 0.0586 | 1.99 | 0.525 |
| ATL HOOKEAR SCULPIN | 0 | 85 | 0.0380 | 0.1723 | 0.05 | 0.825 | 337.2 | 0.058 | 132 | -0.0745 | 0.1256 | 0.35 | 0.611 |
| ATL ROCK CRAB | 0 | 145 | 0.0050 | 0.0058 | 0.77 | 0.499 | 317.9 | 0.010 |  |  |  |  |  |
| Hyas coarctatus | 0 | 913 | 0.0033 | 0.0032 | 1.08 | 0.284 | 2250.9 | 0.010 |  |  |  |  |  |
| Hyas coarctatus | 1 | 846 | 0.0049 | 0.0032 | 2.29 | 0.176 | 2097.5 | 0.001 |  |  |  |  |  |
| NORTHERN STONE CRAB | 0 | 37 | 0.0208 | 0.0158 | 1.73 | 0.368 | 169.4 | 0.120 | 135 | 0.0120 | 0.0073 | 2.73 | 0.083 |
| SNOW CRAB | 0 | 3451 | 0.0006 | 0.0011 | 0.32 | 0.852 | 9754.1 | 0.010 |  |  |  |  |  |
| Hyas araneus | 0 | 173 | 0.0258 | 0.0078 | 10.9 | 0.004 | 563.0 | 0.010 |  |  |  |  |  |
| Hyas araneus | 1 | 159 | 0.0309 | 0.0088 | 12.3 | 0.020 | 541.6 | 0.001 |  |  |  |  |  |
| SHORT-FIN SQUID | 0 | 62 | 0.1695 | 0.1134 | 2.23 | 0.209 | 1489.9 | 0.480 | 466 | -0.0613 | 0.0230 | 7.11 | 0.216 |

Table 8. Results of mixed effects model analyses, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a length-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a length-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value $=0$ ) or excluded (=1) from the analysis. $P$ is the probability value for the $t$-statistic in (1) and (3) and is the probability value for the $F$-statistic based on Type-III tests of fixed effects in (2).

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{\text {length }}$ | SE | $t$ | P | F | P | DF | $\beta_{\text {length }}$ | SE | $t$ | P |
| ATL. COD | 0 | 391 | 0.0035 | 0.0045 | 0.775 | 0.4391 | 1.208 | 0.2722 | 691 | 0.0002 | 0.0037 | 0.047 | 0.9623 |
| ATL. COD | 1 | 372 | 0.0015 | 0.0044 | 0.340 | 0.7337 | 1.259 | 0.2622 | 672 | -0.0008 | 0.0037 | -0.203 | 0.8393 |
| WHITE HAKE | 0 | 139 | -0.0030 | 0.0099 | -0.306 | 0.7598 | 0.021 | 0.8849 | 198 | -0.0038 | 0.0082 | -0.467 | 0.6412 |
| REDFISH <br> (SEBASTES SP.) | 0 | 90 | 0.0078 | 0.0158 | 0.491 | 0.6250 | 0.130 | 0.7186 | 571 | 0.0185 | 0.0084 | 2.207 | 0.0277 |
| REDFISH <br> (SEBASTES SP.) | 1 | 90 |  |  |  |  | 0.018 | 0.8938 | 554 | 0.0163 | 0.0080 | 2.027 | 0.0431 |
| GREENLAND HALIBUT | 0 | 243 | -0.0044 | 0.0063 | -0.707 | 0.4800 | 0.114 | 0.7352 | 565 | -0.0066 | 0.0063 | -1.046 | 0.2962 |
| GREENLAND HALIBUT | 1 | 243 |  |  |  |  | 0.114 | 0.7352 | 565 | -0.0066 | 0.0063 | -1.046 | 0.2962 |
| AMERICAN PLAICE | 0 | 463 | -0.0062 | 0.0041 | -1.519 | 0.1295 | 3.774 | 0.0524 | 938 | -0.0107 | 0.0032 | -3.355 | 0.0008 |
| AMERICAN PLAICE | 1 | 463 |  |  |  |  | 2.150 | 0.1429 | 930 | -0.0097 | 0.0031 | -3.098 | 0.0020 |
| WITCH FLOUNDER | 0 | 230 | -0.0041 | 0.0163 | -0.251 | 0.8023 | 2.864 | 0.0909 | 860 | -0.0068 | 0.0077 | -0.885 | 0.3763 |
| WITCH FLOUNDER | 1 | 194 | -0.0031 | 0.0164 | -0.190 | 0.8497 | 2.341 | 0.1264 | 824 | -0.0078 | 0.0075 | -1.044 | 0.2970 |
| YELLOWTAIL FLOUNDER | 0 | 311 | -0.0124 | 0.0158 | -0.785 | 0.4331 | 0.010 | 0.9219 | 960 | -0.0032 | 0.0089 | -0.362 | 0.7174 |
| YELLOWTAIL FLOUNDER | 1 | 268 | -0.0001 | 0.0171 | -0.007 | 0.9941 | 0.055 | 0.8143 | 877 | 0.0004 | 0.0093 | 0.042 | 0.9666 |
| WINTER FLOUNDER | 0 | 236 | 0.0015 | 0.0171 | 0.085 | 0.9322 | 0.014 | 0.9068 | 439 | -0.0032 | 0.0128 | -0.249 | 0.8034 |
| WINTER FLOUNDER | 1 | 167 | 0.0177 | 0.0190 | 0.931 | 0.3532 | 0.065 | 0.7990 | 370 | 0.0066 | 0.0136 | 0.484 | 0.6285 |
| ATL. HERRING | 0 | 537 | 0.0438 | 0.0208 | 2.105 | 0.0358 | 0.710 | 0.3998 | 1066 | 0.0392 | 0.0186 | 2.108 | 0.0353 |
| ATL. HERRING | 1 | 443 | 0.0356 | 0.0174 | 2.045 | 0.0415 | 0.144 | 0.7042 | 972 | 0.0267 | 0.0177 | 1.507 | 0.1323 |
| GASPEREAU | 0 | 21 | 0.2558 | 0.1001 |  |  | 1.077 | 0.3015 | 117 | 0.1206 | 0.0511 | 2.359 | 0.0200 |
| CAPELIN | 0 | 253 | 0.0076 | 0.0427 | 0.179 | 0.8581 | 0.598 | 0.4397 | 343 | -0.0124 | 0.0385 | -0.322 | 0.7476 |
| CAPELIN | 1 | 253 |  |  |  |  | 1.215 | 0.2712 | 327 | -0.0068 | 0.0384 | -0.177 | 0.8594 |


|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{\text {length }}$ | SE | $t$ | P | F | P | DF | $\beta_{\text {length }}$ | SE | $t$ | P |
| THORNY SKATE | 0 | 135 | 0.0067 | 0.0102 | 0.662 | 0.5092 | 0.230 | 0.6321 | 463 | 0.0016 | 0.0057 | 0.287 | 0.7740 |
| SMOOTH SKATE | 0 | 47 | 0.0358 | 0.0243 |  |  | 0.190 | 0.6638 | 118 | 0.0101 | 0.0138 | 0.733 | 0.4652 |
| SMOOTH SKATE | 1 | 47 |  |  |  |  | 0.634 | 0.4276 | 101 | 0.0096 | 0.0129 | 0.739 | 0.4615 |
| ATLANTIC | 0 | 27 | -0.0296 | 0.0625 |  |  | 1.036 | 0.3120 | 78 | -0.0165 | 0.0382 | -0.431 | 0.6675 |
| HAGFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| LONGHORN | 0 | 70 | -0.0425 | 0.0393 | -1.083 | 0.2826 | 9.342 | 0.0024 |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| LONGHORN | 1 | 45 | -0.0324 | 0.0477 | -0.679 | 0.5006 | 9.810 | 0.0018 |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SHORTHORN | 0 | 65 | 0.0428 | 0.0320 | 1.337 | 0.1860 | 0.009 | 0.9266 | 71 | 0.0382 | 0.0296 | 1.291 | 0.2009 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SHORTHORN | 1 | 54 | 0.0393 | 0.0307 | 1.281 | 0.2058 | 0.191 | 0.6640 | 60 | 0.0336 | 0.0283 | 1.186 | 0.2405 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 0 | 48 | 0.0208 | 0.0528 | 0.393 | 0.6961 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 48 |  |  |  |  |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE | 0 | 170 | 0.0405 | 0.0441 | 0.918 | 0.3597 | 0.455 | 0.5003 | 299 | 0.0918 | 0.0369 | 2.488 | 0.0134 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE | 0 | 82 | 0.1706 | 0.0715 | 2.385 | 0.0194 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 22 | 0.0297 | 0.0426 |  |  | 1.001 | 0.3179 | 254 | -0.0098 | 0.0114 | -0.862 | 0.3893 |
| SEA RAVEN | 1 | 22 |  |  |  |  | 1.506 | 0.2210 | 222 | -0.0040 | 0.0121 | -0.330 | 0.7415 |
| ALLIGATORFISH | 0 | 84 | 0.0910 | 0.0671 | 1.357 | 0.1784 | 0.825 | 0.3651 | 167 | 0.1872 | 0.0472 | 3.962 | 0.0001 |
| ATL SEA POACHER | 0 | 106 | 0.0587 | 0.0301 | 1.954 | 0.0534 | 5.480 | 0.0203 |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 53 | 0.0310 | 0.0431 | 0.720 | 0.4750 | 0.216 | 0.6430 | 101 | 0.0135 | 0.0283 | 0.478 | 0.6338 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 46 | -0.0285 | 0.1257 | -0.227 | 0.8217 | 0.851 | 0.3597 | 69 | -0.0238 | 0.0930 | -0.256 | 0.7990 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 97 | 0.0053 | 0.0295 | 0.178 | 0.8588 | 1.213 | 0.2734 | 99 | 0.0205 | 0.0269 | 0.764 | 0.4468 |
| NORTHERN SAND | 0 | 12 | 0.3891 | 0.2251 |  |  | 2.203 | 0.1388 | 301 | 0.1084 | 0.0392 | 2.767 | 0.0060 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT & 0 & 200 & -0.0052 & 0.0097 & -0.536 & 0.5922 & & & & & & & \\ \hline LAVAL`S EELPOUT | 1 | 187 | -0.0010 | 0.0090 | -0.108 | 0.9140 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 76 | 0.0097 | 0.0291 | 0.335 | 0.7386 | 0.008 | 0.9297 | 198 | 0.0212 | 0.0201 | 1.052 | 0.2940 |
| DAUBED SHANNY | 0 | 233 | 0.1167 | 0.0406 | 2.877 | 0.0044 | 4.145 | 0.0425 |  |  |  |  |  |


|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{\text {length }}$ | SE | $t$ | P | F | P | DF | $\beta_{\text {length }}$ | SE | , | P |
| 4-LINE SNAKE | 0 | 127 | -0.0733 | 0.0585 | -1.253 | 0.2126 |  |  |  |  |  |  |  |
| BLENNY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 109 | 0.0960 | 0.0538 | 1.785 | 0.0770 |  |  |  |  |  |  |  |
| OCEAN | 0 | 12 | -0.0349 | 0.0411 |  |  | 3.027 | 0.0847 | 108 | -0.0170 | 0.0153 | -1.110 | 0.2695 |
| POUT(COMMON) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| OCEAN | 1 | 12 |  |  |  |  | 3.263 | 0.0742 | 92 | -0.0241 | 0.0141 | -1.715 | 0.0896 |
| POUT(COMMON) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| VAHL'S EELPOUT | 0 | 120 | 0.0084 | 0.0238 | 0.353 | 0.7247 | 0.017 | 0.8974 | 303 | 0.0011 | 0.0138 | 0.081 | 0.9358 |
| VAHL'S EELPOUT | 1 | 79 | 0.0181 | 0.0193 | 0.936 | 0.3523 | 0.235 | 0.6283 | 262 | 0.0027 | 0.0136 | 0.202 | 0.8405 |
| ATL. HOOKEAR | 0 | 44 | -0.1827 | 0.2034 | -0.898 | 0.3738 | 0.970 | 0.3288 | 67 | -0.0996 | 0.1590 | -0.626 | 0.5332 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 77 | 0.0052 | 0.0068 | 0.772 | 0.4423 | 4.537 | 0.0352 |  |  |  |  |  |
| HYAS | 0 | 340 | 0.0063 | 0.0061 | 1.034 | 0.3017 | 1.923 | 0.1660 | 576 | 0.0037 | 0.0044 | 0.831 | 0.4064 |
| COARCTATUS |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Hyas coarctatus | 1 | 329 | 0.0059 | 0.0060 | 0.997 | 0.3195 | 2.905 | 0.0889 | 565 | 0.0036 | 0.0044 | 0.813 | 0.4166 |
| NORTHERN STONE | 0 | 24 | 0.0161 | 0.0245 | 0.657 | 0.5173 | 1.512 | 0.2226 | 77 | 0.0113 | 0.0081 | 1.392 | 0.1680 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 640 | 0.0026 | 0.0019 | 1.380 | 0.1680 | 24.987 | 0.0000 |  |  |  |  |  |
| Hyas araneus | 0 | 72 | 0.0332 | 0.0141 | 2.362 | 0.0209 | 2.290 | 0.1325 | 138 | 0.0192 | 0.0102 | 1.876 | 0.0627 |
| Hyas araneus | 1 | 65 | 0.0314 | 0.0145 | 2.170 | 0.0337 | 2.996 | 0.0858 | 131 | 0.0165 | 0.0103 | 1.599 | 0.1121 |
| SHORT-FIN SQUID | 0 | 38 | 0.1532 | 0.1111 | 1.379 | 0.1759 | 0.916 | 0.3392 | 329 | 0.0076 | 0.0185 | 0.413 | 0.6797 |

Table 9. Results of mixed effects model analyses (all relevant set pairs) testing for (1) a length-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a length-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included -0 or excluded -1 from the analysis. $P$ is the probability value for the $t$-statistic in (1) and (3) and is the probability value for the F-statistic based on Type-III tests of fixed effects in (2).

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{\text {length }}$ | SE | $t$ | $\boldsymbol{P}$ | F | $\boldsymbol{P}$ | DF | $\beta_{\text {length }}$ | SE | $t$ | $\boldsymbol{P}$ |
| ATL. COD | 0 | 422 | 0.0018 | 0.0045 | 0.40 | 0.687 | 0.77 | 0.381 | 746 | -0.0002 | 0.0037 | -0.07 | 0.948 |
| ATL. COD | 1 | 403 | -0.0001 | 0.0045 | -0.01 | 0.990 | 0.79 | 0.376 | 727 | -0.0011 | 0.0037 | -0.30 | 0.764 |
| WHITE HAKE | 0 | 139 | -0.0030 | 0.0099 | -0.31 | 0.760 | 0.05 | 0.828 | 210 | -0.0075 | 0.0078 | -0.97 | 0.334 |
| REDFISH (SEBASTES SP.) | 0 | 90 | 0.0092 | 0.0158 | 0.58 | 0.563 | 0.07 | 0.798 | 597 | 0.0151 | 0.0082 | 1.85 | 0.065 |
| REDFISH (SEBASTES SP.) | 1 | 90 | 0.0092 | 0.0158 | 0.58 | 0.563 | 0.00 | 0.965 | 580 | 0.0131 | 0.0079 | 1.66 | 0.097 |
| GREENLAND HALIBUT | 0 | 243 | -0.0047 | 0.0063 | -0.75 | 0.453 | 0.97 | 0.326 | 597 | -0.0106 | 0.0066 | -1.61 | 0.108 |
| GREENLAND HALIBUT | 1 | 243 |  |  |  |  | 0.22 | 0.643 | 566 | -0.0076 | 0.0063 | -1.21 | 0.225 |
| AMERICAN PLAICE | 0 | 493 | -0.0053 | 0.0040 | -1.33 | 0.185 | 4.23 | 0.040 |  |  |  |  |  |
| WITCH FLOUNDER | 0 | 230 | -0.0041 | 0.0163 | -0.25 | 0.802 | 2.88 | 0.090 | 879 | -0.0062 | 0.0076 | -0.82 | 0.414 |
| WITCH FLOUNDER | 1 | 194 | -0.0031 | 0.0164 | -0.19 | 0.850 | 2.36 | 0.125 | 843 | -0.0072 | 0.0074 | -0.98 | 0.328 |
| YELLOWTAIL FLOUNDER | 0 | 337 | -0.0119 | 0.0148 | -0.81 | 0.421 | 0.00 | 0.966 | 990 | -0.0023 | 0.0086 | -0.27 | 0.785 |
| YELLOWTAIL FLOUNDER | 1 | 294 | -0.0001 | 0.0157 | -0.01 | 0.993 | 0.04 | 0.845 | 907 | 0.0012 | 0.0090 | 0.14 | 0.891 |
| WINTER FLOUNDER | 0 | 243 | 0.0020 | 0.0169 | 0.12 | 0.908 | 0.01 | 0.923 | 473 | 0.0011 | 0.0125 | 0.09 | 0.932 |
| WINTER FLOUNDER | 1 | 174 | 0.0182 | 0.0187 | 0.97 | 0.332 | 0.05 | 0.829 | 404 | 0.0108 | 0.0131 | 0.83 | 0.410 |
| ATL. HERRING | 0 | 547 | 0.0440 | 0.0208 | 2.12 | 0.035 | 0.96 | 0.327 | 1082 | 0.0391 | 0.0185 | 2.11 | 0.035 |
| ATL. HERRING | 1 | 453 | 0.0353 | 0.0174 | 2.03 | 0.043 | 0.27 | 0.607 | 988 | 0.0271 | 0.0177 | 1.54 | 0.125 |
| GASPEREAU | 0 | 18 |  |  |  |  | 0.97 | 0.327 | 117 | 0.1176 | 0.0510 | 2.31 | 0.023 |
| CAPELIN | 0 | 262 | 0.0289 | 0.0448 | 0.65 | 0.519 | 0.89 | 0.345 | 356 | 0.0078 | 0.0391 | 0.20 | 0.841 |
| CAPELIN | 1 | 262 |  |  |  |  | 1.50 | 0.221 | 340 | 0.0129 | 0.0392 | 0.33 | 0.741 |
| THORNY SKATE | 0 | 135 | 0.0067 | 0.0102 | 0.66 | 0.509 | 0.20 | 0.653 | 463 | 0.0019 | 0.0056 | 0.33 | 0.740 |
| SMOOTH SKATE | 0 | 47 | 0.0358 | 0.0243 | 1.47 | 0.148 | 0.28 | 0.600 | 119 | 0.0091 | 0.0138 | 0.66 | 0.511 |
| SMOOTH SKATE | 1 | 47 |  |  |  |  | 0.77 | 0.382 | 102 | 0.0087 | 0.0129 | 0.67 | 0.502 |
| ATLANTIC HAGFISH | 0 | 27 | -0.0296 | 0.0625 | -0.47 | 0.640 | 0.85 | 0.361 | 78 | -0.0156 | 0.0383 | -0.41 | 0.684 |
| LONGHORN SCULPIN | 0 | 73 | -0.0376 | 0.0387 | -0.97 | 0.334 | 9.83 | 0.002 |  |  |  |  |  |
| LONGHORN SCULPIN | 1 | 48 | -0.0260 | 0.0460 | -0.57 | 0.574 | 10.21 | 0.001 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 65 | 0.0461 | 0.0320 | 1.44 | 0.155 | 0.01 | 0.929 | 71 | 0.0411 | 0.0296 | 1.39 | 0.169 |
| SHORTHORN SCULPIN | 1 | 54 | 0.0431 | 0.0308 | 1.40 | 0.168 | 0.18 | 0.673 | 60 | 0.0369 | 0.0284 | 1.30 | 0.199 |


|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{\text {length }}$ | SE | $t$ | $\boldsymbol{P}$ | F | $\boldsymbol{P}$ | DF | $\beta_{\text {length }}$ | SE |  | $\boldsymbol{P}$ |
| ARCTIC STAGHORN | 0 | 68 | -0.0029 | 0.0524 | -0.05 | 0.957 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 1 | 58 | -0.0108 | 0.0447 | -0.24 | 0.810 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 186 | 0.0410 | 0.0406 | 1.01 | 0.314 | 0.20 | 0.652 | 315 | 0.0886 | 0.0353 | 2.51 | 0.013 |
| SPATULATE SCULPIN | 0 | 87 | 0.1795 | 0.0702 | 2.56 | 0.012 |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 22 | 0.0296 | 0.0424 | 0.70 | 0.493 | 1.28 | 0.260 | 260 | -0.0113 | 0.0112 | -1.01 | 0.314 |
| SEA RAVEN | 1 | 22 | 0.0296 | 0.0424 | 0.70 | 0.493 | 1.78 | 0.184 | 228 | -0.0050 | 0.0119 | -0.42 | 0.673 |
| ALLIGATORFISH | 0 | 95 | 0.0898 | 0.0646 | 1.39 | 0.168 | 1.74 | 0.189 | 179 | 0.1815 | 0.0464 | 3.91 | <0.00 |
|  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| ATL SEA POACHER | 0 | 106 | 0.0595 | 0.0300 | 1.98 | 0.050 | 5.58 | 0.019 |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 53 | 0.0310 | 0.0431 | 0.72 | 0.475 | 0.53 | 0.469 | 110 | 0.0228 | 0.0257 | 0.89 | 0.377 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY LUMPSUCKER | 0 | 50 | -0.0202 | 0.1204 | -0.17 | 0.867 | 0.49 | 0.484 | 73 | -0.0257 | 0.0908 | -0.28 | 0.778 |
| DUSKY SEASNAIL | 0 | 107 | 0.0061 | 0.0277 | 0.22 | 0.826 | 1.37 | 0.245 | 109 | 0.0201 | 0.0255 | 0.79 | 0.432 |
| NORTHERN SAND LANCE | 0 | 12 |  |  |  |  | 2.82 | 0.094 | 301 | 0.1075 | 0.0391 | 2.75 | 0.006 |
| LAVAL`S EELPOUT & 0 & 230 & -0.0038 & 0.0089 & -0.42 & 0.674 & & & & & & & \\ \hline LAVAL`S EELPOUT | 1 | 217 | 0.0003 | 0.0084 | 0.03 | 0.975 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 76 | 0.0097 | 0.0291 | 0.33 | 0.739 | 0.01 | 0.933 | 201 | 0.0188 | 0.0199 | 0.94 | 0.347 |
| DAUBED SHANNY | 0 | 246 | 0.1315 | 0.0406 | 3.24 | 0.001 | 4.65 | 0.032 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 153 | -0.0463 | 0.0489 | -0.95 | 0.345 |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 111 | 0.0969 | 0.0535 | 1.81 | 0.073 |  |  |  |  |  |  |  |
| OCEAN POUT(COMMON) | 0 | 12 |  |  |  |  | 2.88 | 0.092 | 115 | -0.0114 | 0.0151 | -0.75 | 0.452 |
| OCEAN POUT(COMMON) | 1 | 12 |  |  |  |  | 3.08 | 0.082 | 99 | -0.0175 | 0.0136 | -1.29 | 0.201 |
| VAHL'S EELPOUT | 0 | 120 | 0.0084 | 0.0238 | 0.35 | 0.725 | 0.05 | 0.831 | 316 | 0.0023 | 0.0134 | 0.17 | 0.863 |
| VAHL'S EELPOUT | 1 | 79 | 0.0181 | 0.0193 | 0.94 | 0.352 | 0.34 | 0.563 | 275 | 0.0039 | 0.0132 | 0.30 | 0.768 |
| ATL HOOKEAR SCULPIN | 0 | 49 | -0.1170 | 0.1951 | -0.60 | 0.551 | 0.86 | 0.357 | 72 | -0.0587 | 0.1542 | -0.38 | 0.705 |
| ATL ROCK CRAB | 0 | 77 | 0.0055 | 0.0068 | 0.82 | 0.417 | 3.87 | 0.052 | 122 | 0.0017 | 0.0066 | 0.26 | 0.793 |
| Hyas coarctatus | 0 | 386 | 0.0063 | 0.0058 | 1.08 | 0.282 | 2.78 | 0.096 | 622 | 0.0040 | 0.0043 | 0.93 | 0.353 |
| Hyas coarctatus | 1 | 365 | 0.0072 | 0.0058 | 1.26 | 0.209 | 3.92 | 0.048 |  |  |  |  |  |
| NORTHERN STONE CRAB | 0 | 24 |  |  |  |  | 1.65 | 0.203 | 80 | 0.0120 | 0.0079 | 1.51 | 0.135 |
| SNOW CRAB | 0 | 688 | 0.0037 | 0.0019 | 1.98 | 0.049 | 22.38 | <0.001 |  |  |  |  |  |
| Hyas araneus | 0 | 83 | 0.0333 | 0.0123 | 2.71 | 0.008 | 1.55 | 0.215 | 151 | 0.0198 | 0.0095 | 2.08 | 0.039 |
| Hyas araneus | 1 | 76 | 0.0312 | 0.0124 | 2.51 | 0.014 | 2.08 | 0.152 | 144 | 0.0174 | 0.0096 | 1.82 | 0.071 |
| SHORT-FIN SQUID | 0 | 38 | 0.1561 | 0.1092 | 1.43 | 0.161 | 0.75 | 0.388 | 335 | 0.0071 | 0.0183 | 0.39 | 0.699 |

Table 10. Results of fixed effects model analyses, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a depth-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a depth-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded (=1) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| ATL. COD | 0 | 80 | 0.0076 | 0.0022 | 11.86 | 0.1760 | 2092.5 | 0.2910 | 258 | 0.0065 | 0.0022 | 9.00 | 0.2670 |
| ATL. COD | 1 | 80 |  |  |  |  | 1559.8 | 0.1070 | 255 | 0.0079 | 0.0017 | 20.81 | 0.0850 |
| WHITE HAKE | 0 | 19 | -0.0008 | 0.0009 | 0.91 | 0.4710 | 141.4 | 0.5530 | 49 | -0.0002 | 0.0009 | 0.08 | 0.7760 |
| REDFISH (SEBASTES SP.) | 0 | 22 | -0.0010 | 0.0018 | 0.29 | 0.8410 | 5215.1 | 0.8430 | 254 | -0.0006 | 0.0011 | 0.32 | 0.7090 |
| REDFISH (SEBASTES SP.) | 1 | 22 |  |  |  |  | 4055.1 | 0.3640 | 251 | 0.0007 | 0.0010 | 0.59 | 0.5940 |
| HALIBUT(ATLANTIC) | 0 | 9 |  |  |  |  |  |  | 63 | 0.0014 | 0.0037 | 0.15 | 0.7730 |
| GREENLAND HALIBUT | 0 | 36 | -0.0014 | 0.0016 | 0.79 | 0.7040 | 695.9 | 0.6810 | 103 | -0.0013 | 0.0011 | 1.42 | 0.5740 |
| AMERICAN PLAICE | 0 | 88 | 0.0009 | 0.0012 | 0.60 | 0.5230 | 3022.0 | 0.9900 | 366 | -0.0001 | 0.0010 | 0.02 | 0.9250 |
| WITCH FLOUNDER | 0 | 22 | -0.0005 | 0.0027 | 0.04 | 0.9060 | 1070.9 | 0.7950 | 205 | 0.0006 | 0.0012 | 0.26 | 0.7750 |
| WITCH FLOUNDER | 1 | 21 | 0.0025 | 0.0022 | 1.22 | 0.6440 | 993.0 | 0.9370 | 204 | 0.0005 | 0.0011 | 0.20 | 0.8020 |
| YELLOWTAIL | 0 | 35 | -0.0201 | 0.0078 | 6.56 | 0.1740 | 954.6 | 0.1100 | 171 | -0.0101 | 0.0038 | 7.06 | 0.0860 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 17 | -0.0425 | 0.0296 | 2.06 | 0.4030 | 543.2 | 0.3310 | 68 | -0.0221 | 0.0091 | 5.95 | 0.2780 |
| WINTER FLOUNDER | 1 | 15 | -0.0030 | 0.0216 | 0.02 | 0.9230 | 371.9 | 0.6550 | 66 | -0.0104 | 0.0076 | 1.91 | 0.4290 |
| STRIPED ATL WOLFFISH | 0 | 4 |  |  |  |  |  |  | 60 | -0.0022 | 0.0104 | 0.04 | 0.8750 |
| ATL. HERRING | 0 | 55 | -0.0014 | 0.0116 | 0.01 | 0.9180 | 5445.8 | 0.4230 | 235 | -0.0093 | 0.0029 | 10.19 | 0.1120 |
| ATL. HERRING | 1 | 51 | -0.0069 | 0.0059 | 1.38 | 0.3530 | 3275.8 | 0.2730 | 231 | -0.0087 | 0.0020 | 18.70 | 0.0330 |
| GASPEREAU | 0 | 5 |  |  |  |  |  |  | 49 | 0.0084 | 0.0074 | 1.31 | 0.9690 |
| GASPEREAU | 1 | 4 |  |  |  |  |  |  | 48 | -0.0128 | 0.0060 | 4.59 | 0.1280 |
| CAPELIN | 0 | 58 | 0.0000 | 0.0019 | 0.00 | 0.9930 | 2177.6 | 0.1280 | 102 | 0.0010 | 0.0031 | 0.10 | 0.7810 |
| CAPELIN | 1 | 58 |  |  |  |  | 782.8 | 0.6300 | 94 | -0.0029 | 0.0018 | 2.47 | 0.2980 |
| ATL. MACKEREL | 0 | 16 | 0.0427 | 0.0291 | 2.15 | 0.2240 | 119.8 | 0.4520 | 24 | 0.0420 | 0.0223 | 3.54 | 0.1390 |
| LONGFIN HAKE | 0 | 3 |  |  |  |  |  |  | 48 | -0.0025 | 0.0033 | 0.55 | 0.5570 |
| FOURBEARD ROCKLING | 0 | 16 | 0.0034 | 0.0024 | 2.05 | 0.3320 | 121.6 | 0.5810 | 48 | 0.0023 | 0.0023 | 0.96 | 0.5270 |
| GREENLAND COD | 0 | 24 | -0.0008 | 0.0109 | 0.01 | 0.9450 |  |  |  |  |  |  |  |

|  | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | P | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| THORNY SKATE | 0 | 20 | -0.0009 | 0.0022 | 0.15 | 0.7430 | 491.1 | 0.8690 | 142 | 0.0009 | 0.0017 | 0.31 | 0.5650 |
| SMOOTH SKATE | 0 | 9 | -0.0012 | 0.0033 | 0.14 | 0.6900 | 124.5 | 0.7310 | 57 | -0.0019 | 0.0026 | 0.52 | 0.4920 |
| WINTER SKATE | 0 | 4 |  |  |  |  |  |  | 17 | -0.0038 | 0.0045 | 0.70 | 0.6530 |
| ATLANTIC HAGFISH | 0 | 10 | 0.0078 | 0.0077 | 1.04 | 0.4100 | 138.3 | 0.8560 | 51 | -0.0048 | 0.0027 | 3.07 | 0.4350 |
| LONGHORN SCULPIN | 0 | 17 | -0.0402 | 0.0275 | 2.13 | 0.2730 | 623.2 | 0.2460 | 161 | -0.0157 | 0.0046 | 11.47 | 0.0300 |
| SHORTHORN SCULPIN | 0 | 29 | -0.0073 | 0.0137 | 0.29 | 0.7090 | 73.1 | 0.7940 | 37 | -0.0053 | 0.0127 | 0.17 | 0.7710 |
| ARCTIC STAGHORN | 0 | 30 | -0.0334 | 0.0075 | 20.11 | 0.0390 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 48 | -0.0010 | 0.0048 | 0.05 | 0.8820 | 435.9 | 0.7970 | 122 | 0.0038 | 0.0038 | 1.02 | 0.4440 |
| ARCTIC HOOKEAR | 0 | 19 | -0.0542 | 0.0534 | 1.03 | 0.4050 | 89.5 | 0.3350 | 20 | -0.0497 | 0.0500 | 0.99 | 0.4200 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 37 | -0.0131 | 0.0170 | 0.60 | 0.5240 | 98.9 | 0.5490 | 40 | -0.0133 | 0.0166 | 0.64 | 0.5380 |
| SEA RAVEN | 0 | 11 | 0.0325 | 0.0333 | 0.96 | 0.4230 | 363.7 | 0.2810 | 99 | 0.0170 | 0.0111 | 2.34 | 0.3330 |
| ALLIGATORFISH | 0 | 37 | 0.0008 | 0.0048 | 0.03 | 0.8880 | 247.5 | 0.6680 | 107 | 0.0005 | 0.0039 | 0.02 | 0.9200 |
| ARCTIC | 0 | 25 | -0.0613 | 0.0520 | 1.39 | 0.2910 |  |  |  |  |  |  |  |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 35 | 0.0088 | 0.0069 | 1.61 | 0.1920 | 194.9 | 0.4060 | 77 | 0.0025 | 0.0064 | 0.16 | 0.7020 |
| THREESPINE | 0 | 12 | 0.0013 | 0.0042 | 0.10 | 0.8480 |  |  |  |  |  |  |  |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 6 |  |  |  |  |  |  | 30 | 0.0009 | 0.0032 | 0.09 | 0.7900 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 22 | 0.0194 | 0.0183 | 1.12 | 0.3350 | 90.0 | 0.2290 | 50 | 0.0100 | 0.0115 | 0.75 | 0.4130 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 37 | 0.0036 | 0.0131 | 0.08 | 0.8690 | 109.6 | 0.9350 | 48 | 0.0037 | 0.0118 | 0.10 | 0.8730 |
| NORTHERN SAND | 0 | 9 |  |  |  |  | 2787.9 | 0.8420 | 86 | 0.0116 | 0.0169 | 0.47 | 0.6910 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 11 | 0.0425 | 0.0706 | 0.36 | 0.6790 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 42 | 0.0155 | 0.0083 | 3.52 | 0.1250 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 19 | 0.0045 | 0.0068 | 0.44 | 0.5930 | 214.3 | 0.1350 | 58 | 0.0086 | 0.0079 | 1.18 | 0.2920 |
| DAUBED SHANNY | 0 | 64 | 0.0005 | 0.0028 | 0.03 | 0.8700 | 526.6 | 0.2880 | 117 | -0.0018 | 0.0029 | 0.39 | 0.6530 |
| 4-LINE SNAKE BLENNY | 0 | 23 | -0.0145 | 0.0087 | 2.82 | 0.1720 |  |  | 23 | -0.0145 | 0.0087 | 2.82 | 0.1940 |
| STOUT EELBLENNY | 0 | 27 | 0.0038 | 0.0050 | 0.56 | 0.5490 |  |  | 27 | 0.0038 | 0.0050 | 0.56 | 0.5040 |
| OCEAN POUT(COMMON) | 0 | 7 |  |  |  |  |  |  | 48 | 0.0113 | 0.0123 | 0.84 | 0.4720 |
| VAHL'S EELPOUT | 0 | 16 | -0.0004 | 0.0054 | 0.01 | 0.9630 | 335.2 | 0.9220 | 66 | -0.0015 | 0.0045 | 0.11 | 0.8350 |
| WHITE BARRACUDINA | 0 | 5 |  |  |  |  |  |  | 24 | 0.0118 | 0.0063 | 3.51 | 0.4950 |

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| ATL. HOOKEAR | 0 | 31 | -0.0008 | 0.0059 | 0.02 | 0.9270 | 251.1 | 0.6660 | 57 | -0.0001 | 0.0053 | 0.00 | 0.9840 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 14 | -0.0059 | 0.0146 | 0.16 | 0.6440 | 121.0 | 0.7220 | 40 | -0.0166 | 0.0122 | 1.84 | 0.3720 |
| Hyas coarctatus | 0 | 66 | -0.0049 | 0.0060 | 0.66 | 0.5070 | 1122.8 | 0.5530 | 154 | -0.0105 | 0.0058 | 3.33 | 0.1930 |
| Hyas coarctatus | 1 | 65 | -0.0045 | 0.0054 | 0.68 | 0.5290 | 810.9 | 0.4250 | 151 | -0.0145 | 0.0055 | 6.99 | 0.1270 |
| NORTHERN STONE | 0 | 9 |  |  |  |  | 142.7 | 0.5020 | 54 | 0.0009 | 0.0032 | 0.08 | 0.8000 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 83 | -0.0011 | 0.0034 | 0.10 | 0.8360 | 4002.1 | 0.4500 | 272 | -0.0032 | 0.0025 | 1.67 | 0.4360 |
| SNOW CRAB | 1 | 83 |  |  |  |  | 3472.4 | 0.3500 | 269 | -0.0045 | 0.0024 | 3.47 | 0.3060 |
| Hyas araneus | 0 | 42 | -0.0094 | 0.0110 | 0.73 | 0.5600 | 395.8 | 0.1870 | 90 | -0.0061 | 0.0098 | 0.39 | 0.8050 |
| Hyas araneus | 1 | 42 |  |  |  |  | 350.6 | 0.3450 | 89 | -0.0066 | 0.0092 | 0.52 | 0.7370 |
| SHORT-FIN SQUID | 0 | 22 | -0.0083 | 0.0035 | 5.64 | 0.0320 | 1248.6 | 0.8130 | 187 | -0.0008 | 0.0022 | 0.14 | 0.8630 |

Table 11. Results of fixed effects model analyses (all relevant set pairs) testing for (1) a depth-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a depth-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded (=1) from the analysis. Probability values are based on 999 permutations under the null hypothesis.

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | P | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | P |
| ATL. COD | 0 | 86 | 0.0067 | 0.0023 | 8.23 | 0.212 | 2415.67 | 0.307 | 275 | 0.0070 | 0.0022 | 10.40 | 0.195 |
| ATL. COD | 1 | 86 |  |  |  |  | 1797.4 | 0.079 | 272 | 0.0084 | 0.0018 | 22.62 | 0.070 |
| WHITE HAKE | 0 | 19 | -0.0008 | 0.0009 | 0.91 | 0.473 | 168.4 | 0.480 | 52 | -0.0006 | 0.0009 | 0.49 | 0.528 |
| REDFISH (SEBASTES SP.) | 0 | 23 | -0.0008 | 0.0017 | 0.24 | 0.865 | 5342.2 | 0.799 | 265 | -0.0007 | 0.0011 | 0.44 | 0.665 |
| REDFISH (SEBASTES SP.) | 1 | 23 |  |  |  |  | 4192.6 | 0.432 | 262 | 0.0006 | 0.0010 | 0.41 | 0.686 |
| HALIBUT(ATLANTIC) | 0 | 9 |  |  |  |  | 153.1 | 0.890 | 63 | 0.0014 | 0.0037 | 0.15 | 0.767 |
| GREENLAND HALIBUT | 0 | 37 | -0.0014 | 0.0015 | 0.86 | 0.652 | 862.6 | 0.792 | 108 | -0.0002 | 0.0011 | 0.05 | 0.912 |
| AMERICAN PLAICE | 0 | 94 | 0.0008 | 0.0012 | 0.43 | 0.592 | 3129.6 | 0.964 | 382 | -0.0003 | 0.0010 | 0.08 | 0.838 |
| WITCH FLOUNDER | 0 | 22 | -0.0005 | 0.0027 | 0.04 | 0.917 | 1100.4 | 0.777 | 213 | 0.0006 | 0.0012 | 0.27 | 0.773 |
| WITCH FLOUNDER | 1 | 21 | 0.0025 | 0.0022 | 1.22 | 0.636 | 1022.5 | 0.937 | 212 | 0.0005 | 0.0011 | 0.20 | 0.794 |
| YELLOWTAIL | 0 | 39 | -0.0197 | 0.0072 | 7.49 | 0.149 | 964.2 | 0.082 | 177 | -0.0101 | 0.0037 | 7.39 | 0.069 |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 18 | -0.0429 | 0.0288 | 2.21 | 0.384 | 559.3 | 0.278 | 73 | -0.0236 | 0.0080 | 8.77 | 0.234 |
| WINTER FLOUNDER | 1 | 16 | -0.0035 | 0.0210 | 0.03 | 0.918 | 390.6 | 0.517 | 71 | -0.0138 | 0.0068 | 4.14 | 0.253 |
| STRIPED ATL WOLFFISH | 0 | 4 |  |  |  |  | 168.3 | 0.981 | 61 | -0.0021 | 0.0105 | 0.04 | 0.878 |
| ATL. HERRING | 0 | 59 | -0.0015 | 0.0112 | 0.02 | 0.930 | 5486.0 | 0.441 | 244 | -0.0092 | 0.0029 | 10.44 | 0.114 |
| ATL. HERRING | 1 | 55 | -0.0070 | 0.0057 | 1.51 | 0.334 | 3314.1 | 0.299 | 240 | -0.0087 | 0.0020 | 19.05 | 0.027 |
| GASPEREAU | 0 | 5 |  |  |  |  |  |  | 50 | 0.0085 | 0.0073 | 1.34 | 0.965 |
| GASPEREAU | 1 | 4 |  |  |  |  |  |  | 49 | -0.0127 | 0.0059 | 4.57 | 0.124 |
| CAPELIN | 0 | 60 | 0.0007 | 0.0020 | 0.11 | 0.799 | 2294.7 | 0.126 | 107 | 0.0014 | 0.0030 | 0.20 | 0.683 |
| CAPELIN | 1 | 60 |  |  |  |  | 928.8 | 0.717 | 99 | -0.0024 | 0.0019 | 1.62 | 0.376 |
| ATL. MACKEREL | 0 | 19 | 0.0537 | 0.0267 | 4.05 | 0.101 | 136.7 | 0.279 | 28 | 0.0477 | 0.0219 | 4.74 | 0.081 |
| LONGFIN HAKE | 0 | 3 |  |  |  |  |  |  | 50 | -0.0025 | 0.0033 | 0.56 | 0.539 |
| FOURBEARD ROCKLING | 0 | 17 | 0.0035 | 0.0023 | 2.26 | 0.309 | 128.6 | 0.593 | 52 | 0.0022 | 0.0023 | 0.95 | 0.559 |
| GREENLAND COD | 0 | 27 | 0.0013 | 0.0100 | 0.02 | 0.898 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 20 | -0.0009 | 0.0022 | 0.15 | 0.723 | 492.0 | 0.853 | 144 | 0.0009 | 0.0017 | 0.31 | 0.580 |

|  | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| SMOOTH SKATE | 0 | 9 |  |  |  |  | 129.8 | 0.749 | 59 | -0.0018 | 0.0026 | 0.50 | 0.475 |
| WINTER SKATE | 0 | 4 |  |  |  |  | 40.1 | 0.756 | 17 | -0.0038 | 0.0045 | 0.70 | 0.638 |
| ATLANTIC HAGFISH | 0 | 10 | 0.0078 | 0.0077 | 1.04 | 0.414 | 141.2 | 0.868 | 52 | -0.0047 | 0.0027 | 2.99 | 0.399 |
| LONGHORN SCULPIN | 0 | 19 | -0.0458 | 0.0270 | 2.88 | 0.210 | 648.3 | 0.208 | 169 | -0.0162 | 0.0046 | 12.42 | 0.023 |
| SHORTHORN SCULPIN | 0 | 31 | -0.0093 | 0.0134 | 0.47 | 0.628 | 75.6 | 0.699 | 39 | -0.0070 | 0.0126 | 0.31 | 0.693 |
| ARCTIC STAGHORN | 0 | 35 | -0.0353 | 0.0083 | 18.18 | 0.048 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE SCULPIN | 0 | 50 | -0.0004 | 0.0048 | 0.01 | 0.951 | 439.9 | 0.746 | 125 | 0.0037 | 0.0038 | 0.97 | 0.450 |
| ARCTIC HOOKEAR | 0 | 21 | -0.0361 | 0.0408 | 0.78 | 0.485 | 96.7 | 0.428 | 22 | -0.0332 | 0.0380 | 0.76 | 0.503 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 40 | -0.0053 | 0.0163 | 0.11 | 0.793 | 113.6 | 0.779 | 43 | -0.0056 | 0.0159 | 0.12 | 0.775 |
| SEA RAVEN | 0 | 12 | 0.0332 | 0.0333 | 0.99 | 0.409 | 375.3 | 0.223 | 105 | 0.0174 | 0.0109 | 2.55 | 0.343 |
| ALLIGATORFISH | 0 | 39 | 0.0019 | 0.0049 | 0.14 | 0.739 | 258.2 | 0.797 | 110 | 0.0011 | 0.0039 | 0.07 | 0.808 |
| ARCTIC | 0 | 27 | -0.0673 | 0.0515 | 1.71 | 0.246 |  |  |  |  |  |  |  |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 36 | 0.0084 | 0.0068 | 1.50 | 0.208 | 201.5 | 0.419 | 80 | 0.0022 | 0.0064 | 0.12 | 0.754 |
| THREESPINE | 0 | 12 | 0.0013 | 0.0042 | 0.10 | 0.850 |  |  |  |  |  |  |  |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MARLIN-SPIKE | 0 | 6 |  |  |  |  |  |  | 33 | 0.0003 | 0.0028 | 0.01 | 0.927 |
| GRENADIER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 24 | 0.0193 | 0.0177 | 1.19 | 0.337 | 96.0 | 0.527 | 53 | 0.0064 | 0.0110 | 0.34 | 0.625 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 40 | 0.0049 | 0.0125 | 0.15 | 0.826 | 110.5 | 0.900 | 51 | 0.0049 | 0.0113 | 0.19 | 0.823 |
| NORTHERN SAND | 0 | 10 | 0.0408 | 0.0351 | 1.35 | 0.414 | 2788.3 | 0.814 | 87 | 0.0116 | 0.0169 | 0.47 | 0.701 |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 14 | 0.0477 | 0.0633 | 0.57 | 0.604 | 55.5 | 0.605 | 15 | 0.0366 | 0.0585 | 0.39 | 0.695 |
| LAVAL`S EELPOUT | 0 | 47 | 0.0139 | 0.0080 | 3.06 | 0.127 | 94.8 | 0.131 | 48 | 0.0159 | 0.0078 | 4.15 | 0.077 |
| SNAKEBLENNY | 0 | 19 | 0.0045 | 0.0068 | 0.44 | 0.606 | 219.6 | 0.164 | 60 | 0.0082 | 0.0078 | 1.10 | 0.342 |
| DAUBED SHANNY | 0 | 68 | 0.0020 | 0.0029 | 0.47 | 0.514 | 597.5 | 0.267 | 123 | 0.0000 | 0.0029 | 0.00 | 0.999 |
| 4-LINE SNAKE BLENNY | 0 | 26 | -0.0145 | 0.0073 | 3.95 | 0.102 |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 29 | 0.0037 | 0.0049 | 0.56 | 0.528 |  |  |  |  |  |  |  |
| OCEAN POUT(COMMON) | 0 | 7 |  |  |  |  |  |  | 50 | 0.0123 | 0.0124 | 0.98 | 0.449 |
| VAHL'S EELPOUT | 0 | 16 | -0.0004 | 0.0054 | 0.01 | 0.975 | 351.6 | 0.885 | 71 | -0.0016 | 0.0045 | 0.12 | 0.851 |
| WHITE BARRACUDINA | 0 | 5 |  |  |  |  |  |  | 24 | 0.0118 | 0.0063 | 3.51 | 0.501 |
| ATL. HOOKEAR | 0 | 35 | 0.0000 | 0.0059 | 0.00 | 0.995 | 267.3 | 0.654 | 62 | 0.0003 | 0.0052 | 0.00 | 0.975 |

| Species | 1. September survey data only |  |  |  |  |  | 2.Survey effect |  | 3.September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | P | Deviance | P | DF | $\beta_{\text {depth }}$ | SE | $\chi^{2}$ | $\boldsymbol{P}$ |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 15 | -0.0072 | 0.0142 | 0.25 | 0.590 | 127.1 | 0.731 | 43 | -0.0159 | 0.0118 | 1.81 | 0.358 |
| Hyas coarctatus | 0 | 72 | -0.0039 | 0.0060 | 0.42 | 0.623 | 1205.1 | 0.643 | 162 | -0.0094 | 0.0056 | 2.80 | 0.234 |
| Hyas coarctatus | 1 | 71 | -0.0037 | 0.0056 | 0.44 | 0.599 | 894.2 | 0.474 | 159 | -0.0134 | 0.0054 | 6.06 | 0.146 |
| NORTHERN STONE | 0 | 9 | 0.0019 | 0.0078 | 0.06 | 0.852 | 146.9 | 0.499 | 57 | 0.0010 | 0.0031 | 0.11 | 0.797 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 90 | 0.0003 | 0.0035 | 0.01 | 0.950 | 4283.1 | 0.573 | 284 | -0.0033 | 0.0025 | 1.74 | 0.471 |
| SNOW CRAB | 1 | 90 |  |  |  |  | 3777.0 | 0.486 | 281 | -0.0045 | 0.0024 | 3.45 | 0.288 |
| Hyas araneus | 0 | 47 | -0.0111 | 0.0104 | 1.12 | 0.467 | 424.9 | 0.299 | 98 | -0.0080 | 0.0095 | 0.71 | 0.731 |
| Hyas araneus | 1 | 47 |  |  |  |  | 375.3 | 0.374 | 97 | -0.0081 | 0.0089 | 0.82 | 0.627 |
| SHORT-FIN SQUID | 0 | 23 | -0.0080 | 0.0034 | 5.52 | 0.030 | 1260.5 | 0.818 | 192 | -0.0008 | 0.0022 | 0.13 | 0.870 |

Table 12. Results of mixed effects model analyses, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for (1) a depth-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a depth-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value $=0$ ) or excluded ( $=1$ ) from the analysis. $P$ is the probability value for the $t$-statistic in (1) and (3) and is the probability value for the F-statistic based on Type-III tests of fixed effects in (2).

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {depth }}$ | SE | $t$ | $\boldsymbol{P}$ | F | P | DF | $\beta_{\text {depth }}$ | SE |  | $\boldsymbol{P}$ |
| ATL. COD | 0 | 79 | 0.0006 | 0.0023 | 0.256 | 0.7984 | 0.396 | 0.6742 | 73 | 0.0020 | 0.0023 | 0.886 | 0.3784 |
| ATL. COD | 1 | 79 |  |  |  |  | 0.734 | 0.4837 | 71 | 0.0025 | 0.0022 | 1.173 | 0.2445 |
| WHITE HAKE | 0 | 19 | -0.0006 | 0.0010 | -0.592 | 0.5611 | 0.089 | 0.9149 | 15 | -0.0004 | 0.0010 | -0.425 | 0.6767 |
| REDFISH (SEBASTES SP.) | 0 | 22 | -0.0019 | 0.0019 | -0.995 | 0.3307 | 4.090 | 0.0193 |  |  |  |  |  |
| REDFISH (SEBASTES SP.) | 1 | 22 |  |  |  |  | 4.493 | 0.0133 |  |  |  |  |  |
| GREENLAND HALIBUT | 0 | 36 | -0.0029 | 0.0018 | -1.569 | 0.1253 | 0.065 | 0.9372 | 34 | -0.0006 | 0.0015 | -0.385 | 0.7024 |
| AMERICAN PLAICE | 0 | 88 | 0.0002 | 0.0009 | 0.210 | 0.8338 | 1.497 | 0.2275 | 138 | -0.0001 | 0.0009 | -0.097 | 0.9233 |
| WITCH FLOUNDER | 0 | 22 | -0.0008 | 0.0026 | -0.317 | 0.7542 | 0.875 | 0.4208 | 85 | 0.0010 | 0.0013 | 0.737 | 0.4630 |
| WITCH FLOUNDER | 1 | 21 | 0.0005 | 0.0021 | 0.243 | 0.8103 | 0.462 | 0.6314 | 85 | 0.0009 | 0.0012 | 0.772 | 0.4423 |
| YELLOWTAIL | 0 | 35 | -0.0008 | 0.0103 | -0.080 | 0.9365 | 4.059 | 0.0220 |  |  |  |  |  |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 17 | -0.0348 | 0.0176 | -1.970 | 0.0654 | 2.896 | 0.0798 | 20 | -0.0193 | 0.0092 | -2.093 | 0.0493 |
| WINTER FLOUNDER | 1 | 15 | -0.0308 | 0.0172 | -1.793 | 0.0932 | 2.481 | 0.1103 | 20 | -0.0180 | 0.0091 | -1.973 | 0.0624 |
| ATL. HERRING | 0 | 54 | -0.0086 | 0.0056 | -1.521 | 0.1340 | 1.878 | 0.1596 | 80 | -0.0064 | 0.0037 | -1.749 | 0.0842 |
| ATL. HERRING | 1 | 49 | -0.0081 | 0.0033 | -2.488 | 0.0163 | 1.695 | 0.1901 | 80 | -0.0055 | 0.0031 | -1.801 | 0.0755 |
| CAPELIN | 0 | 57 | 0.0002 | 0.0025 | 0.081 | 0.9360 | 1.974 | 0.1734 | 16 | 0.0005 | 0.0027 | 0.194 | 0.8488 |
| CAPELIN | 1 | 57 |  |  |  |  | 2.353 | 0.1454 | 11 | -0.0012 | 0.0023 | -0.498 | 0.6281 |
| LONGFIN HAKE | 0 | 3 |  |  |  |  | 0.892 | 0.4263 | 20 | 0.0030 | 0.0050 | 0.587 | 0.5635 |
| GREENLAND COD | 0 | 24 | -0.0015 | 0.0111 | -0.132 | 0.8964 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 20 | -0.0006 | 0.0023 | -0.256 | 0.8002 | 0.109 | 0.8972 | 44 | -0.0003 | 0.0013 | -0.197 | 0.8448 |
| SMOOTH SKATE | 0 | 9 |  |  |  |  | 0.268 | 0.7710 | 10 | -0.0015 | 0.0025 | -0.590 | 0.5683 |
| ATLANTIC HAGFISH | 0 | 10 | 0.0060 | 0.0072 | 0.827 | 0.4278 | 0.334 | 0.7245 | 10 | -0.0019 | 0.0031 | -0.605 | 0.5588 |
| LONGHORN SCULPIN | 0 | 17 | -0.0183 | 0.0239 | -0.768 | 0.4531 | 6.002 | 0.0041 |  |  |  |  |  |
| SHORTHORN SCULPIN | 0 | 29 | -0.0001 | 0.0151 | -0.005 | 0.9960 | 0.063 | 0.9390 | 36 | 0.0012 | 0.0142 | 0.083 | 0.9347 |
| ARCTIC STAGHORN SCULPIN | 0 | 30 | -0.0224 | 0.0132 | -1.694 | 0.1006 | 2.870 | 0.1006 | 30 | -0.0224 | 0.0132 | -1.694 | 0.1006 |

| MOUSTACHE SCULPIN | 0 | 48 | 0.0004 | 0.0058 | 0.069 | 0.9456 | 0.477 | 0.6272 | 22 | 0.0021 | 0.0040 | 0.521 | 0.6075 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ARCTIC HOOKEAR | 0 | 19 | -0.0319 | 0.0382 | -0.835 | 0.4141 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE SCULPIN | 0 | 37 | 0.0036 | 0.0161 | 0.224 | 0.8241 |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 11 | 0.0308 | 0.0321 | 0.961 | 0.3571 | 1.865 | 0.1743 | 28 | 0.0160 | 0.0102 | 1.573 | 0.1269 |
| ALLIGATORFISH | 0 | 37 | 0.0028 | 0.0057 | 0.499 | 0.6206 | 0.219 | 0.8062 | 16 | 0.0004 | 0.0040 | 0.089 | 0.9299 |
| ARCTIC | 0 | 25 | -0.0939 | 0.0669 | -1.405 | 0.1724 | 1.974 | 0.1724 | 25 | -0.0939 | 0.0669 | -1.405 | 0.1724 |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 35 | 0.0072 | 0.0072 | 1.000 | 0.3243 | 3.482 | 0.0527 | 19 | 0.0037 | 0.0057 | 0.639 | 0.5303 |
| THREESPINE | 0 | 12 | 0.0060 | 0.0073 | 0.822 | 0.4269 | 0.676 | 0.4269 | 12 | 0.0060 | 0.0073 | 0.822 | 0.4269 |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 22 | 0.0239 | 0.0205 | 1.168 | 0.2551 | 1.038 | 0.5702 | 2 | 0.0091 | 0.0125 | 0.729 | 0.5417 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 37 | -0.0010 | 0.0137 | -0.075 | 0.9405 | 0.165 |  | 1 | -0.0015 | 0.0125 | -0.124 | 0.9215 |
| FISH DOCTOR | 0 | 11 | 0.0937 | 0.0818 | 1.145 | 0.2764 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 42 | 0.0134 | 0.0093 | 1.445 | 0.1557 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 19 | 0.0048 | 0.0099 | 0.488 | 0.6312 | 1.473 | 0.2627 | 15 | 0.0106 | 0.0074 | 1.444 | 0.1694 |
| DAUBED SHANNY | 0 | 64 | -0.0014 | 0.0045 | -0.323 | 0.7477 | 6.003 | 0.0083 |  |  |  |  |  |
| 4-LINE SNAKE BLENNY | 0 | 23 | -0.0143 | 0.0138 | -1.031 | 0.3134 | 1.062 | 0.3134 | 23 | -0.0143 | 0.0138 | -1.031 | 0.3134 |
| STOUT EELBLENNY | 0 | 27 | 0.0097 | 0.0132 | 0.738 | 0.4666 | 0.545 | 0.4666 | 27 | 0.0097 | 0.0132 | 0.738 | 0.4666 |
| OCEAN POUT(COMMON) | 0 | 7 |  |  |  |  | 4.096 | 0.0468 |  |  |  |  |  |
| VAHL'S EELPOUT | 0 | 16 | 0.0004 | 0.0080 | 0.055 | 0.9570 | 0.139 | 0.8709 | 21 | -0.0007 | 0.0050 | -0.130 | 0.8977 |
| WHITE BARRACUDINA | 0 | 5 |  |  |  |  | 0.063 | 0.9401 | 6 | 0.0019 | 0.0068 | 0.286 | 0.7848 |
| ATL. HOOKEAR | 0 | 31 | 0.0009 | 0.0081 | 0.109 | 0.9142 | 0.714 | 0.6419 | 2 | 0.0008 | 0.0069 | 0.111 | 0.9217 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Hyas coarctatus | 0 | 66 | -0.0047 | 0.0050 | -0.943 | 0.3491 | 2.612 | 0.0906 | 30 | -0.0080 | 0.0048 | -1.669 | 0.1055 |
| Hyas coarctatus | 1 | 65 | -0.0042 | 0.0046 | -0.900 | 0.3715 | 5.914 | $\mathbf{0 . 0 0 7 0}$ |  |  |  |  |  |
| NORTHERN STONE | 0 | 9 | 0.0031 | 0.0115 | 0.273 | 0.7912 | 1.753 | 0.2119 | 14 | 0.0001 | 0.0029 | 0.031 | 0.9756 |
| CRAB |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SNOW CRAB | 0 | 83 | 0.0023 | 0.0020 | 1.193 | 0.2361 | 4.991 | 0.0086 |  |  |  |  |  |
| SNOW CRAB | 1 | 83 |  |  |  |  | 7.808 | 0.0007 |  |  |  |  |  |
| Hyas araneus | 0 | 42 | -0.0218 | 0.0174 | -1.256 | 0.2161 | 4.045 | 0.1095 | 5 | 0.0036 | 0.0100 | 0.362 | 0.7319 |
| SHORT-FIN SQUID | 0 | 22 | -0.0081 | 0.0034 | -2.402 | 0.0252 | 0.003 | 0.9974 | 59 | -0.0001 | 0.0027 | -0.033 | 0.9736 |
| SHORT-FIN SQUID | 1 | 22 |  |  |  |  | 0.001 | 0.9995 | 58 | -0.0001 | 0.0026 | -0.032 | 0.9745 |

Table 13. Results of mixed effects model analyses (all relevant set pairs) testing for (1) a depth-dependent difference in catchability between the CCGS Alfred Needler and the CCGS Teleost based on the September 2004 and 2005 comparative fishing experiments, (2) an interaction between the length and survey effect and (3) a depth-dependent difference in catchability based on the combined September and July comparative fishing experiments, where appropriate. The column outlier indicates whether outliers were included (value=0) or excluded ( $=1$ ) from the analysis. $P$ is the probability value for the $t$-statistic in (1) and (3) and is the probability value for the $F$-statistic based on Type-III tests of fixed effects in (2).

| Species | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | outlier | DF | $\beta_{\text {depth }}$ | SE | $t$ | $\boldsymbol{P}$ | F | $\boldsymbol{P}$ | DF | $\beta_{\text {depth }}$ | SE | $t$ | P |
| ATL. COD | 0 | 85 | -0.0008 | 0.0025 | -0.31 | 0.756 | 1.08 | 0.346 | 80 | 0.0024 | 0.0024 | 1.02 | 0.312 |
| ATL. COD | 1 | 85 |  |  |  |  | 0.55 | 0.578 | 78 | 0.0029 | 0.0023 | 1.29 | 0.202 |
| WHITE HAKE | 0 | 19 | -0.0006 | 0.0010 | -0.59 | 0.561 | 0.17 | 0.846 | 17 | -0.0005 | 0.0009 | -0.59 | 0.561 |
| REDFISH <br> (SEBASTES SP.) | 1 | 23 | -0.0015 | 0.0019 | -0.77 | 0.448 | 3.84 | 0.024 |  |  |  |  |  |
| ATLANTIC | 0 | 9 |  |  |  |  | 0.31 | 0.737 | 14 | 0.0028 | 0.0037 | 0.76 | 0.458 |
| HALIBUT |  |  |  |  |  |  |  |  |  |  |  |  |  |
| GREENLAND | 0 | 37 | -0.0030 | 0.0018 | -1.63 | 0.111 | 0.08 | 0.927 | 36 | -0.0002 | 0.0016 | -0.09 | 0.926 |
| HALIBUT |  |  |  |  |  |  |  |  |  |  |  |  |  |
| AMERICAN PLAICE | 0 | 94 | 0.0002 | 0.0009 | 0.20 | 0.841 | 1.85 | 0.161 | 144 | -0.0003 | 0.0009 | -0.33 | 0.744 |
| WITCH FLOUNDER | 0 | 22 | -0.0008 | 0.0026 | -0.32 | 0.754 | 0.49 | 0.612 | 90 | 0.0009 | 0.0013 | 0.71 | 0.480 |
| WITCH FLOUNDER | 1 | 21 | 0.0005 | 0.0021 | 0.24 | 0.810 | 0.95 | 0.389 | 90 | 0.0009 | 0.0012 | 0.75 | 0.457 |
| YELLOWTAIL | 0 | 39 | -0.0025 | 0.0093 | -0.27 | 0.791 | 4.17 | 0.020 |  |  |  |  |  |
| FLOUNDER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WINTER FLOUNDER | 0 | 18 | -0.0350 | 0.0175 | -2.00 | 0.061 | 3.29 | 0.057 | 22 | -0.0198 | 0.0088 | -2.26 | 0.034 |
| WINTER FLOUNDER | 1 | 16 | -0.0308 | 0.0170 | -1.81 | 0.089 | 2.83 | 0.081 | 22 | -0.0185 | 0.0087 | -2.13 | 0.044 |
| STRIPED ATL | 0 | 4 |  |  |  |  | 0.12 | 0.891 | 11 | -0.0037 | 0.0094 | -0.39 | 0.705 |
| WOLFFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL. HERRING | 0 | 58 | -0.0088 | 0.0056 | -1.57 | 0.122 | 1.56 | 0.217 | 82 | -0.0062 | 0.0036 | -1.70 | 0.094 |
| ATL. HERRING | 1 | 53 | -0.0082 | 0.0033 | -2.46 | 0.017 | 1.83 | 0.167 | 82 | -0.0053 | 0.0031 | -1.71 | 0.091 |
| CAPELIN | 0 | 59 | 0.0010 | 0.0029 | 0.36 | 0.717 | 2.26 | 0.135 | 18 | 0.0010 | 0.0029 | 0.36 | 0.725 |
| CAPELIN | 1 | 59 |  |  |  |  | 2.38 | 0.134 | 13 | -0.0006 | 0.0026 | -0.22 | 0.833 |
| ATL. MACKEREL | 0 | 19 | 0.0120 | 0.0436 | 0.28 | 0.785 |  |  |  |  |  |  |  |
| LONGFIN HAKE | 0 | 3 | 0.0151 | 0.0140 | 1.08 | 0.360 | 0.86 | 0.437 | 21 | 0.0028 | 0.0049 | 0.58 | 0.569 |
| FOURBEARD | 0 | 17 | 0.0019 | 0.0033 | 0.57 | 0.578 | 0.12 | 0.886 | 8 | 0.0002 | 0.0027 | 0.09 | 0.929 |
| ROCKLING |  |  |  |  |  |  |  |  |  |  |  |  |  |

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {depth }}$ | SE | $t$ | P | F | P | DF | $\beta_{\text {depth }}$ | SE |  | $\boldsymbol{P}$ |
| GREENLAND COD | 0 | 27 | 0.0012 | 0.0101 | 0.12 | 0.908 |  |  |  |  |  |  |  |
| THORNY SKATE | 0 | 20 | -0.0006 | 0.0023 | -0.26 | 0.800 | 0.09 | 0.911 | 44 | -0.0003 | 0.0013 | -0.20 | 0.844 |
| SMOOTH SKATE | 0 | 9 |  |  |  |  | 0.19 | 0.834 | 10 | -0.0014 | 0.0025 | -0.54 | 0.601 |
| ATLANTIC HAGFISH | 0 | 10 | 0.0060 | 0.0072 | 0.83 | 0.428 | 0.25 | 0.784 | 10 | -0.0017 | 0.0031 | -0.55 | 0.596 |
| LONGHORN | 0 | 19 | -0.0269 | 0.0230 | -1.17 | 0.257 | 6.15 | 0.004 |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SHORTHORN | 0 | 31 | -0.0029 | 0.0149 | -0.20 | 0.845 | 0.08 | 0.922 | 38 | -0.0014 | 0.0140 | -0.10 | 0.920 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC STAGHORN | 0 | 35 | -0.0193 | 0.0142 | -1.36 | 0.182 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| MOUSTACHE | 0 | 50 | 0.0012 | 0.0057 | 0.22 | 0.830 | 0.39 | 0.680 | 22 | 0.0022 | 0.0040 | 0.56 | 0.581 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ARCTIC HOOKEAR | 0 | 21 | -0.0284 | 0.0341 | -0.83 | 0.415 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SPATULATE | 0 | 40 | 0.0134 | 0.0170 | 0.79 | 0.436 |  |  |  |  |  |  |  |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SEA RAVEN | 0 | 12 | 0.0316 | 0.0324 | 0.98 | 0.349 | 2.26 | 0.122 | 31 | 0.0169 | 0.0099 | 1.70 | 0.099 |
| ALLIGATORFISH | 0 | 39 | 0.0039 | 0.0056 | 0.70 | 0.491 | 0.08 | 0.928 | 16 | 0.0010 | 0.0040 | 0.25 | 0.807 |
| ARCTIC | 0 | 27 | -0.0996 | 0.0670 | -1.49 | 0.149 |  |  | 27 | -0.0996 | 0.0670 | -1.49 | 0.149 |
| ALLIGATORFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SEA POACHER | 0 | 36 | 0.0066 | 0.0072 | 0.92 | 0.362 | 3.46 | 0.052 | 20 | 0.0031 | 0.0057 | 0.54 | 0.592 |
| THREESPINE | 0 | 12 | 0.0060 | 0.0073 | 0.82 | 0.427 |  |  | 12 | 0.0060 | 0.0073 | 0.82 | 0.427 |
| STICKLEBACK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL SPINY | 0 | 24 | 0.0233 | 0.0198 | 1.18 | 0.250 | 0.29 | 0.796 | 2 | 0.0041 | 0.0116 | 0.35 | 0.761 |
| LUMPSUCKER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| DUSKY SEASNAIL | 0 | 40 | 0.0011 | 0.0129 | 0.09 | 0.931 |  |  |  |  |  |  |  |
| NORTHERN SAND | 0 | 10 | -0.0042 | 0.0364 | -0.12 | 0.910 | 4.46 | 0.019 |  |  |  |  |  |
| LANCE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FISH DOCTOR | 0 | 14 | 0.0996 | 0.0750 | 1.33 | 0.205 |  |  |  |  |  |  |  |
| LAVAL`S EELPOUT | 0 | 47 | 0.0115 | 0.0090 | 1.29 | 0.204 |  |  |  |  |  |  |  |
| SNAKEBLENNY | 0 | 19 | 0.0048 | 0.0099 | 0.49 | 0.631 | 1.22 | 0.324 | 16 | 0.0097 | 0.0072 | 1.34 | 0.199 |
| DAUBED SHANNY | 0 | 68 | -0.0002 | 0.0048 | -0.04 | 0.972 | 6.32 | 0.007 |  |  |  |  |  |
| 4-LINE SNAKE | 0 | 26 | -0.0154 | 0.0122 | -1.26 | 0.219 |  |  |  |  |  |  |  |
| BLENNY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| STOUT EELBLENNY | 0 | 29 | 0.0074 | 0.0132 | 0.56 | 0.581 |  |  |  |  |  |  |  |
| OCEAN | 0 | 7 |  |  |  |  | 4.58 | 0.033 |  |  |  |  |  |

|  | 1. September survey data only |  |  |  |  |  | 2. Survey effect |  | 3. September \& July data where appropriate |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | outlier | DF | $\beta_{\text {depth }}$ | SE | $t$ | P | F | $\boldsymbol{P}$ | DF | $\beta_{\text {depth }}$ | SE | t | $\boldsymbol{P}$ |
| POUT(COMMON) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| VAHL'S EELPOUT | 0 | 16 | 0.0004 | 0.0080 | 0.05 | 0.957 | 0.16 | 0.854 | 24 | -0.0013 | 0.0050 | -0.27 | 0.793 |
| ATL HOOKEAR | 0 | 35 | 0.0024 | 0.0080 | 0.31 | 0.761 | 0.67 | 0.653 | 2 | 0.0013 | 0.0068 | 0.19 | 0.865 |
| SCULPIN |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ATL ROCK CRAB | 0 | 15 | -0.0060 | 0.0151 | -0.40 | 0.697 | 1.39 | 0.348 | 5 | -0.0143 | 0.0136 | -1.05 | 0.342 |
| Hyas coarctatus | 0 | 72 | -0.0039 | 0.0049 | -0.78 | 0.436 | 6.16 | 0.006 |  |  |  |  |  |
| Hyas coarctatus | 1 | 71 | -0.0034 | 0.0046 | -0.74 | 0.461 | 2.84 | 0.075 | 30 | -0.0087 | 0.0045 | -1.94 | 0.062 |
| SNOW CRAB | 0 | 90 | 0.0027 | 0.0020 | 1.33 | 0.185 | 4.80 | 0.010 |  |  |  |  |  |
| SNOW CRAB | 1 | 90 |  |  |  |  | 7.16 | 0.001 |  |  |  |  |  |
| Hyas araneus | 0 | 47 | -0.0252 | 0.0162 | -1.56 | 0.126 | 3.49 | 0.113 | 6 | 0.0004 | 0.0094 | 0.04 | 0.967 |
| SHORT-FIN SQUID | 0 | 23 | -0.0077 | 0.0033 | -2.36 | 0.027 | 0.01 | 0.989 | 61 | -0.0001 | 0.0026 | -0.02 | 0.980 |
| SHORT-FIN SQUID | 1 | 23 |  |  |  |  | 0.00 | 0.996 | 60 | -0.0001 | 0.0026 | -0.02 | 0.982 |

Table 14. Results of preliminary fixed effects model analyses, with set pairs having $\geq 20 \%$ difference in tow distance removed, testing for a diel difference in the relative catchability of the CCGS Alfred Needler and the CCGS Teleost based on (1) the September 2004 and 2005 comparative fishing experiments and (2) the combined September and July comparative fishing experiments. Probability values are based on 999 permutations under the null hypothesis.

| Species | September |  |  |  | September and July |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $N$ | $\beta_{\text {diel }}$ | S.E. | $P$ | $N$ | $\beta_{\text {diel }}$ | S.E. | $P$ |
| Cod | 81 | 0.757 | 0.196 | 0.544 | 185 | 0.481 | 0.125 | 0.475 |
| outlier removed | 80 | 0.222 | 0.136 | 0.133 | 184 | 0.098 | 0.105 | 0.634 |
| White hake | 21 | -0.029 | 0.192 | 0.892 | 35 | -0.144 | 0.128 | 0.477 |
| Redfish | 24 | -0.372 | 0.379 | 0.370 | 141 | -0.100 | 0.135 | 0.623 |
| Halibut | 11 | -0.647 | 3.754 | 0.514 | 50 | 0.361 | 0.409 | 0.484 |
| Greenland halibut | 38 | 0.331 | 0.473 | 0.620 | 70 | 0.344 | 0.151 | 0.289 |
| American plaice | 90 | 0.116 | 0.102 | 0.281 | 228 | -0.048 | 0.062 | 0.611 |
| Witch flounder | 24 | -0.193 | 0.668 | 0.692 | 120 | 0.064 | 0.136 | 0.821 |
| Yellowtail flounder | 37 | -0.080 | 0.355 | 0.827 | 108 | -0.059 | 0.113 | 0.814 |
| Winter flounder | 19 | 0.435 | 0.630 | 0.556 | 49 | 0.333 | 0.200 | 0.547 |
| Herring | 56 | 0.557 | 0.360 | 0.545 | 155 | 0.679 | 0.207 | 0.302 |
| outlier removed | 55 | 0.772 | 0.677 | 0.280 | 154 | 0.854 | 0.176 | 0.137 |
| Capelin | 59 | -0.348 | 0.225 | 0.742 | 87 | -0.175 | 0.179 | 0.820 |
| outlier removed | 58 | 0.240 | 0.318 | 0.469 | 86 | 0.249 | 0.150 | 0.359 |
| Mackerel | 18 | -2.124 | 4.896 | 0.299 | 24 | -1.348 | 0.992 | 0.478 |
| Longfin hake |  |  |  |  | 28 | -1.363 | 0.592 | 0.059 |
| Fourbeard rockling | 18 | 1.322 | 0.746 | 0.337 | 42 | 0.465 | 0.489 | 0.546 |
| Greenland cod | 26 | 1.560 | 0.808 | 0.061 |  |  |  |  |
| Thorny skate | 22 | -0.359 | 0.374 | 0.313 | 98 | 0.141 | 0.200 | 0.590 |
| Smooth skate | 11 | -0.241 | 0.443 | 0.562 | 47 | 0.034 | 0.406 | 0.952 |
| Winter skate |  |  |  |  | 17 | -0.513 | 1.058 | 0.708 |
| Spiny dogfish |  |  |  |  | 31 | -0.199 | 0.149 | 0.426 |
| Atlantic hagfish | 12 | -0.236 | 0.852 | 0.837 | 41 | 0.127 | 0.492 | 0.838 |
| Longhorn sculpin | 19 | 0.548 | 0.679 | 0.743 | 97 | -0.171 | 0.250 | 0.659 |
| Shorthorn sculpin | 31 | -0.283 | 0.653 | 0.765 | 38 | -0.240 | 0.514 | 0.796 |
| Arctic staghorn sculpin | 32 | -1.293 | 1.222 | 0.769 |  |  |  |  |
| Moustache sculpin | 50 | 1.394 | 0.480 | 0.173 | 101 | 1.236 | 0.457 | 0.058 |
| Arctic hookear sculpin | 21 | -0.169 | 4.024 | 0.956 |  |  |  |  |
| Atl hookear sculpin | 33 | 0.141 | 2.059 | 0.905 |  |  |  |  |
| Hookear sculpins, ns | 39 | 0.071 | 1.034 | 0.965 | 68 | 0.427 | 0.622 | 0.766 |
| Spatulate sculpin | 39 | -0.329 | 0.521 | 0.574 |  |  |  |  |
| Sea raven | 13 | -0.864 | 0.817 | 0.437 | 71 | -1.001 | 0.210 | 0.044 |
| outlier removed |  |  | s above |  | 70 | -0.759 | 0.256 | 0.026 |
| Alligatorfish | 39 | -0.813 | 0.393 | 0.078 | 91 | -0.251 | 0.229 | 0.439 |
| Arctic alligatorfish | 27 | 0.445 | 0.877 | 0.773 |  |  |  |  |


| Species | September |  |  |  | September and July |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $N$ | $\beta_{\text {diel }}$ | S.E. | $P$ | $N$ | $\beta_{\text {diel }}$ | S.E. | $P$ |
| Alligatorfishes, n.s. | 53 | -0.202 | 0.494 | 0.820 |  |  |  |  |
| Atl sea poacher | 37 | 0.434 | 0.329 | 0.160 | 59 | 0.389 | 0.182 | 0.120 |
| M.-s. grenadier | 8 | 0.932 | 7.138 | 0.302 | 23 | 0.678 | 0.297 | 0.155 |
| Spiny lumpsucker | 24 | -0.384 | 2.572 | 0.768 | 49 | -0.418 | 0.574 | 0.661 |
| Dusky seasnail | 39 | -0.525 | 0.471 | 0.592 | 48 | -0.456 | 0.307 | 0.620 |
| outlier removed | 38 | -1.107 | 0.563 | 0.054 | 47 | -1.026 | 0.417 | 0.055 |
| Sand lance | 11 | -1.535 | 2.049 | 0.743 | 53 | -0.197 | 0.065 | 0.631 |
| Laval's eelpout | 44 | -0.067 | 0.439 | 0.911 |  |  |  |  |
| Snakeblenny | 21 | 2.141 | 0.708 | 0.016 | 44 | 0.202 | 0.245 | 0.632 |
| Daubed shanny | 66 | 0.228 | 0.240 | 0.401 | 95 | 0.041 | 0.047 | 0.854 |
| 4-line snake blenny | 25 | -0.730 | 0.785 | 0.467 |  |  |  |  |
| Stout eelblenny | 29 | 0.064 | 0.252 | 0.770 |  |  |  |  |
| Ocean pout | 9 | 25.585 | 8.768 | 0.942 | 37 | -0.094 | 0.333 | 0.886 |
| Shanny-n.s. | 66 | 0.006 | 0.120 | 0.974 |  |  |  |  |
| outliers removed | 63 | -0.140 | 0.130 | 0.271 |  |  |  |  |
| Vahl's eelpout | 18 | 0.372 | 0.480 | 0.862 | 46 | 0.289 | 0.164 | 0.532 |
| outlier removed | 17 | -0.591 | 0.373 | 0.085 | 45 | -0.069 | 0.271 | 0.847 |
| White barracudina | 7 | 0.153 | 5.747 | 0.692 | 19 | 1.563 | 0.217 | 0.195 |
| Atlantic rock crab | 16 | -0.196 | 0.456 | 0.636 | 37 | -0.703 | 0.257 | 0.496 |
| outlier removed | 15 | -0.093 | 0.513 | 0.867 | 36 | -0.209 | 0.386 | 0.651 |
| Hyas coarctatus | 68 | 0.167 | 0.271 | 0.641 | 125 | 0.277 | 0.083 | 0.339 |
| outliers removed | 67 | 0.318 | 0.323 | 0.356 | 122 | 0.370 | 0.170 | 0.117 |
| Northern stone crab | 11 | 1.249 | 8.369 | 0.567 | 41 | 0.396 | 0.383 | 0.435 |
| Snow crab | 85 | 0.230 | 0.269 | 0.413 | 172 | -0.020 | 0.034 | 0.929 |
| outlier removed |  |  | s above |  | 171 | 0.080 | 0.116 | 0.663 |
| Hyas araneus | 44 | -0.115 | 1.146 | 0.870 | 85 | 0.113 | 0.190 | 0.904 |
| outliers removed |  |  | s above |  | 83 | 0.476 | 0.361 | 0.454 |
| Short-fin squid | 24 | -0.655 | 1.290 | 0.422 | 129 | 0.229 | 0.081 | 0.548 |
| outlier removed |  |  | s above |  | 128 | 0.389 | 0.178 | 0.280 |

Table 15. Results of (1.) fixed and (2.) mixed effects model analyses testing for a difference in the catchability of certain invertebrate taxa between the CCGS Alfred Needler and the CCGS Teleost. Analyses are based on the September 2004 and 2005 comparative fishing experiments, with set pairs having $\geq 20 \%$ difference in tow distance removed. The column outlier indicates whether the analysis is based on all available data (0), removing outliers only (1), removing only sets in which much unsorted remaining catch was reported (2) removing outliers and sets with unsorted catch (3). Probability values are based on 999 permutations under the null hypothesis in (1) and represent the probability value for the $t$-statistic (2.).

| species | outlier | DF | 1. Fixed-effects model analysis |  |  |  | 2. Mixed-effects model analysis |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\beta_{v}$ | SE | $F$ | $P_{\text {rand }}$ | $\beta_{v}$ | SE | $F$ | P |
| Whelk eggs | 0 | 53 | -0.313 | 0.426 | 0.45 | 0.512 | -0.313 | 0.426 | 0.54 | 0.466 |
|  |  | 52 | -0.426 | 0.419 | 1.18 | 0.293 | -0.426 | 0.419 | 1.03 | 0.314 |
|  | 2 | 46 | -0.148 | 0.439 | 0.10 | 0.766 | -0.148 | 0.439 | 0.11 | 0.738 |
|  | 3 | 45 | -0.275 | 0.430 | 0.55 | 0.483 | -0.275 | 0.430 | 0.41 | 0.526 |
| Sea potato | 0 | 64 | -0.602 | 0.437 | 0.63 | 0.438 | -0.602 | 0.437 | 1.90 | 0.173 |
|  | 1 | 62 | -0.773 | 0.411 | 2.85 | 0.104 | -0.773 | 0.411 | 3.53 | 0.065 |
|  | 2 | 48 | -0.213 | 0.483 | 0.00 | 0.997 | -0.213 | 0.483 | 0.19 | 0.662 |
|  | 3 | 46 | -0.425 | 0.443 | 0.66 | 0.421 | -0.425 | 0.443 | 0.92 | 0.342 |
| Shrimp | 0 | 88 | -0.806 | 0.253 | 0.67 | 0.414 | -0.806 | 0.253 | 10.15 | 0.002 |
|  |  | 86 | -0.827 | 0.258 | 0.87 | 0.332 | -0.827 | 0.258 | 10.31 | 0.002 |
|  | 2 | 78 | -0.292 | 0.219 | 0.83 | 0.372 | -0.292 | 0.219 | 1.77 | 0.187 |
|  | 3 | 76 | -0.302 | 0.224 | 0.75 | 0.408 | -0.302 | 0.224 | 1.83 | 0.181 |
| Hermit crabs | 0 | 67 | -0.459 | 0.326 | 2.51 | 0.103 | -0.459 | 0.326 | 1.98 | 0.164 |
|  | 2 | 52 | -0.294 | 0.287 | 1.62 | 0.222 | -0.294 | 0.287 | 1.05 | 0.310 |
| Polycheates | 0 | 51 | -0.920 | 0.175 | 35.70 | <0.001 | -0.920 | 0.175 | 27.72 | <0.001 |
|  | 1 | 49 | -1.055 | 0.153 | 52.65 | <0.001 | -1.055 | 0.153 | 47.58 | <0.001 |
|  | 2 | 30 | -0.655 | 0.252 | 8.04 | 0.010 | -0.655 | 0.252 | 6.73 | 0.015 |
|  | 3 | 28 | -0.868 | 0.218 | 14.73 | <0.001 | -0.868 | 0.218 | 15.89 | 0.000 |
| Sea mouse | 0 | 12 | -0.329 | 0.922 | 0.07 | 0.788 | -0.329 | 0.922 | 0.13 | 0.728 |
|  | 2 | 10 | -0.350 | 0.974 | 0.07 | 0.806 | -0.350 | 0.974 | 0.13 | 0.727 |
| Mollusk shells | 0 | 86 | -0.606 | 0.205 | 9.88 | 0.003 | -0.606 | 0.205 | 8.73 | 0.004 |
|  | 1 | 85 | -0.631 | 0.206 | 11.38 | 0.003 | -0.631 | 0.206 | 9.36 | 0.003 |
|  | 2 | 79 | -0.590 | 0.171 | 13.10 | 0.003 | -0.590 | 0.171 | 11.91 | 0.001 |
|  | 3 | 78 | -0.616 | 0.171 | 15.56 | <0.001 | -0.616 | 0.171 | 12.99 | 0.001 |
| Whelks | 0 | 69 | -0.343 | 0.256 | 1.90 | 0.167 | -0.343 | 0.256 | 1.79 | 0.185 |
|  | 1 | 68 | -0.373 | 0.258 | 2.47 | 0.130 | -0.373 | 0.258 | 2.10 | 0.152 |
|  | 2 | 63 | -0.176 | 0.226 | 0.68 | 0.408 | -0.176 | 0.226 | 0.61 | 0.439 |
|  | 3 | 62 | -0.207 | 0.227 | 1.10 | 0.286 | -0.207 | 0.227 | 0.83 | 0.367 |
| Iceland scallop | 0 | 19 | -0.906 | 0.756 | 1.58 | 0.243 | -0.906 | 0.756 | 1.44 | 0.246 |
|  | 1 | 18 | -1.249 | 0.711 | 3.50 | 0.085 | -1.249 | 0.711 | 3.09 | 0.096 |
|  | 2 | 13 | -1.995 | 0.673 | 10.03 | 0.016 | -1.995 | 0.673 | 8.80 | 0.011 |
| Octopus | 0 | 14 | -0.091 | 0.633 | 0.00 | 0.997 | -0.091 | 0.633 | 0.02 | 0.888 |
|  | 2 | 13 | 0.204 | 0.602 | 0.25 | 0.609 | 0.204 | 0.602 | 0.11 | 0.740 |


|  |  |  | 1. Fixed-effects model analysis |  |  |  | 2. Mixed-effects model analysis |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| species | outlier | DF | $\beta_{v}$ | SE | $F$ | $P_{\text {rand }}$ | $\beta_{v}$ | SE | F | P |
| Starfish | 0 | 70 | 0.124 | 0.371 | 0.14 | 0.703 | 0.124 | 0.371 | 0.11 | 0.740 |
|  | 1 | 67 | 0.141 | 0.345 | 0.22 | 0.646 | 0.141 | 0.345 | 0.17 | 0.684 |
|  | 2 | 56 | 0.204 | 0.366 | 0.28 | 0.605 | 0.204 | 0.366 | 0.31 | 0.580 |
|  |  | 54 | 0.387 | 0.347 | 1.60 | 0.229 | 0.387 | 0.347 | 1.24 | 0.270 |
| Mud star | 0 | 29 | -0.254 | 0.507 | 0.30 | 0.549 | -0.254 | 0.507 | 0.25 | 0.620 |
|  | 2 | 23 | 0.317 | 0.558 | 0.32 | 0.598 | 0.317 | 0.558 | 0.32 | 0.576 |
| Sunstars | 0 | 79 | -0.465 | 0.243 | 5.29 | 0.026 | -0.465 | 0.243 | 3.65 | 0.060 |
|  | 2 | 74 | -0.206 | 0.203 | 2.40 | 0.131 | -0.206 | 0.203 | 1.03 | 0.314 |
| Brittle stars | 0 | 38 | -2.655 | 0.439 | 22.84 | <0.001 | -2.655 | 0.439 | 36.62 | <0.001 |
|  | 1 | 37 | -2.498 | 0.421 | 22.45 | <0.001 | -2.498 | 0.421 | 35.26 | <0.001 |
|  | 2 | 26 | -2.496 | 0.574 | 10.60 | 0.002 | -2.496 | 0.574 | 18.93 | <0.001 |
|  | 3 | 25 | -2.261 | 0.544 | 9.40 | 0.005 | -2.261 | 0.544 | 17.28 | <0.001 |
| Basket stars | 0 | 67 | -0.487 | 0.369 | 3.10 | 0.098 | -0.487 | 0.369 | 1.74 | 0.192 |
|  | 2 | 61 | -0.571 | 0.340 | 3.95 | 0.054 | -0.571 | 0.340 | 2.81 | 0.099 |
| Sea urchins | 0 | 83 | -0.248 | 0.217 | 0.38 | 0.542 | -0.248 | 0.217 | 1.30 | 0.257 |
|  | 2 | 80 | -0.206 | 0.203 | 0.24 | 0.639 | -0.206 | 0.203 | 1.02 | 0.315 |
| Sand dollars | 0 | 48 | -0.155 | 0.344 | 1.18 | 0.279 | -0.155 | 0.344 | 0.20 | 0.655 |
|  | 1 | 47 | -0.102 | 0.347 | 0.58 | 0.445 | -0.102 | 0.347 | 0.09 | 0.769 |
|  | 2 | 44 | -0.018 | 0.342 | 0.62 | 0.450 | -0.018 | 0.342 | 0.00 | 0.958 |
|  | 3 | 43 | 0.042 | 0.344 | 0.18 | 0.676 | 0.042 | 0.344 | 0.01 | 0.904 |
| Sea | 0 | 25 | -0.477 | 0.794 | 0.81 | 0.394 | -0.477 | 0.794 | 0.36 | 0.553 |
|  | 1 | 24 | -0.521 | 0.826 | 1.05 | 0.308 | -0.521 | 0.826 | 0.40 | 0.534 |
|  | 2 | 19 | -0.385 | 0.715 | 0.68 | 0.370 | -0.385 | 0.715 | 0.29 | 0.596 |
|  | 3 | 18 | -0.437 | 0.751 | 1.01 | 0.347 | -0.437 | 0.751 | 0.34 | 0.568 |
| Sea anemones | 0 | 75 | 0.081 | 0.338 | 0.04 | 0.836 | 0.081 | 0.338 | 0.06 | 0.811 |
|  | 1 | 74 | -0.029 | 0.324 | 0.05 | 0.806 | -0.029 | 0.324 | 0.01 | 0.929 |
|  | 2 | 61 | 0.215 | 0.317 | 0.39 | 0.523 | 0.215 | 0.317 | 0.46 | 0.499 |
| Sea pen | 0 | 11 | -0.783 | 0.848 | 2.25 | 0.145 | -0.783 | 0.848 | 0.85 | 0.375 |
| Sea cauliflower | 0 | 50 | -0.983 | 0.334 | 9.52 | 0.004 | -0.983 | 0.334 | 8.65 | 0.005 |
|  | 2 | 40 | -0.706 | 0.345 | 4.88 | 0.043 | -0.706 | 0.345 | 4.18 | 0.048 |
| Jellyfish | 0 | 25 | 1.348 | 1.037 | 1.52 | 0.228 | 1.348 | 1.037 | 1.69 | 0.206 |
|  | 2 | 18 | 1.399 | 1.117 | 1.18 | 0.308 | 1.399 | 1.117 | 1.57 | 0.226 |
| Sponges | 0 | 66 | -0.939 | 0.504 | 3.62 | 0.066 | -0.939 | 0.504 | 3.47 | 0.067 |
|  | 2 | 56 | -0.814 | 0.528 | 2.57 | 0.121 | -0.814 | 0.528 | 2.38 | 0.129 |
| Plants \& algae | 0 | 82 | -1.119 | 0.358 | 6.46 | 0.017 | -1.119 | 0.358 | 9.77 | 0.002 |
|  |  | 81 | -1.152 | 0.361 | 7.82 | 0.009 | -1.152 | 0.361 | 10.18 | 0.002 |
|  | 2 | 71 | -0.799 | 0.354 | 2.88 | 0.094 | -0.799 | 0.354 | 5.09 | 0.027 |
|  | 3 | 70 | -0.832 | 0.357 | 3.84 | 0.063 | -0.832 | 0.357 | 5.42 | 0.023 |
| Wood debris | 0 | 67 | -1.082 | 0.420 | 7.77 | 0.002 | -1.082 | 0.420 | 6.62 | 0.012 |
|  | 1 | 64 | -1.200 | 0.430 | 9.60 | 0.005 | -1.200 | 0.430 | 7.77 | 0.007 |
|  | 2 | 49 | -1.010 | 0.426 | 6.45 | 0.011 | -1.010 | 0.426 | 5.61 | 0.022 |
|  | 3 | 46 | -1.168 | 0.436 | 8.96 | 0.002 | -1.168 | 0.436 | 7.17 | 0.010 |

Table 16. Locations of stations sampled non-simultaneously by both the Alfred Needler $(N)$ and Teleost ( $T$ ) in 2004. Because these stations were not fished synchronously by both vessels, they were not included as part of the comparative fishing experiments but nonetheless constitute repeat sets for the regular survey sampling and therefore need to be weighted accordingly (see text section 4.2)

| Stratum | $N$ set \# | T set \# | latitude | longitude |
| :---: | :---: | :---: | :---: | :---: |
| 432 | 5 | 5 | 45.5239 | 62.3815 |
| 402 | 3 | 6 | 45.5602 | 63.2006 |
| 402 | 2 | 7 | 46.0605 | 63.2385 |
| 432 | 4 | 10 | 45.5622 | 63.0356 |
| 433 | 6 | 12 | 45.5415 | 62.2456 |
| 403 | 7 | 16 | 45.4971 | 61.5001 |
| 403 | 9 | 18 | 45.4806 | 61.444 |
| 433 | 10 | 19 | 46.0501 | 61.4851 |
| 433 | 12 | 20 | 46.1289 | 61.5014 |
| 433 | 13 | 22 | 46.2265 | 61.5417 |
| 434 | 14 | 23 | 46.2572 | 61.4797 |
| 429 | 19 | 113 | 46.5947 | 62.4232 |
| 423 | 21 | 117 | 47.1639 | 62.2689 |
| 417 | 106 | 124 | 48.1002 | 63.5647 |
| 418 | 113 | 126 | 47.5509 | 64.539 |
| 419 | 110 | 127 | 47.535 | 65.2347 |
| 419 | 111 | 129 | 47.578 | 65.4563 |
| 418 | 108 | 132 | 48.0615 | 64.4768 |
| 417 | 105 | 134 | 48.2443 | 64.0522 |
| 416 | 104 | 136 | 48.2721 | 63.5269 |
| 416 | 103 | 137 | 48.381 | 63.5043 |
| 415 | 99 | 139 | 48.5268 | 63.4588 |
| 415 | 98 | 141 | 48.5089 | 63.3769 |
| 416 | 97 | 146 | 48.2781 | 63.3682 |
| 416 | 96 | 147 | 48.2247 | 63.4215 |
| 424 | 95 | 154 | 47.5629 | 63.0693 |
| 431 | 20 | 175 | 47.0413 | 62.2176 |
| 431 | 18 | 179 | 46.4609 | 62.2098 |
| 431 | 15 | 181 | 46.3335 | 62.1497 |
| 431 | 17 | 182 | 46.4184 | 62.1081 |

a)

b)


Figure 1. (a) Stratum boundaries for the southern Gulf of St. Lawrence September bottom-trawl survey, and (b) location of fishing sets in the 2004 (+) and 2005 (o) comparative fishing experiments.


Figure 2. Histogram of the percentage difference in towing distance within relevant set pairs in the September and July comparative fishing experiments.


Figure 3. The following general caption, describing the contents of each panel, applies to the individual species figures that follow.
Data:
a) Total standardized catches from paired sets by the Teleost vs. Alfred Needler from the September (o) and July ( + ) experiments. Dashed line is the 1:1 relationship.
b) Relative total length frequencies for Needler (solid line) \& Teleost (dashed) in September.
c) Relative total length frequencies in the July experiments.

Length-aggregated analysis:
d) Histogram of standardized $\chi^{2}$ residuals from the fixed-effects analysis.
e) Histogram of standardized $\chi^{2}$ residuals from the mixed-effects analysis.
f) Histogram of random effects from the mixed-effects analysis.

Analysis of length-dependent relative catchability:
g) General additive model (GAM; with $95 \%$ confidence intervals) fit of standardized conditional $\chi^{2}$ residuals from the fixed-effects analysis versus length.
h) Standardized total conditional $\chi^{2}$ residuals from each set in the fixed-effects analysis.
i) Like panel (g), but for the mixed-effects analysis.
j) Like panel (h), but for the mixed-effects analysis.
k) Predicted random effects versus length for each set.

Analysis of depth-dependent relative catchability:
I) Standardized conditional $\chi^{2}$ residuals from the fixed-effects analysis (+) along with GAM fit with $95 \%$ confidence intervals (lines) versus depth.
m) Like panel (I), but for the mixed-effects analysis.
n) Histogram of random effects from the mixed-effects analysis.


Figure 4. Comparative fishing analysis results for Atlantic cod (see Fig. 3 for details on the panel contents).


Figure 5. Comparative fishing analysis results for white hake (see Fig. 3 for details on the panel contents).


Figure 6. Comparative fishing analysis results for redfish (see Fig. 3 for details on the panel contents).


Figure 7. Comparative fishing analysis results for Atlantic halibut (see Fig. 3 for details on the panel contents).


Figure 8. Comparative fishing analysis results for Greenland halibut (see Fig. 3 for details on the panel contents).


Figure 9. Comparative fishing analysis results for American plaice (see Fig. 3 for details on the panel contents).


Figure 10. Comparative fishing analysis results for witch flounder (see Fig. 3 for details on the panel contents).


Figure 11. Comparative fishing analysis results for yellowtail flounder (see Fig. 3 for details on the panel contents).


Figure 12. Comparative fishing analysis results for winter flounder (see Fig. 3 for details on the panel contents).


Figure 13. Comparative fishing analysis results for herring (see Fig. 3 for details on the panel contents).


Figure 14. Comparative fishing analysis results for gaspereau (see Fig. 3 for details on the panel contents).


Figure 15. Comparative fishing analysis results for capelin (see Fig. 3 for details on the panel contents).


Figure 16. Comparative fishing analysis results for mackerel (see Fig. 3 for details on the panel contents).


Figure 17. Comparative fishing analysis results for fourbeard rockling (see Fig. 3 for details on the panel contents).




Figure 18. Comparative fishing analysis results for Greenland cod (see Fig. 3 for details on the panel contents).


Figure 19. Comparative fishing analysis results for thorny skate (see Fig. 3 for details on the panel contents).


Figure 20. Comparative fishing analysis results for smooth skate (see Fig. 3 for details on the panel contents).


Figure 21. Comparative fishing analysis results for winter skate (see Fig. 3 for details on the panel contents).


Figure 22. Comparative fishing analysis results for hagfish (see Fig. 3 for details on the panel contents).


Figure 23. Comparative fishing analysis results for longhorn sculpin (see Fig. 3 for details on the panel contents).


Figure 24. Comparative fishing analysis results for shorthorn sculpin (see Fig. 3 for details on the panel contents).


Figure 25. Comparative fishing analysis results for arctic staghorn sculpin (see Fig. 3 for details on the panel contents).


Figure 26. Comparative fishing analysis results for mailed sculpin (see Fig. 3 for details on the panel contents).


Figure 27. Comparative fishing analysis results for Arctic hookear sculpin (see Fig. 3 for details on the panel contents).


Figure 28. Comparative fishing analysis results for Atlantic hookear sculpin (see Fig. 3 for details on the panel contents).


Figure 29. Comparative fishing analysis results for spatulate sculpin (see Fig. 3 for details on the panel contents).


Figure 30. Comparative fishing analysis results for sea raven (see Fig. 3 for details on the panel contents).


Figure 31. Comparative fishing analysis results for alligatorfish (see Fig. 3 for details on the panel contents).


Figure 32. Comparative fishing analysis results for sea poacher (see Fig. 3 for details on the panel contents).


Figure 33. Comparative fishing analysis results for spiny lumpsucker (see Fig. 3 for details on the panel contents).


Figure 34. Comparative fishing analysis results for dusky seasnail (see Fig. 3 for details on the panel contents).


Figure 35. Comparative fishing analysis results for sandlance (see Fig. 3 for details on the panel contents).


Figure 36. Comparative fishing analysis results for fish doctor (see Fig. 3 for details on the panel contents).


Figure 37. Comparative fishing analysis results for Laval's eelpout (see Fig. 3 for details on the panel contents).


Figure 38. Comparative fishing analysis results for snakeblenny (see Fig. 3 for details on the panel contents).


Figure 39. Comparative fishing analysis results for daubed shanny (see Fig. 3 for details on the panel contents).


Figure 40. Comparative fishing analysis results for fourline snakeblenny (see Fig. 3 for details on the panel contents).


Figure 41. Comparative fishing analysis results for stout eelblenny (see Fig. 3 for details on the panel contents).


Figure 42. Comparative fishing analysis results for ocean pout (see Fig. 3 for details on the panel contents).


Figure 43. Comparative fishing analysis results for Vahl's eelpout (see Fig. 3 for details on the panel contents).


Figure 44. Comparative fishing analysis results for rock crab (see Fig. 3 for details on the panel contents).


Figure 45. Comparative fishing analysis results for Hyas coarctatus (see Fig. 3 for details on the panel contents).


Figure 46. Comparative fishing analysis results for northern stone crab (see Fig. 3 for details on the panel contents).


Figure 47. Comparative fishing analysis results for snow crab (see Fig. 3 for details on the panel contents).


Figure 48. Comparative fishing analysis results for Hyas araneus (see Fig. 3 for details on the panel contents).


Figure 49. Comparative fishing analysis results for shortfin squid (see Fig. 3 for details on the panel contents).


Figure 50. Total standardized catches of various invertebrate species from paired sets by the Teleost vs. Alfred Needler from the September experiments.


Figure 51. Total standardized catches of various invertebrate species and other biological material from paired sets by the Teleost vs. Alfred Needler from the September experiments.


Figure 52. Comparison of fixed-effects analysis results ( $\pm$ SE) when all relevant set pairs are included and excluding sets with a difference in vessel distance towed $>20 \%$.


Figure 53. Comparison of fixed-effects and mixed-effects analysis results ( $\pm$ SE; excluding sets with a difference in distance towed between vessels $>20 \%$ ).


Figure 54. Total standardized catches-at-length by the Teleost, Alfred Needler (uncorrected), Alfred Needler corrected by $\beta_{v}$ and Alfred Needler corrected by $\beta_{l}$ for those fishes for which a significant difference in catchability between vessels was found. Data from outlier sets (see Appendix II) have been removed.

Appendix I. Numerical codes, taxonomic names and common names of taxa covered by the 2004-2005 comparative fishing experiment analyses.

| code | Species name | Common name |
| :---: | :---: | :---: |
| 10 | Gadus morhua | Atlantic Cod |
| 12 | Urophycis tenuis | White hake |
| 23 | Sebastes sp. | Redfish |
| 30 | Hippoglossus hippoglossus | Atlantic halibut |
| 31 | Reinhardtius hippoglossoides | Greenland halibut |
| 40 | Hippoglossoides platessoides | American plaice |
| 41 | Glyptocephalus cynoglossus | Witch flounder |
| 42 | Limanda ferruginea | Yellowtail flounder |
| 43 | Pseudopleuronectes americanus | Winter flounder |
| 50 | Anarhichas lupus | Striped Atlantic wolffish |
| 60 | Clupea harengus | Atlantic Herring |
| 62 | Alosa pseudoharengus | Gaspereau |
| 64 | Mallotus villosus | Capelin |
| 70 | Scomber scombrus | Atlantic Mackerel |
| 112 | Phycis chesteri | Long fin hake |
| 114 | Enchelyopus cimbrius | Fourbeard rockling |
| 118 | Gadus ogac | Greenland cod |
| 201 | Amblyraja radiata | Thorny skate |
| 202 | Malacoraja senta | Smooth skate |
| 204 | Leucoraja ocellata | Winter skate |
| 220 | Squalus acanthias | Spiny dogfish |
| 241 | Myxine glutinosa | Atlantic hagfish |
| 300 | Myoxocephalus octodecemspinosus | Longhorn sculpin |
| 301 | Myoxocephalus scorpius | Shorthorn sculpin |
| 302 | Gymnocanthus tricuspis | Arctic staghorn sculpin |
| 304 | Triglops murrayi | Moustache (mailed) sculpin |
| 306 | Artediellus uncinatus | Arctic hookear sculpin |
| 314 | Icelus spatula | Spatulate sculpin |
| 320 | Hemitripterus americanus | Sea raven |
| 340 | Aspidophoroides monopterygius | Alligatorfish |
| 341 | Uleina olrikii | Arctic alligatorfish |
| 350 | Leptagonus decagonus | Atlantic sea poacher |
| 361 | Gasterosteus aculeatus aculeatus | Threespine stickleback |
| 400 | Lophius americanus | Monkfish,goosefish,angler |
| 410 | Nezumia bairdii | Marlin-spike grenadier |
| 501 | Cyclopterus lumpus | Lumpfish |
| 502 | Eumicrotremus spinosus | Atlantic spiny lumpsucker |
| 512 | Liparis gibbus | Dusky seasnail, |
| 610 | Ammodytes dubius | Northern sand lance |
| 616 | Gymnelis viridis | Fish doctor |
| 620 | Lycodes lavalaei | Laval`s eelpout |
| 622 | Lumpenus lumpretaeformis | Snakeblenny |
| 623 | Leptoclinus maculatus | Daubed shanny |
| 626 | Eumesogrammus praecisus | Fourline snake blenny |
| 632 | Lumpenus medius | Stout eelblenny |

| code | Species name | Common name |
| ---: | :--- | :--- |
| 640 | Zoarces americanus | Ocean pout |
| 646 | Melanostigma atlanticum | Atlantic soft pout |
| 647 | Lycodes vahlii | Checker eelpout (Vahl's) |
| 712 | Notolepis rissoi kroyeri | White barracudina |
| 880 | Artediellus atlanticus | Atlantic Hookear sculpin |
| 1510 | Buccinidae eggs | Whelk eggs |
| 1823 | Boltenia sp. | Sea potato |
| 2100 | Decapoda Order | Shrimps |
| 2513 | Cancer irroratus | Atlantic rock crab |
| 2521 | Hyas coarctatus | Lyre crab |
| 2523 | Lithodes maja | Northern stone crab |
| 2526 | Chionoecetes opilio | Snow crab |
| 2527 | Hyas araneus | Toad crab |
| 2560 | Paguroidea | Hermit crabs |
| 3100 | Polychaeta | Bristle worms |
| 3212 | Aphrodita sp. | Sea mouse |
| 4000 | Mollusca | Mollusc shells |
| 4210 | Buccinum sp. | Whelks |
| 4322 | Chlamys islandicus | Iceland scallop |
| 4511 | Illex illecebrosus | Short-fin squid |
| 4521 | Octopoda | Octopus |
| 6100 | Asteroidea | Starfish |
| 6115 | Ctenodiscus crispatus | Mud star |
| 6120 | Solaster sp. | Sunstars |
| 6200 | Ophiuroidea | Brittle stars |
| 6300 | Gorgonocephalidae, Asteronychidae | Basket stars |
| 6400 | Strongylocentrotus sp. | Sea urchins |
| 6500 | Clypeasteroida | Sand dollars |
| 6600 | Holothuroidea | Sea cucumbers |
| 8300 | Anthozoa | Sea anemone |
| 8318 | Pennatula borealis | Sea pen |
| 8324 | Gersemia rubiformis | Sea cauliflower |
| 8500 | Scyphozoa | Jellyfishes |
| 8600 | Porifera | Sponges |
| 9300 | Thallophyta | Marine plants \& algae |
|  |  |  |

Appendix II. Sets that were identified as outliers or as having considerable leverage in the analyses without and with covariates. Catches are standardized to a 1.75 nautical mile tow and are in numbers for fish, large crabs and squid, and in kg for the other taxa.

| analysis | code | Species | survey | year | Set | Needler catch | Teleos t catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| length-aggregated | 10 | Cod | 2 | 2005 | 44 | 14.0 | 216.8 |
| length-aggregated | 10 | Cod | 1 | 2005 | 59 | 345.6 | 776.7 |
| length-aggregated | 10 | Cod | 2 | 2005 | 83 | 130.5 | 328.5 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 27 | 274.0 | 30.3 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 35 | 249.0 | 20.6 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 36 | 22.0 | 248.5 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1033 | 32.0 | 265.4 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1034 | 109.0 | 338.7 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1036 | 284.0 | 404.6 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1051 | 33.8 | 278.2 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1056 | 252.0 | 0.0 |
| length-aggregated | 23 | Redfish | 2 | 2005 | 1069 | 272.2 | 307.5 |
| length-aggregated | 40 | American plaice | 2 | 2005 | 36 | 49.0 | 229.8 |
| length-aggregated | 41 | witch flounder | 1 | 2005 | 39 | 104.0 | 8.2 |
| length-aggregated | 43 | winter flounder | 1 | 2005 | 32 | 243.6 | 193.9 |
| length-aggregated | 43 | winter flounder | 1 | 2005 | 96 | 105.3 | 233.3 |
| length-aggregated | 60 | herring | 1 | 2005 | 36 | 265.5 | 30.3 |
| length-aggregated | 60 | herring | 1 | 2005 | 56 | 202.6 | 197.4 |
| length-aggregated | 60 | herring | 1 | 2005 | 57 | 195.4 | 1.1 |
| length-aggregated | 60 | herring | 1 | 2005 | 78 | 404.2 | 382.1 |
| length-aggregated | 60 | herring | 1 | 2005 | 81 | 5.6 | 374.8 |
| length-aggregated | 60 | herring | 1 | 2005 | 112 | 179.0 | 200.3 |
| length-aggregated | 64 | capelin | 2 | 2005 | 14 | 293.0 | 166.9 |
| length-aggregated | 64 | capelin | 2 | 2005 | 16 | 51.0 | 131.5 |
| length-aggregated | 64 | capelin | 2 | 2005 | 17 | 148.0 | 107.5 |
| length-aggregated | 64 | capelin | 2 | 2005 | 24 | 62.3 | 222.0 |
| length-aggregated | 301 | shorthorn sculpin | 1 | 2005 | 32 | 15.9 | 0.0 |
| length-aggregated | 302 | arctic staghorn sculpin | 1 | 2005 | 109 | 39.7 | 75.1 |
| length-aggregated | 320 | sea raven | 2 | 2005 | 44 | 4.0 | 42.8 |
| length-aggregated | 320 | sea raven | 2 | 2005 | 1040 | 37.0 | 3.0 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 36 | 205.0 | 0.0 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 37 | 177.0 | 114.0 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 43 | 200.0 | 0.0 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 67 | 0.0 | 52.0 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 69 | 130.0 | 82.4 |
| length-aggregated | 610 | sand lance | 2 | 2005 | 76 | 126.4 | 91.1 |
| length-aggregated | 647 | Vahl's eelpout | 1 | 2005 | 39 | 59.7 | 2.0 |
| length-aggregated | 2521 | Lyre crabs | 2 | 2005 | 12 | 0.0 | 70.2 |
| length-aggregated | 2521 | Lyre crabs | 2 | 2005 | 16 | 0.0 | 50.6 |
| length-aggregated | 2521 | Lyre crabs | 1 | 2004 | 90 | 58.3 | 0.0 |
| length-aggregated | 2526 | snow crab | 1 | 2004 | 86 | 220.3 | 78.4 |
| length-aggregated | 2526 | snow crab | 1 | 2005 | 89 | 255.7 | 115.7 |


| analysis | code | Species | survey | year | Set | Needler catch | Teleos t catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| length-aggregated | 4511 | short-fin squid | 2 | 2005 | 1005 | 78.0 | 73.7 |
| length-aggregated | 1510 | Whelk eggs | 1 | 2004 | 91 | 3.76 | 0.01 |
| length-aggregated | 1823 | Sea potato | 1 | 2005 | 38 | 52.2 | 0 |
| length-aggregated | 1823 | Sea potato | 1 | 2005 | 49 | 1.7 | 6.4 |
| length-aggregated | 2100 | Shrimps | 1 | 2005 | 101 | 93 | 22.8 |
| length-aggregated | 2100 | Shrimps | 1 | 2005 | 108 | 8.9 | 30.1 |
| length-aggregated | 2526 | Snow crab | 2 | 2005 | 17 | 8 | 288.6 |
| length-aggregated | 3100 | Bristle worms | 1 | 2005 | 100 | 0.0153 | 0 |
| length-aggregated | 3100 | Bristle worms | 1 | 2005 | 114 | 0.015 | 0.001 |
| length-aggregated | 4000 | Mollusc shells | 1 | 2005 | 86 | 3 | 0.7 |
| length-aggregated | 4210 | Whelks | 1 | 2004 | 91 | 1.3 | 0.2 |
| length-aggregated | 4322 | Iceland scallop | 1 | 2005 | 68 | 0.3 | 0 |
| length-aggregated | 6100 | Starfish | 1 | 2005 | 41 | 7.1 | 0 |
| length-aggregated | 6100 | Starfish | 1 | 2005 | 49 | 0.3 | 2 |
| length-aggregated | 6100 | Starfish | 1 | 2005 | 78 | 0 | 2.3 |
| length-aggregated | 6200 | Brittle star | 1 | 2005 | 42 | 0 | 5.5 |
| length-aggregated | 6500 | Sand dollars | 1 | 2005 | 68 | 0.8 | 11.9 |
| length-aggregated | 6600 | Sea cucumbers | 1 | 2005 | 96 | 35.3 | 19.2 |
| length-aggregated | 8300 | Sea anemone | 1 | 2005 | 109 | 4.1 | 0 |
| length-dependent | 10 | Cod | 1 | 2005 | 59 | 345.6 | 776.7 |
| length-dependent | 23 | Redfish | 2 | 2005 | 1056 | 252.0 | 0.0 |
| length-dependent | 23 | Redfish | 2 | 2005 | 1069 | 272.2 | 307.5 |
| length-dependent | 40 | American plaice | 2 | 2005 | 36 | 49.0 | 229.8 |
| length-dependent | 41 | witch flounder | 1 | 2005 | 39 | 104.0 | 8.2 |
| length-dependent | 41 | witch flounder | 1 | 2005 | 154 | 46.7 | 120.6 |
| length-dependent | 42 | yellowtail flounder | 2 | 2005 | 40 | 66.1 | 13.0 |
| length-dependent | 42 | yellowtail flounder | 1 | 2005 | 59 | 422.1 | 335.2 |
| length-dependent | 42 | yellowtail flounder | 1 | 2005 | 60 | 406.0 | 290.8 |
| length-dependent | 42 | yellowtail flounder | 2 | 2005 | 60 | 109.0 | 186.4 |
| length-dependent | 43 | winter flounder | 1 | 2005 | 31 | 137.1 | 18.6 |
| length-dependent | 43 | winter flounder | 1 | 2005 | 32 | 243.6 | 193.9 |
| length-dependent | 43 | winter flounder | 1 | 2005 | 96 | 105.3 | 233.3 |
| length-dependent | 60 | herring | 1 | 2005 | 36 | 265.5 | 30.3 |
| length-dependent | 60 | herring | 1 | 2005 | 56 | 202.6 | 197.4 |
| length-dependent | 60 | herring | 1 | 2005 | 57 | 195.4 | 1.1 |
| length-dependent | 60 | herring | 1 | 2005 | 78 | 404.2 | 382.1 |
| length-dependent | 60 | herring | 1 | 2005 | 81 | 5.6 | 374.8 |
| length-dependent | 64 | capelin | 2 | 2005 | 14 | 293.0 | 166.9 |
| length-dependent | 64 | capelin | 2 | 2005 | 24 | 62.3 | 222.0 |
| length-dependent | 114 | fourbeard rockling | 1 | 2005 | 108 | 15.2 | 29.2 |
| length-dependent | 202 | smooth skate | 2 | 2005 | 22 | 9.0 | 1.0 |
| length-dependent | 202 | smooth skate | 2 | 2005 | 84 | 12.5 | 3.1 |
| length-dependent | 300 | longhorn sculpin | 1 | 2005 | 32 | 37.8 | 4.0 |
| length-dependent | 300 | longhorn sculpin | 1 | 2005 | 96 | 8.9 | 18.6 |
| length-dependent | 301 | shorthorn sculpin | 1 | 2005 | 32 | 15.9 | 0.0 |
| length-dependent | 320 | sea raven | 2 | 2005 | 44 | 4.0 | 42.8 |
| length-dependent | 320 | sea raven | 2 | 2005 | 1040 | 37.0 | 3.0 |
| length-dependent | 620 | Laval's eelpout | 1 | 2005 | 85 | 25.6 | 1.3 |


| analysis | code | Species | survey | year | Set | Needler <br> catch | Teleos <br> t catch |
| :--- | :---: | :--- | :---: | :---: | :---: | :---: | :---: |
| length-dependent | 640 | Ocean pout | 2 | 2005 | 1044 | 2.0 | 19.9 |
| length-dependent | 647 | Vahl's eelpout | 1 | 2005 | 39 | 59.7 | 2.0 |
| length-dependent | 647 | Vahl's eelpout | 1 | 2005 | 154 | 8.8 | 17.8 |
| length-dependent | 2521 | Lyre crab | 1 | 2004 | 90 | 58.3 | 0.0 |
| length-dependent | 2527 | Toad crab | 1 | 2005 | 38 | 13.0 | 4.1 |
| depth-dependent | 10 | Cod | 2 | 2005 | 83 | 130.5 | 328.5 |
| depth-dependent | 23 | Redfish | 2 | 2005 | 1056 | 252.0 | 0.0 |
| depth-dependent | 23 | Redfish | 2 | 2005 | 1069 | 272.2 | 307.5 |
| depth-dependent | 41 | witch flounder | 1 | 2005 | 39 | 104.0 | 8.2 |
| depth-dependent | 43 | winter flounder | 1 | 2005 | 32 | 243.6 | 193.9 |
| depth-dependent | 43 | winter flounder | 1 | 2005 | 96 | 105.3 | 233.3 |
| depth-dependent | 60 | herring | 1 | 2005 | 36 | 265.5 | 30.3 |
| depth-dependent | 60 | herring | 1 | 2005 | 56 | 202.6 | 197.4 |
| depth-dependent | 60 | herring | 1 | 2005 | 78 | 404.2 | 382.1 |
| depth-dependent | 60 | herring | 1 | 2005 | 81 | 5.6 | 374.8 |
| depth-dependent | 62 | gaspereau | 1 | 2005 | 31 | 1.9 | 108.5 |
| depth-dependent | 64 | capelin | 2 | 2005 | 14 | 293.0 | 166.9 |
| depth-dependent | 64 | capelin | 2 | 2005 | 16 | 51.0 | 131.5 |
| depth-dependent | 64 | capelin | 2 | 2005 | 24 | 62.3 | 222.0 |
| depth-dependent | 2521 | Lyre crab | 2 | 2005 | 12 | 0.0 | 70.2 |
| depth-dependent | 2521 | Lyre crab | 2 | 2005 | 16 | 0.0 | 50.6 |
| depth-dependent | 2521 | Lyre crab | 1 | 2004 | 90 | 58.3 | 0.0 |
| depth-dependent | 2526 | snow crab | 2 | 2005 | 17 | 8.0 | 288.6 |
| depth-dependent | 2527 | Toad crab | 2 | 2005 | 12 | 69.0 | 0.0 |

Appendix III. Proportion of total catch in each of the September and July comparative fishing experiments that was occurred in set pairs where the difference in tow distance between vessels was $\geq 20 \%$. Note that such a difference occurred in seven of 101 (7\%) sets in September and four of 173 ( $\sim 2 \%$ ) sets in July.

| code | species | Proportion September | Proportio <br> n July |
| :---: | :---: | :---: | :---: |
| 10 | COD(ATLANTIC) | 0.062 | 0.062 |
| 12 | WHITE HAKE | 0.000 | 0.467 |
| 23 | REDFISH UNSEPARATED | 0.001 | 0.024 |
| 30 | HALIBUT(ATLANTIC) | 0.000 | 0.000 |
| 31 | TURBOT,GREENLAND HALIBUT | 0.001 | 0.112 |
| 40 | AMERICAN PLAICE | 0.086 | 0.019 |
| 41 | WITCH FLOUNDER | 0.000 | 0.020 |
| 42 | YELLOWTAIL FLOUNDER | 0.080 | 0.001 |
| 43 | WINTER FLOUNDER | 0.004 | 0.226 |
| 50 | STRIPED ATL WOLFFISH | 0.000 | 0.018 |
| 60 | HERRING(ATLANTIC) | 0.003 | 0.004 |
| 62 | ALEWIFE | 0.000 | 0.004 |
| 63 | RAINBOW SMELT | 0.000 |  |
| 64 | CAPELIN | 0.025 | 0.020 |
| 70 | MACKEREL(ATLANTIC) | 0.115 | 0.045 |
| 110 | ARCTIC COD | 0.000 |  |
| 112 | LONGFIN HAKE | 0.000 | 0.005 |
| 114 | FOURBEARD ROCKLING | 0.008 | 0.081 |
| 118 | GREENLAND COD | 0.314 |  |
| 122 | CUNNER | 0.000 | 0.000 |
| 143 | BRILL/WINDOWPANE | 0.000 | 0.000 |
| 201 | THORNY SKATE | 0.000 | 0.006 |
| 202 | SMOOTH SKATE | 0.000 | 0.030 |
| 204 | WINTER SKATE | 0.000 | 0.000 |
| 241 | ATLANTIC HAGFISH | 0.000 | 0.010 |
| 300 | LONGHORN SCULPIN | 0.018 | 0.040 |
| 301 | SHORTHORN SCULPIN | 0.017 | 0.000 |
| 302 | ARCTIC STAGHORN SCULPIN | 0.135 |  |
| 304 | MOUSTACHE (MAILED) SCULPIN | 0.086 | 0.004 |
| 306 | ARCTIC HOOKEAR SCULPIN | 0.085 | 0.000 |
| 313 | TWOHORN SCULPIN | 0.000 |  |
| 314 | SPATULATE SCULPIN | 0.056 | 0.000 |
| 320 | SEA RAVEN | 0.021 | 0.029 |
| 340 | ALLIGATORFISH | 0.144 | 0.008 |
| 341 | ARCTIC ALLIGATORFISH | 0.031 |  |
| 350 | ATL SEA POACHER | 0.003 | 0.025 |
| 361 | THREESPINE STICKLEBACK | 0.000 |  |
| 410 | MARLIN-SPIKE GRENADIER | 0.000 | 0.202 |
| 502 | ATL SPINY LUMPSUCKER | 0.070 | 0.023 |
| 505 | SEASNAIL,GELATINOUS | 0.000 |  |
| 512 | SEASNAIL,DUSKY | 0.062 | 0.000 |


| code | species | Proportion <br> September | Proportio <br> $n$ July |
| :--- | :--- | ---: | ---: |
| 610 | NORTHERN SAND LANCE | 0.008 | 0.000 |
| 616 | FISH DOCTOR | 0.070 | 0.000 |
| 620 | LAVAL`S EELPOUT | 0.168 | 0.000 |
| 622 | SNAKEBLENNY | 0.000 | 0.019 |
| 623 | DAUBED SHANNY | 0.028 | 0.032 |
| 626 | 4-LINE SNAKE BLENNY | 0.382 |  |
| 630 | WRYMOUTH | 0.000 |  |
| 632 | STOUT EELBLENNY | 0.004 |  |
| 640 | OCEAN POUT(COMMON) | 0.000 | 0.062 |
| 647 | CHECKER EELPOUT(VAHL) | 0.000 | 0.051 |
| 712 | WHITE BARRACUDINA | 0.000 | 0.000 |
| 880 | HOOKEAR SCULPIN,ATL. | 0.085 | 0.012 |
| 2513 | ATL ROCK CRAB | 0.004 | 0.027 |
| 2521 | HYAS COARCTATUS | 0.208 | 0.002 |
| 2523 | NORTHERN STONE CRAB | 0.000 | 0.077 |
| 2526 | SNOW CRAB (QUEEN) | 0.078 | 0.050 |
| 2527 | TOAD CRAB | 0.130 | 0.041 |
| 2550 | AMERICAN LOBSTER | 0.001 |  |
| 4511 | SHORT-FIN SQUID | 0.018 | 0.011 |


[^0]:    * This series documents the scientific basis for the evaluation of fisheries resources in Canada. As such, it addresses the issues of the day in the time frames required and the documents it contains are not intended as definitive statements on the subjects addressed but rather as progress reports on ongoing investigations.
    * La présente série documente les bases scientifiques des évaluations des ressources halieutiques du Canada. Elle traite des problèmes courants selon les échéanciers dictés. Les documents qu'elle contient ne doivent pas être considérés comme des énoncés définitifs sur les sujets traités, mais plutôt comme des rapports d'étape sur les études en cours.

    Les documents de recherche sont publiés dans la langue officielle utilisée dans le manuscrit envoyé au Secrétariat.

    Ce document est disponible sur l'Internet à:
    http://www.dfo-mpo.gc.ca/csas/

