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Research Document 2004/028

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# Summary of Results of the 2003 Queen Charlotte Sound Bottom Trawl Survey 

## Sommaire des résultats du relevé au chalut de fond effectué en 2003 dans le bassin de la Reine-Charlotte

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http://www.dfo-mpo.gc.ca/csas/
ISSN 1499-3848 (Printed / Imprimé)
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#### Abstract

This document summarizes the general methods and results from the 2003 groundfish bottom trawl survey in Queen Charlotte Sound and southern Hecate Strait. The survey conducted 239 useable tows in depths of 50-500 m from July 3August 9 on board the F/V Viking Storm. The survey was jointly conducted and funded by the Canadian Research and Conservation Society, and Fisheries and Oceans Canada. The objective of this year's survey was to examine its capability to provide long-term indices of relative abundance for fish species affected by bottom trawling, primarily in the survey area.

Results indicate that if the survey were repeated in its current design it could meet its primary objective and would cost approximately $\$ 312,000 / \mathrm{y}$. It will also provide a research platform that will contribute essential biological samples, and oceanographic information. The document recommends that the survey be continued for the planned three years with minor modifications that will be identified with additional analyses of the 2003 results. The additional years will provide insight into the magnitude of the interannual process error and thus be used to determine the optimal frequency of the survey.


## Résumé

Ce document résume les méthodes et les résultats du relevé au chalut du poisson de fond réalisé en 2003 dans le bassin de la Reine-Charlotte et le sud du détroit d'Hécate. Dans le cadre du relevé, le navire de pêche F/VViking Storm a effectué 239 traits de chalut utilisables, à des profondeurs de 50 à 500 m , du 3 juillet au 9 août. La Canadian Research and Conservation Society et Pêches et Océans Canada ont réalisé et financé conjointement le relevé. Cette année, le relevé avait pour objectif d'examiner sa capacité de fournir des indices à long terme de l'abondance relative d'espèces de poissons touchées par le chalutage de fond, principalement dans la région du relevé.

Les résultats indiquent que si l'on répétait le relevé selon son plan actuel, il permettrait d'atteindre son objectif principal et coûterait environ 312000 dollars par année. Le relevé constituera une plate-forme de recherche qui fournira des échantillons biologiques essentiels et des données océanographiques. Le document recommande de poursuivre le relevé durant les trois années prévues en y apportant de légères modifications qui seront précisées à la suite d'analyses supplémentaires des résultats de 2003. Les années de relevé supplémentaires permettront d'estimer l'erreur de méthode d'une année à l'autre et d'ainsi déterminer la fréquence optimale du relevé.

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## Introduction

A previous working paper (Sinclair et al. 2003) recommended further development of fisheries independent, relative abundance indices using bottom trawl surveys in British Columbia (B.C.) waters. It presented an analysis predicting that a coastwide sampling density of 1,000 bottom tows would provide adequate estimates of relative abundance over time to support stock assessment for a majority of the groundfish species affected by trawling. However, the report emphasized that these benefits would only accrue over the longer term (10-20 y). The precision would be too low to accurately characterize modest changes in abundance over short periods for most species.

The document concluded that this approach should be initiated for most of the B.C. coast not currently covered by other surveys. As an interim step, it recommended that a pilot survey be conducted in a reduced area, at the same spatial sampling intensity, to verify the predicted precision. Furthermore it recommended that this survey should be conducted in PMFC major areas 5A and 5B (Queen Charlotte Sound: QCSd). This area was recommended for a number of reasons. Firstly, it was not covered by an existing bottom trawl survey. Secondly, this area represents a significant portion of the bottom trawl fishery. Thirdly, the large proportion of trawlable bottom in this area would allow the survey to cover the full spectrum of commercial species in the $50-500 \mathrm{~m}$ depth range. Finally, as the central region of the coast, it would be a useful starting point for building relative indices for tracking abundance and collecting samples of relatively minor species that might be affected by trawling but without commercial value.

The Pacific Scientific Advice Review Committee (PSARC) and the Regional Management Executive Committee (RMEC) accepted these recommendations. In February 2003, the Canadian Groundfish Research and Conservation Society (CGRCS) committed to funding the vessel and net costs, and a significant portion of the staffing costs needed to conduct the survey and analyze the results. The Science Branch of Fisheries and Oceans Canada (DFO) committed to funding additional scientific staff, and to provide the scientific sampling equipment.

Final design details were approved in May and the survey was conducted from July to August of 2003. This document summarizes the general methodology, results and costs of the 2003 survey, and comments that it appears to be a cost-effective means to provide adequate relative abundance indices of groundfish stocks in the survey area. Full results and methodology will be published under separate cover. If the survey is to be continued, a more detailed analysis will be conducted in early 2004 to fine-tune the design and the onboard activities.

## Methods

## General survey outline

The survey was conducted aboard the commercial stern trawler, F/V Viking Storm, from July 3 to August 9. Mid to late summer was chosen to ensure that the winter to summer movements of the fish had stabilized and to take advantage of better weather. The personnel consisted of the fishing captain, four deckhands and five scientific staff. The survey was conducted in four legs of eight, 11, 11 and eight days.

## Fishing Design

The study area extended from 50-500 m in bottom depth. It did not include inlets and enclosed waters on the eastern borders of QCSd (Figure 2). The southern boundary extends from Hope Island in the southeast corner at the entrance of Queen Charlotte Strait to Cape Scott, and then follows Cape Scott Islands to $50^{\circ} 52^{\prime} \mathrm{N}$, then west to 500 m . The western boundary continues to the northwest from $50^{\circ} 52^{\prime} \mathrm{N}$ following the 500 m contour off the west coast of Queen Charlotte Islands to include $52^{\circ} 20^{\prime} \mathrm{N}$ (the new Moresby Gully Pacific ocean perch boundary part way up the west coast of the Queen Charlotte Islands). The northern Boundary cuts across Hecate Strait at $52^{\circ} 40^{\prime}$ N.

We moved north from the 5B-5C boundary to be approximately contiguous with the southern extent of Hecate Strait Assemblage survey. It thus includes the South Moresby grounds (Oil Drum, Hippa Spot), Ramsey Island, and NW Middle Bank. We have excluded the protected sponge reefs and the hook-and-line rockfish closed areas near Cape Scott and Cape St. James (Figure 2). The overall survey surface area (50-500 m) is about $28,308 \mathrm{~km}^{2}$ (Table 1).

Most of the survey, excluding the southern Hecate Strait portion area falls within Major Areas 5A and 5B. In 2003, total groundfish catches (excluding Halibut) were about $18,000 t$ with Pacific ocean perch the dominant species.

| Pacific ocean perch | $4,382.3$ |
| :--- | ---: |
| Arrowtooth flounder | $1,942.8$ |
| Yellowmouth rockfish | $1,485.7$ |
| Yellowtail rockfish | $1,329.0$ |
| Widow rockfish | $1,178.6$ |
| Lingcod | $1,071.9$ |
| Big skate | $1,049.9$ |
| Rock sole | 770.0 |
| Redbanded rockfish | 660.7 |
| Silvergray rockfish | 546.7 |
| Other species | $3,883.9$ |
|  |  |
| Total | $1,8301.4$ |

We used a sampling target of 240 tows for the proposed area under allocation scheme five from Sinclair et al. (2003). This target is less than the 400 tows recommended. Noting that DFO-Groundfish staff conducted 11 other surveys in 2003, we concluded that the 400 tows was not a sampling effort we could hope to sustain nor a sampling density we could hope to extend to other areas. We also concluded that the marginal improvement in precision of adding 160 more tows would not justify the expense. Since sample precision increases with $\sqrt{n}$, increasing from 240 tows to 400 tows would improve sampling precision by only $30 \%$. Furthermore, assuming an additional process error CV of 0.2 (Francis et al. 2003), the marginal gain would be less than $30 \%$. We used the depth strata recommended in Sinclair et al. 2003, namely:

$$
\begin{aligned}
& 50 \mathrm{~m}<D \leq 125 \mathrm{~m}(27 \mathrm{fm}-68 \mathrm{fm}) \\
& 125 \mathrm{~m}<D \leq 200 \mathrm{~m}(68 \mathrm{fm}-109 \mathrm{fm}) \\
& 200 \mathrm{~m}<D \leq 330 \mathrm{~m}(109 \mathrm{fm}-180 \mathrm{fm}) \\
& 330 \mathrm{~m}<D \leq 500 \mathrm{~m} \quad(180 \mathrm{fm}-273 \mathrm{fm})
\end{aligned}
$$

We used two spatial strata (one interior boundary) separated by Mitchell's and Reed Troughs (Figure 2). This separates a core Queen Charlotte Sound area, including Cape Scott Bank and Goose Bank from a southern Hecate Strait/Cape St. James/Middle Bank area. The size and importance of these two regions appear congruent with the northern Hecate Strait region covered by the Hecate Strait Assemblage study. Species are more mixed in the southern portion; the northern portion tends to be dominated by rockfish (Figure 2). We therefore have eight strata within the QCSd Survey. Using the gullies to bound zones is a poor decision for deeper, gully-dwelling creatures like Pacific ocean perch, but more appropriate for the shallower species. For example, rock sole populations are more likely to conform to the banks than the gullies.

We used Option Five allocation strategy by depth (Sinclair et al. 2003). This allocation scheme is influenced by the variance in total catch in the commercial tows as follows:
"The fifth allocation scheme requires an additional analysis from historical data, in which all biomass is treated as if it were one species. This gives estimates of the parameters $\theta_{h s}$ for a hypothetical species $s$ composed of all fish biomass captured by the tows under consideration" (Sinclair et al. 2003: p7).

This choice contradicted the recommendation in the earlier document to weigh by area of the strata (Option Two). We chose Option Five to shift the focus to areas that are fished more intensively, and to provide more observations on rockfish species, which tend to exhibit more variable catch rates. From Fig. 9 of Sinclair et al. 2003, it was apparent that Option Five is superior for the deeper rockfish species at the expense of shallower nonrockfish species. The main effect was to allocate a few tows away from the shallow stratum to the two intermediate depth strata (Figure 3).

## Selection of blocks, tow definition and tow location

The sampling element was defined as a $4 \mathrm{~km}^{2}$ block. Figure 4 shows the initial random selection of 240 blocks following the weighted selection strategy outlined above. We anticipated having to reject blocks owing to untrawlable bottom, therefore we selected an additional 80 secondary blocks, in advance, to be used to replace rejected original blocks. The protocol was to choose the nearest secondary block to the rejected primary block in the same stratum. Near the end of the survey through a combination of using up all secondary tows in some strata, and having to make pragmatic choices in order avoid excessive travelling time, we adopted an ad hoc-random approach. In this case, as we rejected a block, we identified from the same stratum the nearest polygon (10-20 blocks) of trawlable blocks as identified by the fishing skipper. We then chose, at random, a replacement from within that polygon.

Fishing commenced at sunrise and finished before sunset (approximately 06002130). For each tow, sensors in the net recorded depth, temperature, door spread, headrope height, and degree of bottom contact, at 1-second time intervals. Redundant records of location and net mensuration were manually recorded at approximately 5-minute time intervals in the wheelhouse, as was the time of winch release and lockup. We also collected temperature at depth using a SEABIRD 39 temperature/depth probe attached near the headrope.

Target tow length was 20 minutes on bottom with minimum on-bottom duration of 15 minutes. At least half of the bottom-time of the tow had to be within the block, and the tow was to follow the depth contours and pass through the centre of the block, if possible. The fishing scope (amount of trawl warp deployed) was at the discretion of the captain. If the net hung up after 15 minutes, the tow was still considered useable, as long as the net was retrieved quickly and without significant damage. If a hang-up occurred earlier than this, the tow was rejected and the survey block was either re-attempted or rejected. Some blocks were rejected prior to fishing, when the fishing captain deemed the block to be untrawlable based on sounder information and prior knowledge. All rejected blocks were replaced by alternates. In total, 253 survey tows were attempted. Fourteen tows had to be aborted. Second attempts to fish the same block, and in one case three attempts, succeeded in providing useable tows for four of the 11 blocks.

The catches from 233 of the 239 useable tows were completely sorted and weighed to species. The total catch in six large tows (> $1,500 \mathrm{~kg}$ ) was estimated by the skipper. We then sorted and weighed the total catch of all species except the dominant species, which was estimated by subtraction. Between tows, we sampled as many species as possible, some for length/sex/maturity and some for length/sex/maturity/age. No other specialty sampling or activities were conducted during the survey.

Choice of species to sample was ad hoc. We attempted to sample all the dominant species in the tows plus additional species that were considered higher priority (i.e., lingcod, bocaccio, Pacific cod). If the survey is continued, we plan to use the results of 2003 to provide a more rigorous basis for selecting samples. This will include a review of the objectives of the sampling with respect to the purpose of this survey as well as the overall needs for groundfish research.

## Estimation of survey relative errors in estimating catch density

We estimated survey precision by calculating the coefficient of variation (CV) around the estimates of biomass. The swept area biomass estimates were determined as the catch rate per swept area expanded by the total area in each stratum. The area covered by each tow (swept area) was calculated as the mean doorspread (the distance between the trawl doors) multiplied by the distance that the vessel travelled during the tow. Doorspread measures were obtained electronically from sensors and were recorded throughout each tow at one-second time intervals. The distance travelled during each tow was determined using the Great Circle Distance (GSD) formula with latitude and longitude obtained from GPS and recorded at one second time intervals. Details of the GCD calculation can be found at http://mathworld.wolfram.com/GreatCircle.html.

The observations were analysed using the following equations. The biomass in any year $y$ was obtained by summing the product of the CPUE and the area surveyed across the surveyed strata $i$ for each species $s$ :

$$
\begin{aligned}
B_{s}=\sum_{i} C_{s_{i}} A_{i}= & \sum_{i} B_{s_{i}} \\
\text { where } \quad C_{s_{i}} & =\text { mean CPUE density }\left(\mathrm{kg} / \mathrm{km}^{2}\right) \text { for species } s \text { in stratum } i \\
A_{i} & =\text { area of stratum } i\left(\mathrm{~km}^{2}\right), \text { and } \\
B_{s_{i}} & =\text { estimated biomass of species } s \text { in stratum } i .
\end{aligned}
$$

Eq. 1

The variance of the survey biomass estimate $V_{B_{s}}$ for species $s$ is calculated in $\mathrm{kg}^{2}$ as follows:

$$
V_{B_{s}}=\sum_{i} \sigma_{s_{i}}^{2} A_{i}^{2} / n_{i}
$$

Eq. 2
where $\quad \sigma_{s_{i}}^{2}=$ variance of CPUE $\left(\mathrm{kg}^{2} / \mathrm{km}^{4}\right)$ for species $s$ in stratum $i$
$n_{i} \quad=$ number of tows in stratum $i$
CPUE ( $C_{s_{i}}$ ) was calculated as a density in $\mathrm{kg} / \mathrm{km}^{2}$ by

$$
C_{s_{i}}=\frac{\sum_{j=1}^{n_{i}}\left(W_{s_{i} j} / D_{i j} w_{i j}\right)}{n_{i}}
$$

Eq. 3

$$
\text { where } \begin{array}{ll}
W_{s_{i} j} & =\text { catch weight }(\mathrm{kg}) \text { for species } s \text { in stratum } i \text { and tow } j \\
& D_{i j} \\
=\text { distance travelled }(\mathrm{km}) \text { by tow } j \text { in stratum } i \\
w_{i j} & =\text { wingspread width }(\mathrm{km}) \text { for tow } j \text { in stratum } i \\
& n_{i} \\
=\text { number of tows for stratum } i
\end{array}
$$

The CV for each species s was calculated as follows:

$$
\begin{equation*}
C V_{s}=\frac{\sqrt{V_{B_{s}}}}{B_{s}} \tag{Eq. 4}
\end{equation*}
$$

Five thousand bootstrap replicates with replacement were made on the survey data to estimate bias corrected $95 \%$ confidence regions for each survey species (Efron 1982).

## Estimation of relative errors in commercial catch rates

For a comparison of survey CVs with the CVs one could predict from commercial catch information, we used the method described by Schnute and Haigh (2003) as was used to design this survey (Sinclair et al. 2003). This model is derived from a compound binomial/gamma distribution. The binomial component is used to accommodate the large proportion of zero catches that are typical of research trawl surveys. This distribution is defined in terms of three parameters: $p_{s_{i}}$, the proportion of zero tows in stratum $i$ for species $s ; \mu_{s_{i}}$, the mean density $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$ of the non-zero tows in stratum $i$ for species $s$; and $\rho_{s_{i}}$, the coefficient of variation of the non-zero tows in stratum $i$ for species $s$.

The Schnute and Haigh (2003) model then uses these three input parameter values to calculate biomass and variance predictions for each stratum and for the entire survey, given a specified allocation of tows $n_{i}$ and an area $A_{i}$ for each stratum and an underlying compound binomial/gamma distribution. First, the quantities $\delta_{s_{i}}$ and $v_{s_{i}}$ are calculated for each species $s$ in stratum $i$ :

$$
\begin{align*}
& \delta_{s_{i}}=\left(1-p_{s_{i}}\right) \mu_{s_{i}} \\
& v_{s_{i}}=1 / \rho^{2} \tag{Eq. 5}
\end{align*}
$$

The biomass for species $s$ in stratum $i$ is then:

$$
B_{s_{i}}=\delta_{s_{i}} A_{i}
$$

Eq. 6
and $\sigma_{s_{i}}^{2}$, the variance associated with the CPUE for species $s$ in stratum $i$ is:

$$
\sigma_{s_{i}}^{2}=\left(1-p_{s_{i}}\right)\left(1+v_{s_{i}} p_{s_{i}} \mu_{s_{i}}^{2} / v_{s_{i}}^{2},\right.
$$

Eq. 7
The total biomass, variance and CV for species $s$ using the tow allocation scheme $n_{1} \ldots n_{j}$ and the areas $A_{1} \ldots A_{j}$ for $j$ strata can then be calculated using equations 1,2 and 3 .

Values for $p_{s_{i}}, \mu_{s_{i}}$, and $\rho_{s_{i}}$ were calculated for each of the 24 species examined in Sinclair et al. 2003 and based on commercial catch information for tows conducted in each aerial stratum between 1997 and 2002 at each of the four depth ranges used in this survey. Because the survey aerial strata do not conform to standard DFO management units, only tows with valid latitude and longitude positions could be included. These parameter values were then combined with information on the survey area and the number of tows achieved during the 2003 QCSd survey to obtain predicted CVs. This set of parameters constitutes the data that would have been used if the method used in Sinclair et al. (2003) had been applied to the specific area surveyed in 2003.

## Results

## General results

From the total of 38 days, 3.25 days were required for travel at the start and end, three days were required for unloading and personnel change during the survey, $0.753 / 4$ day was lost to equipment breakdown on the vessel, and three days were lost to weather. Our original estimate was six to seven weeks. We completed 239 successful survey tows plus three successful camera tows for a gross average of 6.4 tows/y. We averaged over nine useable survey tows/y on full fishing days.

From the total of 239 useable tows, 202 were chosen from the original set, 22 from the secondary set and 15 were chosen using the ad hoc approach. Blocks were initially assigned to depth strata based on nautical chart depths. Owing to actual placement of the tow and errors in the interpolated depths in the charts, some completed tows had to be reassigned to different depth strata (Table 2). Each tow's depth stratum was determined as the modal stratum indicated in the SEABIRD 1-second fixes. Fishing skippers generally attempted to fish the centre of the block parallel to the depth contours (Figure 6).

## Estimation of trawlable area or blocks

Since the primary blocks were chosen randomly from within each stratum, we can use the proportion deemed untrawlable to calculate the proportion of the survey area that is untrawlable. From Table 3, over $83 \%$ was shown to be trawlable. While we have few observations in some strata, these same strata represent a small proportion of the survey area, thus we can be confident that this is a reasonable estimate. Given there are 7,552 4$\mathrm{km}^{2}$ blocks in the survey area; we can assume that about 1,280 blocks are untrawlable. It is tempting to suggest that conducting the survey will be come more efficient, as we eliminate untrawlable blocks, however, since we only removed 40 blocks in 2003, this will take a long time.

During the planning phase, fishing skippers noted that while they could predict blocks that were likely untrawlable, they should all be examined as a few should prove fishable for 20 -minute tows. This assumption was borne out during the survey.

## The variability in survey estimates

The 239 useable tows captured 88 t of fish, $37 \%$ of this was Pacific ocean perch (Table 4, Appendix Table 10 and Appendix Table 11). The average fish catch in the useable tows was 377 kg (Figure 7). We recorded 105 fish species or species groups. Of these, 92 were identified to species and 13 were grouped (e.g., pricklebacks). We typically observed 10-20 species per tow (Figure 8).

We examined the precision for the original 23 species considered during the design phase (Sinclair et al. 2003) and an additional 21 species, which had at least ten observations among the 239 useable survey tows (about $4 \%$ of the total tows). Three of the species in the original list of 24 species did not meet this 10 -tow criterion (shortraker rockfish: eight tows; wolf eel: three tows; and sand sole: 0 tows) but were kept, as they were part of the first list.

The estimates of survey precision and area swept biomass are summarized for the 44 species, ranked in order of precision (Table 5). For those species, which we can assume might be assessed separately by region, we have reported the results for the two spatial strata separately. Results for the same species for the entire survey area combined are provided in Table 7 and Figure 9 (see Table 6 for definition of acronyms and Appendix Table 12and Appendix Table 13 for results by area). Of the 52 population or stocks indicated in the Table 5, 14 indicate a CV less than 0.20 . These populations account for about $60 \%$ of both retained and total trawl catch produced by commercial bottom trawling in the region covered by the survey. An additional 20 of these populations are associated with CV's of 0.20-0.40. Combined with the previous category, the 34 stocks account for about $80 \%$ of all retained and total commercial bottom trawl catch from the survey area.

## Comparison with predictions from Sinclair et al. 2003

This comparison of sampling error predictions between those provided by Sinclair et al. 2003 and the 2003 survey results is confined to the list of 24 species developed by Sinclair et al. (2003). One species (sand sole) in the original list of 24 species was not taken at all in the survey, but this outcome was predicted in Sinclair et al (2003). Eight survey CVs were below a $20 \%$ threshold, with the CVs for 23 species ranging from $11 \%$ for arrowtooth flounder to 69\% for wolf eel (Figure 9).

Only 3 of the 23 species had predicted survey CVs based on the Schnute and Haigh (2003) model within $10 \%$ of the observed CV and 13 of 23 were within $50 \%$ of the observed CV (Figure 10 and Figure 11). Eight of the observed CVs were larger than the predicted CVs while 15 were smaller. This result probably indicates that the method does not seem to consistently over- or under-predict the survey CVs.

There are eleven species for which the predicted CVs are higher than the bootstrap confidence bounds for the analytic CVs and only one where the predicted CV is below the lower limit of the bootstrap confidence bounds ( Table 7) indicating that this method fails for some species. These species, for which the CVs are higher than the survey bootstrap confidence bounds tend to be species, which are either discarded or actively avoided (Pacific cod, sablefish, dogfish, arrowtooth flounder). This indicates that the CVs for these species may be inflated in the commercial data owing to the behaviour of the commercial fleet rather than to the underlying population variability and which leads to bias in the CV prediction.

Nevertheless, the results indicate that the design number of tows was probably not excessive, given that only eight of the 23 survey CVs were below the $20 \%$ threshold and only a further two more were between 20 and $30 \%$ (Table 5). This indicates that basing the design of a multispecies survey on a suite of species taken in the commercial fishery is a reasonable tool to use in situations where there are few alternative data sources.

In addition to using commercial catch rates to predict precision, one can use other surveys. For the overlapping species, results from the QCSd survey appear similar to those for the NMFS 2001 survey for the Vancouver Area (Weinberg 2001) (Figure 12). The NMFS survey CV's have been prorated from 79 tows to 239, assuming that precision is proportional to $\sqrt{n}$.

## Graphical simulation

To illustrate how well the proposed survey would track a known population, we developed a simulator that uses sample size (number of tows), sampling error for each stock in the proposed survey area from the 2003 results, an estimate of process error of 0.2 (Francis et al. 2003), and a specified biomass trend (Figure 13 and Figure 14). As noted in Sinclair et al. 2003, the effective precision of a survey is actually determined by the combined effects of the sampling error, discussed at length above, and process error.

$$
C V_{\text {TOTAL }}=\sqrt{\left(C V_{\text {SAMPLING }}\right)^{2}+\left(C V_{\text {PRoCESS }}\right)^{2}}
$$

The sample error is proportional to variability among observations within one year and sample size. The process error is the additional variability added by those influences that can vary among years, excluding changes in actual abundance. This could be caused by changes in fishing power brought about by different nets, vessels or captains. It can also be caused by variation in the availability or vulnerability of the fish owing to variation in their environment from one year to the next. We have no means of estimating this component at present so have assumed a general value of 0.20 from Francis et al. 2003. It is somewhat simplistic to assume a constant $C V_{\text {PROCESS }}$; it is more likely to be proportional to $C V_{\text {SAMPING }}$, however we cannot assess this at this time. The effect on $C V_{\text {TotaL }}$ for various levels of $C V_{\text {SAMPLING }}$ is shown below:

| $C V_{\text {SAMPLING }}$ | $C V_{\text {PROCESS }}$ | $C V_{\text {TOTAL }}$ |
| :---: | :---: | :---: |
| 0.1 | 0.2 | 0.22 |
| 0.2 | 0.2 | 0.28 |
| 0.3 | 0.2 | 0.36 |
| 0.4 | 0.2 | 0.45 |
| 0.5 | 0.2 | 0.54 |
| 0.6 | 0.2 | 0.63 |

Each display is tailored to the presumptive stock assessment needs for a given species. Pacific ocean perch is modelled over 20 years, for the southern survey section only (Sea Otter and Reed Troughs) ( $\mathrm{n}=121$ ) with abundance increasing $5 \% / \mathrm{y}$. Rock sole abundance for the whole survey area ( $\mathrm{n}=239$ ), and presumably more dynamic over time, is simulated over 10 years, and we assume abundance is decreasing $15 \% / \mathrm{y}$. The LOWESS fit indicates how an assessment might "perceive" the biomass trajectory as opposed to the true trajectory.

The impressions these graphics provide are obviously influenced by choice of scaling, magnitude and shape of the trends, and the duration. Nevertheless they are presented to provide insight as to how survey results might "appear" as each is added to a stock assessment model. We can easily generate alternative scenarios or more examples on request. The random examples shown were the first three and were not selected.

The simulations indicate that for those species with higher precision ( $C V_{\text {SAMPLING }}<0.3$ ), as for Pacific ocean perch and rock sole, the survey will successfully capture general trends through time, but individual survey points will often appear anomalous. For a $C V_{\text {SAMPLING }}$ of 0.369 for canary rockfish, the general trend can still be captured, but for these species short term trends of 2-3 years can appear to be moving in opposite direction of the overall trend. For species with even lower precision, the survey may incorrectly indicate trends for up to seven years. It is clear that for the species associated with low precision, the survey will be useful only in identifying large population shifts over the long term.

## Additional survey results

The survey provided 1,372 samples of over 30,000 specimens (Table 8), averaging about six species/tow. About $55 \%$ of the samples included ageing parts in addition to length, sex and maturity. The average number of pieces per sample was 20 and 25 specimens for ageing and length/sex samples respectively. This is smaller than the targeted 50 pieces per sample because many samples represented all the specimens captured in the tow.

Bottom contact reading and temperature at depth were recorded during the survey on virtually every tow. One example of how survey data could be integrated to study the sources of inter-annual variance ('process error') in abundance is provided in Figure 15.

## Summary of costs

We estimate that the ongoing costs of the charter vessel will be about $\$ 7,250 / \mathrm{y}$ for an annual cost for a 6 -week charter of about $\$ 304,500$. The annualized costs of purchasing and maintaining two fishing trawls would be about $\$ 5,000$. This estimate is based on a purchase price for two complete nets and footropes of \$50,000, a depreciation of $10 \% / \mathrm{y}$, and net maintenance costs of $\$ 5,000 / \mathrm{y}$. It also assumes that the nets will be used $50 \%$ of the time by other surveys, thus, use by the QCSd survey will account for $50 \%$ of the annualized costs. We have ignored the purchase and maintenance price of net sensors as the charter vessels will continue to use these during commercial fishing.

We estimate that the annualized cost for purchasing and maintaining scientific field equipment would be approximately $\$ 5,500$. This estimate is derived by taking $6.7 \%$ of the replacement price of $\$ 83,000$ for a full set (including backups) of all survey equipment (i.e., contact sensors, temperature/depth probes, motion compensating balances, automatic fish measuring boards, laptops). The $6.7 \%$ factor is derived from an estimated depreciation rate of $20 \%$, divided by three, based on the assumption that the QCSd survey will represent $33 \%$ of the usage.

If we assume a cost of $\$ 500 / \mathrm{y}$ to put a scientific staff member on the survey then staffing the survey with five scientific staff for a six week cruise equals an annual cost of $\$ 21,000$. Cruise preparation will require about 15 working days at $\$ 400 / \mathrm{y}$ equalling $\$ 6,000$. Data processing after the cruise will require 40 working days at $\$ 400 / \mathrm{y}$ or approximately $\$ 16,000$. The total costs of scientific staff will be approximately $\$ 43,000 / \mathrm{y}$. We have not included the cost of analyses.

The cost estimates summarized above indicate a total annual expenditure by all participants combined of about $\$ 358,000$. Funds returned by the sale of fish in 2003 equalled $\$ 45,707$, therefore we can expect the annual costs of this charter in 2003 to be about $\$ 312,000$.

## Discussion

## Was the survey an operational success?

There were no problems in conducting the survey. The proposed design of 240 tows, allocated among eight strata was completed in 38 days, less time than predicted. There were no surprise costs and little time was lost to equipment breakdown or weather. The lengths of trip legs ( $9-11$ days) helped significantly to reduce discarding of commercially valuable catches. The majority of the grounds, over $80 \%$, appear to be trawlable given the choice of trawl net and footrope.

The net was not only able to negotiate most of the bottom but appeared to provide an acceptable "catchability" of the main species. The net in combination with the choice of $20-$ minute tows provided an acceptable catch rate for most species of interest and yielded an adequate number of biological sampling opportunities. Shorter tow lengths ( 20 vs. 30 minutes) successfully reduced the likelihood of very large tows, produced a workable average catch, probably increased the proportion of trawlable ground, and allowed the survey to complete an extra 0.25-0.50 tows/y.

## Was it a success from indexing point of view?

We do not know of an accepted standard for whether a survey provides adequate precision. Sinclair et al. (2003) proposed a target CV of 0.20 . While this is a reasonable standard for considering one species at a time, the total benefit accruing from a survey must also address the number of different indices being generated. Secondly, as commented by a reviewer of Sinclair et al. (2003), a 0.20 standard might be overly rigorous for long-lived and less variable species which might only need assessment on five, ten or $15-y$ intervals. Assessment of these species, which vary in abundance more slowly over time, benefit from the "pooling" effect of repeated annual estimates. This in turn indicates that a CV standard is dependent on the frequency of the survey. Finally, the adequacy of a standard must also address what alternatives are available. Recent discussion involving species-at-risk issues (Stanley et al. 2003) would indicate that pertinent insight has been derived from surveys with very poor precision.

Notwithstanding the comments above, if we assume that the survey frequency will be every year or at least every two years, we suggest for discussion purposes that we assign the following descriptors for the range of observed stock $\mathrm{CVs}\left(C V_{\text {SAMPLING }}\right)$ :

- "excellent" = <0.20;
- "good" $=$ 0.20-30;
- "adequate" = 0.30-0.40;
- "poor" $=$ 0.40-0.60;
- "very poor" $=>0.60$.

Using the above descriptors, Table 5 indicates that 240 tows in the QCSd survey can provide at least adequate precision for 34 stocks, which represent $80 \%$ of the biomass landed or captured in the study area.

We note that the above group includes numerous species/stocks associated with contentious TACs. It also includes species, such as some skates, of relevance to SARA issues. The general impression is that the survey is worth the cost, provided the added imprecision owing to $C V_{\text {PROCESS }}$ does not overwhelm the ability of the survey to track trends. We should derive insight into the $C V_{\text {PROCESS }}$ by comparing 2003 results with those to be gathered in 2004 and 2005.

The survey might be improved by re-allocating 240 tows among the existing strata, but the gain will likely be modest and only improve precision for some species at the expense of others. These issues will be examined if the survey is continued. The overall target sampling density of 240 appears appropriate. We recommend against a reduction. Furthermore, a modest increase in sampling effort will have negligible impact and not cause the survey to cross some critical precision threshold.

## What additional benefits will be derived from the survey?

The large number of biological samples collected will clearly benefit groundfish research, especially in conjunction with the ongoing collection of samples from the commercial fishery. Since the survey samples would be collected from a nearly constant sampling design they are more comparable over time and possibly more representative of the actual population. The commercial samples, which are more representative of the harvested portion of the population, are influenced by spatial and temporal trends in the fishery.

As trends emerge in the relative index, the biological sampling will provide a better understanding of why the trends occur. In particular they will help distinguish between the impacts of fishing and "natural" variation in recruitment.

We demonstrated in Figure 15 the potential for collecting additional information during the survey. The accuracy of these oceanographic data is acceptable for physical oceanographic research, and these data will be appended to large-scale global oceanographic databases. They may also be useful for explaining short term anomalies in the relative abundance trends (i.e., explaining some of the process error) in future stock assessments as well as detecting ocean climate changes on the actual fishing grounds.

While we were close to working capacity in the 2003 survey, we expect to modestly expand survey activities and take better advantage of the research platform provide by this survey. This could include specialty sampling such as obtaining genetics samples or tagging. We note however, that the ability to multi-task depends on the vessel chosen for the survey.

## Predictions based on commercial data and other surveys

In the process of examining the feasibility and designing a coastwide synoptic survey approach, Sinclair et al. (2003) used commercial catch and effort data to predict survey precision as well as explore alternative stratification and tow allocation models. The data collected by the 2003 QCSd survey was used to test the predictions made by Sinclair et al. (2003) and to refine the use of commercial catch and effort data in designing future bottom trawl surveys on the British Columbia coast. It is clear that commercial fishery data can be useful for drawing inference about the feasibility of survey designs especially when there are no alternative data sources.

Precision of the NMFS triennial survey in INPFC Vancouver region also appeared congruent with results of the 2003 QCSd survey. This is not surprising since the US survey targets the same depth range and has similar objectives. We can assume that the NMFS survey will be especially useful for designing a Canadian survey of the Vancouver area, off the west coast of Vancouver Island. While not presented in this document, we also briefly looked at Hecate Strait results in comparison with 2003 QCSd results. As might be expected, the shallower Hecate Strait survey provided more precise estimates for shallowwater species than did the QCSd survey, but proved less precise for the deeper species.

## Absolute biomass estimates

Only 5 species registered mean bootstrap biomass estimates greater than $2,000 \mathrm{t}$, with the largest biomass levels associated with Pacific ocean perch and arrowtooth flounder (Table 5). A plot of the biomass estimates with associated bias corrected $95 \%$ confidence intervals shows proportionately wide bounds for all 44 species (Figure 16).

Not only is there considerable uncertainty around the estimates of biomass, we emphasize strongly that, at best, they represent minimum estimates of actual biomass. There are many assumptions involved with inferring absolute abundance from these data. Among them is the assumption that the net captures $100 \%$ (catchability $=1.0$ ) of all specimens in the total area between the doors over the length of the tow, and from the entire water column. It also assumes that the trawlable fraction of the survey area ( $80 \%$ ) is representative of the untrawlable fraction. These assumptions are certainly untrue for most species. For many species, catchability may be less than $25 \%$ or even $1 \%$ for those species not observed in the survey catches (e.g., prowfish) or some of those represented by fewer than 10 tows (Table 11) less than $1 \%$. Catchability will also be proportional to size within a species. For example, it can be assumed that catchability will differ between large and small halibut. By using distance between the doors and not the wingtips, it is probably reasonable to assume for most species that these biomass estimates are "minimum" estimates of biomass but it is to be remembered that actual biomass values may be orders of magnitude greater. These estimates should not be used directly and in isolation of other data to infer status of stocks relative to current harvests.

While we discourage placing confidence in the present estimates of absolute biomass, this conceptual approach could prove beneficial in a limited number of cases and will become increasingly important over time. For those species which are highly vulnerable to trawl gear and might be assumed to have catchabilities that approach 1.0, the "minimum" biomass estimates may provide some guidance during assessment. Over time, it also may be possible to estimate the catchability of some species by comparing stock assessment estimates of biomass with the swept-area estimates from this survey (Millar and Methot, 2002).

## Recommendations

1. The QCSd survey should be continued as proposed for 2004 and 2005 and use the same target of 240 tows. At the end of the three years, with the added insight about interannual variance, we will be better able to assess the precision of the survey and recommend the optimal frequency within the context of an overall groundfish survey strategy (see Recommendation \#4). We suggest that the provision of relative indices of these populations will greatly improve the stock assessment advice that will be provided to managers for all these populations, but caution that it will require many years for that impact to be manifest.
2. The 2003 results should be examined to determine whether a reallocation of the tows among strata will improve precision prior to conducting the 2004 survey.
3. Based on the better knowledge of the expected number of sampling opportunities and on-board sampling capability during the charter, the current ad hoc method of choosing samples should be made more rigorous to ensure comparability over time.
4. DFO in collaboration with its research partners should develop a PSARC document outlining a comprehensive, coastwide groundfish survey strategy for submission to 2004 PSARC meeting.

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## Acknowledgements

The authors wish to acknowledge assistance from the overall working group, which also included Jon Schnute, Sandy McFarlane, Alan Sinclair, Greg Workman, Brian Krishka, Brian Dickens, Brian Mose, Kelly Anderson and Chris Roberts. We also appreciate the hard and effective work by the field staff and boat crew on board the F/V Viking Storm. The planning of the survey also received valuable assistance from Mark Wilkins and Guy Fleischer from the U.S. National Marine Fisheries Service. Field operations and trip preparation was also assisted by Bruce Turris of the CGRCS and Kate Rutherford of DFO.

## Request for Working Paper

## Date Submitted: October 28, 2003

Individual or group requesting advice: Science, Management and Fishing Industry

## Proposed PSARC Presentation Date: December 2003

## Subject of Paper:

## Summary of results of the 2003 Queen Charlotte Sound Bottom trawl survey

Science Lead Author: Rick Stanley

## Rationale for request:

A 5-week bottom trawl survey of Queen Charlotte Sound and Southern Hecate Strait was conducted in 2003 by Fisheries and Oceans Canada and the Canadian Groundfish Research and Conservation Society. The primary objective of the 2003 survey was to examine the feasibility of using this survey to provide long-term relative abundance indices for groundfish populations in this area. It was also expected that such a survey would provide a standardized vehicle for collecting groundfish biological samples as well as having other collateral benefits.

While there is little doubt that such a survey would assist groundfish research and management, repeating this survey at regular intervals represents a major long term commitment of groundfish research resources. Therefore there is an immediate need to examine the 2003 results and determine whether the survey should be continued and, if so, identify how the design could be improved.

## Objective of Working Paper:

The primary purpose of this document is to examine whether the survey will meet its objective of providing a relative index of abundance for a suite of groundfish species/populations at acceptable levels of precision while keeping costs to a reasonable level. The second purpose of this study is to examine whether the design is optimized and what changes could be made to improve the precision. The report will conclude with recommendations for whether the survey should be continued, what the long-term costs will be, and, if it is to be continued, what major changes should be made.

## Question(s) to be addressed in the Working Paper:

1. What will be the expected precision in the relative abundance indices for each groundfish species/population monitored by this survey?
2. With respect to precision, does the survey appear to be capable of meeting the intended stock assessment and groundfish management objectives?
3. Can the survey be modified to improve precision without increasing costs?
4. Was the survey design followed during the execution of the survey?
5. What were the costs of the survey and what are the implications of these costs to meet the survey objectives?
6. What additional benefits will the survey provide?
7. Should additional activities be added to the survey?
8. Should the survey be continued and at what frequency?

## Stakeholders Affected:

This survey is intended to provide trends in abundance and biological composition for all groundfish species vulnerable to bottom trawl, which inhabit depths of 50-500 m , in the central region of the coast. It therefore will provide critical assessment information for virtually all exploited groundfish stocks in this area. Additionally, since the survey covers a large portion of the central B.C. coast, it will be used to index coastwide abundance for those species that lack spatial structure or those for which the stock structure is unknown. The latter group of species has particular relevance with respect to SARA-related issues.

Therefore, the survey should benefit all groundfish harvesters (commercial, First Nations and recreational) directly and will address a significant element of the SARA information needs as they pertain to groundfish.

## How Advice May Impact the Development of a Fishing Plan:

Results of the survey will have little immediate impact on Fishing Plans. It is expected that commercial catches in the survey are too small to affect TAC or IVQ management.

Over the longer term, results of the survey will have major impact on the stock assessment advice provided prior to development of fishing plans.

## Timing Issues Related to When Advice is Necessary:

The decisions on whether to proceed with this survey must be available quickly so that preparations can begin for the 2004 groundfish field season.

Table 1. Stratum designations, number of useable tows, and total area for each stratum in the 2003 Queen Charlotte Sound survey.

| Stratum <br> number | Area <br> designation | Depth <br> zone | Number <br> tows | Area <br> $\left(\mathrm{km}^{2}\right)$ |
| :---: | :--- | :--- | ---: | ---: |
| 18 | $5 \mathrm{AB}-1$ | $50-125 \mathrm{~m}$ | 29 | 5428 |
| 19 |  | $125-200 \mathrm{~m}$ | 56 | 5700 |
| 20 |  | $200-330 \mathrm{~m}$ | 30 | 3136 |
| 21 |  | $330-500 \mathrm{~m}$ | 6 | 556 |
| 22 | 5 AB-2 | $50-125 \mathrm{~m}$ | 6 | 2308 |
| 23 |  | $125-200 \mathrm{~m}$ | 39 | 5120 |
| 24 |  | $200-330 \mathrm{~m}$ | 54 | 4760 |
| 25 |  | $330-500 \mathrm{~m}$ | 19 | 1300 |

Table 2. Target number of tows per stratum and actual number of tows per stratum.

| Area <br> Stratum | Depth <br> Stratum | Target <br> Number of Sets | Delivered <br> Number of Sets | Difference |
| :--- | ---: | ---: | ---: | ---: | ---: |

Table 3. Number of blocks trawlable by stratum and estimates of untrawlable area by stratum.

| Area | Depth <br> Stratum (m) | Number <br> of Blocks | Num Blocks <br> Trawlable | Num Blocks <br> Untrawlable | Percent <br> Trawlable | Trawlable <br> Area $(\mathrm{km} 2)$ | Untrawlable <br> Area (km2) |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Stratum | $50-125$ | 26 | 20 | 6 | $76.9 \%$ | 80 | 24 |
| 5AB South | $125-200$ | 60 | 49 | 11 | $81.7 \%$ | 196 | 44 |
|  | $200-330$ | 28 | 22 | 6 | $78.6 \%$ | 88 | 24 |
|  | $330-500$ | 3 | 3 |  | $100.0 \%$ | 12 | 0 |
| Subtotal: |  | 117 | 94 | 23 | $80.3 \%$ | 376 | 92 |
| 5AB North | $50-125$ | 8 | 6 | 2 | $75.0 \%$ | 24 | 8 |
|  | $125-200$ | 42 | 33 | 9 | $78.6 \%$ | 132 | 36 |
|  | $200-330$ | 53 | 49 | 4 | $92.5 \%$ | 196 | 16 |
|  | $330-500$ | 20 | 18 | 2 | $90.0 \%$ | 72 | 8 |
|  |  | 123 | 106 | 17 | $86.2 \%$ | 424 | 68 |
| Subtotal: |  | 240 | 200 | 40 | $83.3 \%$ | 800 | 160 |
| Total: |  |  |  |  |  |  |  |

Table 4. Retained, discarded, total catch (kg) and frequency of occurrence (tows) by fish species.

| Species | Retained Weight (kg) | Discarded Weight (kg) | Total Weight (kg) | Frequency |
| :--- | ---: | ---: | ---: | ---: |
| Arrowtooth Flounder | 167.3 | $7,945.7$ | $8,113.0$ | 212 |
| Rex Sole | 143.2 | $3,142.0$ | $3,285.2$ | 200 |
| Pacific Ocean Perch | $18,608.2$ | $17,385.7$ | $35,993.9$ | 180 |
| Dover Sole | 338.5 | $2,426.9$ | $2,765.4$ | 172 |
| Spotted Ratfish |  | $1,637.2$ | $1,637.2$ | 151 |
| Sablefish |  | $1,975.0$ | $1,975.0$ | 135 |
| Redbanded Rockfish | 396.8 | $1,283.6$ | $1,680.4$ | 129 |
| Silvergray Rockfish | 487.7 | $2,584.2$ | $3,071.9$ | 127 |
| Spiny Dogfish |  | $2,170.0$ | $2,170.0$ | 126 |
| Shortspine Thornyhead | 312.3 | $1,400.4$ | $1,712.7$ | 106 |
| Pacific Cod | 77.2 | 797.7 | 874.9 | 104 |
| Walleye Pollock | 5.9 | 319.2 | 325.1 | 92 |
| Pacific Hake |  | $1,772.9$ | $1,772.9$ | 87 |
| Flathead Sole | 0.5 | 718.0 | 718.5 | 82 |
| Longnose Skate |  | 594.9 | 594.9 | 82 |
| Rougheye Rockfish | 227.3 | $1,323.1$ | $1,550.4$ | 91 |
| English Sole | 14.6 | 925.6 | 940.2 | 78 |
| Petrale Sole | 1.8 | 274.7 | 276.5 | 75 |
| Lingcod | 94.5 | 486.1 | 57 |  |
| Greenstriped Rockfish | 1.0 | 150.3 | 580.6 | 56 |
| Redstripe Rockfish | $2,919.5$ | 483.8 | 151.3 | 51 |
| Pacific Halibut |  | 729.2 | $3,403.3$ | 729.2 |

Table 5. Estimated survey CVs for all stocks represented in the survey observations.

| Species | Stock | Minimum biomass | CVs |
| :---: | :---: | :---: | :---: |
| Rex Sole | Combined | 2438.2 | 0.100 |
| Shortspine Thornyhead | Combined | 1034.8 | 0.108 |
| Arrowtooth Flounder | Combined | 6058.0 | 0.108 |
| Sablefish | Combined | 1149.3 | 0.131 |
| Dover Sole | South | 1531.1 | 0.136 |
| Dover Sole | North | 308.9 | 0.143 |
| Pacific Hake | Combined | 1193.4 | 0.145 |
| Redbanded Rockfish | Combined | 1094.2 | 0.163 |
| Pacific Ocean Perch | South | 17409.4 | 0.167 |
| Walleye Pollock | Combined | 268.6 | 0.183 |
| Flathead Sole | Combined | 564.1 | 0.187 |
| Pacific Ocean Perch | North | 6765.7 | 0.193 |
| Silvergray Rockfish | North | 1964.7 | 0.194 |
| Petrale Sole | Combined | 315.4 | 0.196 |
| Longnose Skate | Combined | 504.6 | 0.208 |
| Pacific Cod | North | 526.0 | 0.213 |
| English Sole | South | 773.3 | 0.229 |
| Pacific Halibut | Combined | 853.2 | 0.237 |
| Lingcod | South | 381.9 | 0.250 |
| Rosethorn Rockfish | Combined | 88.8 | 0.253 |
| Silvergray Rockfish | South | 639.2 | 0.262 |
| Slender Sole | Combined | 86.9 | 0.267 |
| Rock Sole | South | 676.9 | 0.268 |
| Pacific Cod | South | 353.3 | 0.278 |
| Greenstripe Rockfish | Combined | 122.9 | 0.278 |
| Blackfin Sculpin | Combined | 9.8 | 0.288 |
| Pacific Sanddab | Combined | 1185.8 | 0.288 |
| Sandpaper Skate | Combined | 24.4 | 0.312 |
| Curlfin Sole | Combined | 14.4 | 0.313 |
| Blackbelly Eelpout | Combined | 51.8 | 0.334 |
| Eulachon | Combined | 34.7 | 0.342 |
| Yellowmouth Rockfish | Combined | 1714.5 | 0.346 |
| Canary Rockfish | Combined | 1331.3 | 0.358 |
| Threadfin Sculpin | Combined | 9.3 | 0.384 |
| Yelloweye Rockfish | Combined | 256.4 | 0.408 |
| Spiny Dogfish | Combined | 2799.9 | 0.408 |
| Rougheye Rockfish | Combined | 982.9 | 0.436 |
| Yellowtail Rockfish | Combined | 989.3 | 0.442 |
| Shortraker Rockfish | Combined | 71.0 | 0.455 |
| Sharpchin Rockfish | Combined | 751.0 | 0.479 |
| Darkblotched Rockfish | Combined | 138.2 | 0.500 |
| English Sole | North | 377.1 | 0.536 |
| Redstripe Rockfish | Combined | 2828.0 | 0.550 |
| Widow Rockfish | Combined | 182.2 | 0.584 |
| Bocaccio | Combined | 137.3 | 0.661 |
| Lingcod | North | 543.7 | 0.685 |
| Big Skate | Combined | 643.5 | 0.690 |
| Wolf Eel | Combined | 10.0 | 0.704 |
| Rock Sole | North | 57.1 | 0.761 |
| Splitnose Rockfish | Combined | 2934.2 | 0.793 |
| Spotted Ratfish | Combined | 3605.1 | 0.815 |

Table 6. Acronyms for species examined for estimation of survey relative error.

| Arrowtooth Flounder | ARF Rex Sole | RXL |
| :--- | :---: | :---: |
| Big Skate | BIS Rock Sole | ROL |
| Blackbelly Eelpout | BEP Rosethorn Rockfish | RTR |
| Blackfin Sculpin | BSN Rougheye Rockfish | RER |
| Bocaccio | BOR Sablefish | SBF |
| Canary Rockfish | CAR Sand Sole | SAL |
| Curlfin Sole | CUL Sandpaper Skate | SPS |
| Darkblotched Rockfish | DBR Sharpchin Rockfish | SCR |
| Dover Sole | DOL Shortraker Rockfish | SRR |
| English Sole | ENL Shortspine Thornyhead | SSY |
| Eulachon | EUN Silvergray Rockfish | SGR |
| Flathead Sole | FHL Slender Sole | SLL |
| Greenstripe Rockfish | GSR Spiny Dogfish | DOG |
| Lingcod | LIN Splitnose Rockfish | SNR |
| Longnose Skate | LNS Spotted Ratfish | RAT |
| Pacific Cod | PAC Threadfin Sculpin | TSN |
| Pacific Hake | PAK Walleye Pollock | WAP |
| Pacific Halibut | PAH Widow Rockfish | WWR |
| Pacific Ocean Perch | POP Wolf Eel | WOE |
| Pacific Sanddab | PAD Yelloweye Rockfish | YYR |
| Petrale Sole | PEL Yellowmouth Rockfish | YMR |
| Redbanded Rockfish | RBR Yellowtail Rockfish | YTR |
| Redstripe Rockfish | RSR |  |

Table 7. Analytic and bootstrap results for 44 species from the total 2003 Queen Charlotte Sound survey. Analytic results are presented for the biomass (Eq. 1) and the CV (Eq. 4). Bootstrap results are for 5,000 replicate samples taken with replacement. Bias corrected $95 \%$ confidence intervals are presented and the bootstrap CV is calculated relative to the bootstrap mean biomass.
$\left.\begin{array}{lrrrrrr}\hline & & & \text { Bootstrap } \\ \text { mean } \\ \text { Biomass }\end{array} \begin{array}{r}\text { Bootstrap } \\ \text { lower } \\ \text { biomass (t) }\end{array} \quad \begin{array}{r}\text { Bootstrap } \\ \text { upper } \\ \text { bound }\end{array}\right)$

Table 8. Summary of biological samples collected during the 2003 QCSd survey.

| Species name | L/S/M/Age |  |  | L/S |  | Number of tows species caught | $\begin{array}{r} \% \text { of tows } \\ \text { sampled } \\ \hline \end{array}$ | \% of total catch wt. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total Samples | Number of samples | Number of specimens | Number of samples | Number of specimens |  |  |  |
| Arrowtooth flounder | 94 | 29 | 1077 | 65 | 2367 | 213 | 44.1 | 16.6 |
| Aurora rockfish | 1 | 1 | 1 |  |  | 1 | 100.0 | 100.0 |
| Big skate | 12 |  |  | 12 | 49 | 12 | 100.0 | 62.2 |
| Bocaccio | 10 | 10 | 38 |  |  | 12 | 83.3 | 89.1 |
| Canary rockfish | 11 | 10 | 323 | 1 | 1 | 37 | 29.7 | 47.0 |
| Chub mackerel | 1 | 1 | 25 |  |  | 1 | 100.0 | 62.8 |
| Chum salmon | 1 |  |  | 1 | 7 | 3 | 33.3 | 63.7 |
| Curlfin sole | 3 | 2 | 4 | 1 | 2 | 11 | 27.3 | 26.0 |
| Darkblotched rockfish | 13 | 13 | 88 |  |  | 25 | 52.0 | 48.7 |
| Dover sole | 55 | 18 | 684 | 37 | 1009 | 172 | 32.0 | 34.4 |
| Dusky rockfish | 1 | 1 | 1 |  |  | 1 | 100.0 | 100.0 |
| English sole | 29 | 13 | 597 | 16 | 582 | 75 | 38.7 | 34.9 |
| Eulachon | 3 |  |  | 3 | 329 | 40 | 7.5 | 6.4 |
| Flathead sole | 30 | 10 | 385 | 20 | 712 | 82 | 36.6 | 28.9 |
| Greenstriped rockfish | 41 | 39 | 318 | 2 | 20 | 51 | 80.4 | 83.3 |
| Lingcod | 55 | 54 | 189 | 1 | 1 | 56 | 98.2 | 97.6 |
| Longnose skate | 72 |  |  | 72 | 112 | 81 | 88.9 | 86.5 |
| Pacific cod | 100 | 2 | 77 | 98 | 1307 | 105 | 95.2 | 81.3 |
| Pacific hake | 7 | 6 | 178 | 1 | 13 | 87 | 8.0 | 10.2 |
| Pacific halibut | 44 |  |  | 44 | 108 | 47 | 93.6 | 80.5 |
| Pacific ocean perch | 101 | 74 | 3067 | 27 | 846 | 181 | 55.8 | 9.2 |
| Pacific sand lance | 1 |  |  | 1 | 50 | 6 | 16.7 | 85.7 |
| Pacific sanddab | 13 | 5 | 229 | 8 | 395 | 33 | 39.4 | 10.0 |
| Pacific tomcod | 1 |  |  | 1 | 6 | 3 | 33.3 | 72.7 |
| Petrale sole | 64 | 62 | 381 | 2 | 5 | 67 | 95.5 | 96.7 |
| Pink salmon | 1 |  |  | 1 | 1 | 5 | 20.0 | 23.9 |
| Pygmy rockfish | 1 | 1 | 100 |  |  | 5 | 20.0 | 21.0 |
| Quillback rockfish | 5 | 4 | 30 | 1 | 30 | 6 | 83.3 | 57.6 |
| Redbanded rockfish | 121 | 121 | 942 |  |  | 129 | 93.8 | 81.1 |
| Redstripe rockfish | 17 | 11 | 485 | 6 | 330 | 47 | 36.2 | 7.7 |
| Rex sole | 87 | 23 | 934 | 64 | 2788 | 201 | 43.3 | 20.8 |
| Rock sole | 16 | 11 | 319 | 5 | 135 | 35 | 45.7 | 44.4 |
| Rosethorn rockfish | 28 | 27 | 329 | 1 | 18 | 36 | 77.8 | 75.3 |
| Rougheye rockfish | 51 | 51 | 383 |  |  | 78 | 65.4 | 36.1 |
| Roughtail skate | 4 |  |  | 4 | 4 | 4 | 100.0 | 100.0 |
| Sablefish | 24 | 17 | 435 | 7 | 134 | 135 | 17.8 | 44.2 |
| Sandpaper skate | 23 |  |  | 23 | 28 | 26 | 88.5 | 77.5 |
| Sharpchin rockfish | 16 | 8 | 279 | 8 | 348 | 43 | 37.2 | 20.7 |
| Shortraker rockfish | 7 | 7 | 25 |  |  | 8 | 87.5 | 93.7 |
| Shortspine thornyhead | 52 | 23 | 833 | 29 | 1129 | 106 | 49.1 | 35.5 |
| Silvergrey rockfish | 25 | 22 | 456 | 3 | 90 | 127 | 19.7 | 34.0 |
| Slender sole | 9 | 3 | 114 | 6 | 303 | 95 | 9.5 | 34.4 |
| Spiny dogfish | 11 |  |  | 11 | 430 | 126 | 8.7 | 24.6 |
| Splitnose rockfish | 11 | 9 | 319 | 2 | 83 | 28 | 39.3 | 2.9 |
| Spotted ratfish | 20 |  |  | 20 | 1044 | 151 | 13.2 | 19.0 |
| Threadfin sculpin | 1 |  |  | 1 | 6 | 21 | 4.8 | 9.4 |
| Walleye pollock | 19 | 7 | 155 | 12 | 500 | 92 | 20.7 | 43.4 |
| Widow rockfish | 5 | 5 | 75 |  |  | 13 | 38.5 | 52.7 |
| Yelloweye rockfish | 21 | 21 | 79 |  |  | 21 | 100.0 | 100.0 |
| Yellowmouth rockfish | 24 | 24 | 543 |  |  | 40 | 60.0 | 30.9 |
| Yellowtail rockfish | 10 | 8 | 231 | 2 | 45 | 32 | 31.3 | 27.7 |
| Total | 1372 | 753 | 14728 | 619 | 15367 |  |  |  |



Figure 1. Pictures of tire gear in the bosom of the footrope.


Figure 2. Proposed QCSd survey area. Yellow striped polygons represent sponge reef areas. Orange striped zones represent Hook-and-line rockfish closed areas. Catch weights for 38 commercially important species were obtained from the PacHarvTrawl database. These weights were apportioned based on trawl locations, onto a grid consisting of $4-\mathrm{km} 2$ blocks. The total weight in each block was summed and individual species weights were then converted to proportions. Cluster analysis was performed on the resulting dataset using the "clara" (clustering large applications) method (Kaufman and Rousseeuw, 1990).


Figure 3. One simulation of tow placement using allocation Options two and five.


Figure 4. Initial random selection of 240 primary (red blocks) and 60 secondary (yellow) blocks.


Figure 5. Actual surveyed blocks with rejected and aborted blocks.


Figure 6. Location of tows within blocks from a southeast section of the survey area.


Figure 7. Frequency distribution of observed catches by tow.


Figure 8. frequency distribution of number of species per tow


Figure 9. Plot of CV estimates for the combined aerial strata of the 2003 Queen Charlotte Sound trawl survey (with bias corrected $95 \%$ confidence intervals from 5,000 bootstrap replicates for 44 species (see Table 6 for definition of the acronyms).


Figure 10. Direct comparison of the analytical survey CVs (Eq. 4) with the CVs predicted based on commercial catch and effort using the method described by Schnute \& Haigh (2003). The species acronyms are provided in Table 6.


Figure 11. Comparison of the analytical survey CVs (Eq. 4) with the CVs predicted based on commercial catch and effort using the method described by Schnute \& Haigh (2003). Dashed line is the one-to-one line and the species plotting symbols are provided in Table 6.


Figure 12. Plot of CV's 27 species of fish captured in both the 2003 QCSd survey (CVa) and the 2001 NMFS survey of $f$ the Vancouver Region (CVb) for prorated to 239 tows. The solid line indicates the fitted regression while the dotted line indicates the 1:1 line.


Figure 13. Three simulated time series of surveys based on sampling CV from the QCSd survey and Process CV of 0.20 . Black line is the actual trend. Blue line is LOWESS fit. Top: Pacific ocean perch (QCSd-South), $\mathrm{K}=121$, $\mathrm{CV}=0.167$; actual abundance increasing $5 \% / \mathrm{y}$ for 20 y ; Mid: Rock sole (QCSd), K=239, CVs=0.257; actual abundance decreasing 15\%/y for 10y; Top: Canary rockfish (QCSd), K=249, CVs=0.369; actual abundance increasing $10 \% / \mathrm{y}$ for $15 y$.


Figure 14. Two simulated time series of surveys based on sampling CV from the QCSd survey and Process CV of 0.20 . Black line is the actual trend. Blue line is LOWESS fit. Top: Redstripe rockfish, $\mathrm{K}=239$, $\mathrm{CVs}=0.550$; actual abundance declining $10 \% / \mathrm{y}$ for $15 y$; Bottom Mid: Big Skate (QCSd), $\mathrm{K}=239, \mathrm{CVs}=0.690$; actual abundance stable for 20 y .


Figure 15. Catch per tow of Pacific ocean perch displayed over bottom bathymetry and interpolated bottom temperature.


Figure 16. Plot of biomass estimates for the combined aerial strata of the 2003 Queen Charlotte Sound trawl survey with bias corrected $95 \%$ confidence intervals from 5,000 bootstrap replicates for 44 species.

## Appendices

Table 9. Net specifications

| Part | Standard Length | Material | Metric Length | Units | Material |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Rigging |  |  |  |  |  |
| Sweep Line | 90 | 7/8 cable | 27.4 | m | 22 mm cable |
| Upper bridle | 90 | 3/4 cable | 27.4 | m | 19 mm cable |
| Lower bridle | 90 | 7/8 cable | 27.4 | m | 22 mm cable |
| Door Legs | 36 | 7/8 cable | 11.0 | m | 22 mm cable |
| Pickups | 42 | 7/8 cable | 12.8 | m | 22 mm cable |
| Hook ups | 8.8 t | BMMDV80 | 8 mt | mt | BMMDV80 |
| Net frame |  |  |  |  |  |
| Headline | 74.5 | 5/8 cable | 22.7 | m | 16 mm cable |
| Headline floats | 90 | 8" plastic Spheres |  |  | 200 mm plastic spheres |
| Riblines |  | 1" Polysteel rope |  |  | 25 mm polysteel rope |
| Bolsch Line | 68.33 | 9/8" poly steel rope | 20.8 | m | 29 mm polysteel rope |
| Fishing Line | 107.33 | 14 mm long link chain | 32.7 | m | 14 mm long link chain |
| Foot Rope |  |  |  |  |  |
| Foot Rope | 107.33 | 5/8 Chain <br> 16 in Tire gear with 2 in | 32.7 | m | 16 mm chain 400 mm tire gear with 50 mm |
| Foot rope bosom | 14 | Spacing <br> 18" rock hopper, 18 " | 4.3 | m | spacing 450 mm rock hopper 450 mm |
| Root rope wing1 | 18.33 | disks spaced 18 " apart 18" rock hopper, 18 " | 5.6 | m | spacing 450 mm rock hopper 450 mm |
| Root rope wing2 | 8.83 | disks spaced 18 " apart | 2.7 | m | spacing |
| Web |  |  |  |  |  |
| Belly | 5" | 3.5 mm Euroline | 127 | mm | 3.5 mm Euroline |
| Square | 5" | 3.5 mm Euroline | 127 | mm | 3.5 mm Euroline |
| Side Panel | 5" | 3.5 mm Euroline | 127 | mm | 3.5 mm Euroline |
| Taper | 4.5" | 3.5 mm Euroline | 114 | mm | 3.5 mm Euroline |
| Intermediate | 4.5 " | 3.5 mm Euroline | 114 | mm | 3.5 mm Euroline |
| Codend | 4.5 " | 3.5 mm Euroline | 114 | mm | 3.5 mm Euroline |
| Guard Mesh | 4.5 or 5 " | Double 4.5 mm Euroline | 114 or 127 | mm | Double 4.5 mm Euroline |
| Liner | 3/4" | Notless Nylon | 19 | mm | Notless Nylon |

Table 10. Retained, discarded, and total catch by fish species from 239 useable tows.

| Species | Retained Weight (kg) | Discarded <br> Weight (kg) | Total <br> Weight (kg) |
| :---: | :---: | :---: | :---: |
| Arrowtooth Flounder | 167.3 | 7,945.7 | 8,113.0 |
| Aurora Rockfish |  | 0.7 | 0.7 |
| Barracudinas |  | 0.0 | 0.0 |
| Big Skate |  | 369.6 | 369.6 |
| Bigfin Eelpout |  | 5.2 | 5.2 |
| Bigmouth Sculpin |  | 10.3 | 10.3 |
| Black Eelpout |  | 1.0 | 1.0 |
| Black Rockfish |  | 0.8 | 0.8 |
| Blackbelly Eelpout |  | 78.0 | 78.0 |
| Blackfin Poacher |  | 0.3 | 0.3 |
| Blackfin Sculpin |  | 15.5 | 15.5 |
| Blackgill Rockfish |  | 1.8 | 1.8 |
| Blacktail Snailfish |  | 0.1 | 0.1 |
| Blacktip Poacher |  | 0.0 | 0.0 |
| Blue Lanternfish |  | 0.0 | 0.0 |
| Bluespotted Poacher |  | 0.0 | 0.0 |
| Bocaccio | 13.6 | 143.1 | 156.7 |
| Brown Irish Lord |  | 3.7 | 3.7 |
| Brown Rockfish |  | 7.0 | 7.0 |
| Butter Sole |  | 2.2 | 2.2 |
| Canary Rockfish | 61.3 | 1,355.7 | 1,417.0 |
| Chilipepper |  | 3.2 | 3.2 |
| Chub Mackerel |  | 74.7 | 74.7 |
| Chum Salmon |  | 47.0 | 47.0 |
| Curlfin Sole |  | 8.9 | 8.9 |
| Darkblotched Rockfish | 46.8 | 167.6 | 214.4 |
| Daubed Shanny |  | 0.0 | 0.0 |
| Dover Sole | 338.5 | 2,426.9 | 2,765.4 |
| Dusky Rockfish |  | 2.0 | 2.0 |
| Eelpouts |  | 2.4 | 2.4 |
| English Sole | 14.6 | 925.6 | 940.2 |
| Eulachon |  | 46.8 | 46.8 |
| Fish Eggs |  | 0.0 | 0.0 |
| Flathead Sole | 0.5 | 718.0 | 718.5 |
| Goldfish |  | 2.3 | 2.3 |
| Greenstriped Rockfish | 1.0 | 150.3 | 151.3 |
| Hagfishes |  | 0.1 | 0.1 |
| Harlequin Rockfish |  | 2.4 | 2.4 |
| Inanimate Object(s) |  | 0.6 | 0.6 |
| Jacks |  | 1.7 | 1.7 |
| Kelp Greenling |  | 4.6 | 4.6 |
| Lampreys |  | 0.0 | 0.0 |
| Lanternfishes |  | 0.2 | 0.2 |
| Lingcod | 94.5 | 480.3 | 574.8 |
| Longnose Skate |  | 594.9 | 594.9 |
| Northern Lampfish |  | 1.2 | 1.2 |
| Northern Ronquil |  | 0.0 | 0.0 |
| Northern Spearnose Poacher |  | 0.0 | 0.0 |
| Pacific Cod | 77.2 | 797.7 | 874.9 |
| Pacific Hake |  | 1,772.9 | 1,772.9 |
| Pacific Halibut |  | 729.2 | 729.2 |
| Pacific Herring |  | 9.8 | 9.8 |
| Pacific Ocean Perch | 18,608.2 | 17,385.7 | 35,993.9 |
| Pacific Sand Lance |  | 0.7 | 0.7 |
| Pacific Sanddab |  | 839.7 | 839.7 |
| Pacific Staghorn Sculpin |  | 0.0 | 0.0 |
| Pacific Tomcod |  | 1.1 | 1.1 |
| Pacific Viperfish |  | 0.0 | 0.0 |
| Pearly Prickleback |  | 0.1 | 0.1 |
| Perches |  | 0.3 | 0.3 |
| Petrale Sole | 1.8 | 274.7 | 276.5 |
| Pink Salmon |  | 7.1 | 7.1 |


| Species | Retained <br> Weight (kg) | Discarded <br> Weight (kg) | Total <br> Weight (kg) |
| :---: | :---: | :---: | :---: |
| Poachers |  | 0.1 | 0.1 |
| Pricklebacks |  | 0.0 | 0.0 |
| Pygmy Poacher |  | 0.4 | 0.4 |
| Pygmy Rockfish |  | 16.2 | 16.2 |
| Quillback Rockfish |  | 63.1 | 63.1 |
| Ragfish |  | 0.0 | 0.0 |
| Redbanded Rockfish | 396.8 | 1,283.6 | 1,680.4 |
| Redstripe Rockfish | 2,919.5 | 481.5 | 3,401.0 |
| Rex Sole | 143.2 | 3,142.0 | 3,285.2 |
| Ribbed Sculpin |  | 0.0 | 0.0 |
| Ribbon Barracudina |  | 0.1 | 0.1 |
| Ridgeheads |  | 0.0 | 0.0 |
| Rock Sole | 39.1 | 468.2 | 507.3 |
| Rosethorn Rockfish | 21.4 | 113.3 | 134.7 |
| Roughback Sculpin |  | 0.0 | 0.0 |
| Rougheye Rockfish | 227.3 | 1,323.1 | 1,550.4 |
| Roughtail Skate |  | 25.0 | 25.0 |
| Sablefish |  | 1,975.0 | 1,975.0 |
| Sandpaper Skate |  | 39.8 | 39.8 |
| Sculpins |  | 0.6 | 0.6 |
| Sharpchin Rockfish | 210.5 | 931.8 | 1,142.3 |
| Shining Tubeshoulder |  | 0.0 | 0.0 |
| Shortbelly Rockfish |  | 13.3 | 13.3 |
| Shortfin Eelpout |  | 0.3 | 0.3 |
| Shortraker Rockfish | 42.3 | 81.6 | 123.9 |
| Shortspine Thornyhead | 312.3 | 1,400.4 | 1,712.7 |
| Silvergray Rockfish | 487.7 | 2,584.2 | 3,071.9 |
| Slender Sole |  | 100.6 | 100.6 |
| Slim Sculpin |  | 0.2 | 0.2 |
| Smalldisk Snailfish |  | 0.0 | 0.0 |
| Spiny Dogfish |  | 2,170.0 | 2,170.0 |
| Spinyhead Sculpin |  | 0.0 | 0.0 |
| Splitnose Rockfish | 344.1 | 4,374.0 | 4,718.1 |
| Spotfin Sculpin |  | 0.8 | 0.8 |
| Spotted Ratfish |  | 1,637.2 | 1,637.2 |
| Sturgeon Poacher |  | 0.3 | 0.3 |
| Threadfin Sculpin |  | 8.9 | 8.9 |
| Tubeshoulders |  | 0.0 | 0.0 |
| Viperfishes |  | 0.0 | 0.0 |
| Walleye Pollock | 5.9 | 319.2 | 325.1 |
| Wattled Eelpout |  | 0.4 | 0.4 |
| Widow Rockfish | 14.6 | 169.9 | 184.5 |
| Wolf Eel |  | 7.2 | 7.2 |
| Yelloweye Rockfish | 4.1 | 266.6 | 270.7 |
| Yellowmouth Rockfish | 169.6 | 2,321.1 | 2,490.7 |
| Yellowtail Rockfish | 689.2 | 565.5 | 1,254.7 |
| Subtotal: | 25,453 | 63,307 | 88,760 |
| Non-Finfish |  | 1,361.7 | 1,361.7 |
| Total: | 25,453 | 64,669 | 90,122 |

*Note: 0.0 indicate amounts of less than 0.05 kg

Table 11. Frequency of occurrence in useable tows for all species/groups by area and depth stratum.

| Species | Area Stratum 1: 5AB South |  |  |  | Area Stratum 2: 5AB North |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Depth Strata |  |  |  | Depth Strata |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 |  |
| Anemone | 3 |  | 2 | 2 |  | 1 |  |  | 8 |
| Anthozoa | 1 |  |  |  |  |  | 2 | 1 | 4 |
| Arminidae | 1 |  |  |  |  |  |  |  | 1 |
| Arrowtooth Flounder | 13 | 56 | 30 | 4 | 5 | 39 | 48 | 17 | 212 |
| Arthropoda | 1 |  |  |  |  |  |  |  | 1 |
| Ascidians And Tunicates | 3 | 8 | 2 |  |  | 3 | 7 | 1 | 24 |
| Aurora Rockfish |  |  |  |  |  |  | 1 |  | 1 |
| Barracudinas |  |  |  |  |  |  |  | 1 | 1 |
| Basket Stars | 5 | 7 | 1 |  |  | 3 | 3 | 1 | 20 |
| Big Skate | 11 |  |  |  | 1 |  |  |  | 12 |
| Bigfin Eelpout |  | 1 | 2 |  | 2 | 2 | 1 | 1 | 9 |
| Bigmouth Sculpin |  |  | 1 |  |  | 2 |  |  | 3 |
| Black Eelpout |  |  |  |  |  |  |  | 5 | 5 |
| Black Rockfish |  |  |  |  |  | 1 |  |  | 1 |
| Blackbelly Eelpout | 7 | 36 | 9 | 4 | 1 | 2 | 4 | 3 | 66 |
| Blackfin Poacher |  | 1 | 4 | 1 |  |  |  |  | 6 |
| Blackfin Sculpin |  | 1 | 5 |  |  | 4 | 15 | 1 | 26 |
| Blackgill Rockfish |  |  |  |  |  |  | 1 |  |  |
| Blacktail Snailfish |  |  | 1 |  |  |  | 1 | 1 | 3 |
| Blacktip Poacher |  |  | 1 |  |  |  | 1 |  | 2 |
| Blood Star |  |  | 1 |  |  |  |  |  | 1 |
| Blue Lanternfish |  |  | 2 |  |  |  | 1 |  | 3 |
| Bluespotted Poacher |  |  |  |  |  | 2 |  |  | 2 |
| Bocaccio |  | 4 | 2 |  |  | 5 | 1 |  | 12 |
| Box Crabs |  | 3 |  |  |  |  |  |  | 3 |
| Bristly Crab |  |  | 1 |  |  |  |  |  | 1 |
| Brittle Stars |  |  | 2 |  |  | 5 | 2 | 3 | 12 |
| Brown Irish Lord |  |  |  |  |  | 1 |  |  | 1 |
| Brown Rockfish |  |  |  |  |  | 1 | 2 |  | 3 |
| Butter Sole | 2 |  |  |  |  |  |  |  | 2 |
| Canary Rockfish | 6 | 12 | 4 |  | 1 | 13 | 1 |  | 37 |
| Cancer Branneri | 2 |  |  |  |  |  |  |  | 2 |
| Chilipepper |  | 1 |  |  |  | 1 |  |  | 2 |
| Chitons |  | 1 |  |  |  |  |  |  | 1 |
| Chub Mackerel | 1 |  |  |  |  |  |  |  | 1 |
| Chum Salmon |  | 1 |  |  |  | 1 | 1 |  | 3 |
| Cookie Star |  |  | 1 |  | 1 |  |  |  | 2 |
| Coonstripe Shrimp |  |  |  |  |  | 2 |  |  | 2 |
| Curlfin Sole | 9 |  |  |  |  | 2 |  |  | 11 |
| Cushion Star |  | 2 | 2 |  |  | 3 | 2 | 1 | 10 |
| Darkblotched Rockfish |  | 3 | 9 | 3 |  |  | 8 | 2 | 25 |
| Daubed Shanny | 1 |  |  |  |  |  |  |  | 1 |
| Decorator Crab | , | 2 | 1 |  |  | 2 | 1 |  | 7 |
| Dover Sole | 9 | 48 | 28 | 4 | 3 | 18 | 44 | 18 | 172 |
| Dusky Rockfish |  |  |  |  |  | 1 |  |  | 1 |
| Eelpouts |  | 4 |  | 1 |  | 1 | 1 | 1 | 8 |
| English Sole | 15 | 34 | 8 |  | 4 | 12 | 2 |  | 75 |
| Eulachon | 2 | 16 | 1 |  | 1 | 2 | 13 | 5 | 40 |
| Fish Eggs | 1 |  |  |  | 1 |  | 1 |  | 3 |
| Fish-eating Star |  |  |  |  |  | 2 |  |  | 2 |
| Flathead Sole | 5 | 46 | 9 |  | 3 | 6 | 13 |  | 82 |
| Fragile Urchin |  | 31 | 8 | 4 |  | 11 | 23 | 10 | 87 |
| Gastropods |  | 2 |  |  |  | 2 |  | 1 | 5 |
| Giant Pacific Octopus |  |  | 2 | 1 |  |  | 1 |  | 4 |
| Giant Red Sea Cucumber |  |  | 1 |  |  |  |  |  | 1 |
| Glass Shrimp |  |  |  |  |  |  |  | 1 | - |
| Glass Sponges | 1 |  |  |  |  |  |  |  | 1 |
| Goldfish |  |  |  |  |  |  | 1 |  | 1 |
| Gorgonian Corals |  |  |  |  |  | 1 | 4 | 2 | 7 |
| Green Urchin |  |  |  |  |  | 1 |  |  | 1 |



| Species | Area Stratum 1: 5AB South |  |  |  | Area Stratum 2: 5AB North |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Depth Strata |  |  | 4 | Depth Strata |  |  |  |  |
|  | 1 | 2 | 3 |  | 1 | 2 | 3 | 4 |  |
| Sablefish | 1 | 37 | 27 | 6 | 1 | 10 | 34 | 19 | 135 |
| Sand Star |  |  |  |  |  |  |  |  | 1 |
| Sandpaper Skate |  | 5 | 1 | 1 |  | 1 | 7 | 11 | 26 |
| Scallop | 1 |  |  |  |  | 2 |  |  | 3 |
| Sculpins |  | 1 |  |  |  |  | 1 |  | 2 |
| Sea Cucumber | 4 | 8 | 9 | 2 |  | 1 | 5 | 2 | 31 |
| Sea Lilies And Feather Stars |  |  |  |  |  |  |  | 1 | 1 |
| Sea Pen |  | 2 |  |  |  |  |  |  | 2 |
| Sea Pens | 1 |  |  |  |  |  |  |  | 1 |
| Sea Urchins | 1 | 6 | 1 |  | 1 | 6 | 1 |  | 16 |
| Sea Whip | 2 | 3 |  | 1 | 1 | 3 | 5 |  | 15 |
| Seaslugs |  |  |  |  |  | 1 |  | 1 | 2 |
| Sharpchin Rockfish | 1 | 7 | 6 |  |  | 12 | 14 | 3 | 43 |
| Shining Tubeshoulder |  |  |  |  |  |  | 1 |  | 1 |
| Shortbelly Rockfish |  | 1 |  |  |  | 1 |  |  | 2 |
| Shortfin Eelpout |  | 2 |  |  |  |  | 1 | 2 | 5 |
| Shortraker Rockfish |  |  | 1 | 2 |  |  | 2 | 3 | 8 |
| Shortspine Thornyhead |  | 6 | 28 | 6 |  |  | 47 | 19 | 106 |
| Shrimp |  |  |  |  |  |  | 1 | 1 | 2 |
| Sidestripe Shrimp |  | 13 | 7 |  | 1 | 4 | 33 | 2 | 60 |
| Silvergray Rockfish | 5 | 35 | 11 | 1 | 3 | 33 | 38 | 1 | 127 |
| Slender Sole | 4 | 35 | 17 |  | , | 10 | 27 | 1 | 95 |
| Slim Sculpin | 1 |  |  |  | 1 |  |  |  | 2 |
| Smalldisk Snailfish |  |  |  |  |  |  |  | 1 | 1 |
| Solasteridae | 1 |  |  |  |  |  |  |  | 1 |
| Spike Shrimp (horned Shrimp) |  |  |  |  | 1 |  |  |  | 1 |
| Spiny Dogfish | 12 | 37 | 9 |  | 5 | 34 | 22 | 7 | 126 |
| Spiny Red Sea Star |  |  | 2 |  |  | 2 | 1 |  | 5 |
| Spinyhead Sculpin |  | 1 |  |  |  |  |  |  | 1 |
| Splitnose Rockfish |  |  | 7 |  |  |  | 18 | 3 | 28 |
| Sponges | 3 | 9 | 7 | 4 | 2 | 14 | 16 | 5 | 60 |
| Spotfin Sculpin | 2 |  |  |  |  |  |  |  | 2 |
| Spotted Ratfish | 19 | 51 | 14 | 2 | 6 | 37 | 18 | 4 | 151 |
| Squat Lobster | 1 | 1 | 1 |  |  |  |  |  | 3 |
| Squids |  | 7 | 16 | 5 |  | 3 | 24 | 18 | 74 |
| Starfish | 2 | 5 | 1 | 1 | 1 | 6 | 4 | 3 | 23 |
| Sturgeon Poacher |  |  |  |  | 1 | 1 | 3 |  | 5 |
| Threadfin Sculpin | 7 | 5 | 1 |  | 1 | 6 |  | 1 | 21 |
| Tubeshoulders |  |  | 1 |  |  |  |  |  | 1 |
| Vermillion Starfish |  | 1 | 1 |  |  |  | 1 |  | 3 |
| Viperfishes |  |  |  |  |  |  | 1 | 1 | 2 |
| Walleye Pollock | 4 | 18 | 6 | 2 | 4 | 26 | 28 | 4 | 92 |
| Wattled Eelpout |  | 3 |  |  |  | 2 |  |  | 5 |
| Widow Rockfish | 1 | 4 | 2 | 1 |  | 1 | 4 |  | 13 |
| Wolf Eel | 2 |  | 1 |  |  |  |  |  | 3 |
| Yelloweye Rockfish | 1 | 7 |  |  | 1 | 12 |  |  | 21 |
| Yellowmouth Rockfish |  | 9 | 14 | 1 |  | 1 | 14 | 1 | 40 |
| Yellowtail Rockfish | 1 | 17 | 1 |  |  | 8 | 4 |  | 31 |
| Yelloweye Rockfish | 1 | 9 |  |  | 1 | 14 | 1 |  | 26 |
| Yellowmouth Rockfish |  | 12 | 14 | 1 |  | 1 | 17 | 1 | 46 |
| Yellowtail Rockfish | 1 | 21 | 1 |  |  | 8 | 4 |  | 35 |
| Total: | 370 | 1093 | 531 | 94 | 102 | 673 | 955 | 308 | 4126 |

Table 12. Analytic and bootstrap results for 44 species from the southern aerial stratum of the 2003 QCSd survey.

Analytic results are presented for the biomass (Eq. 1) and the CV (Eq. 4). Bootstrap results are for 5,000 replicate samples taken with replacement. Bias corrected $95 \%$ confidence intervals are presented and the bootstrap CV is calculated relative to the bootstrap mean biomass south

| Species | Biomass <br> (t) | $\begin{array}{r} \text { Bootstrap } \\ \text { mean } \\ \text { biomass }(\mathbf{t}) \\ \hline \end{array}$ | Bootstrap lower bound | Bootstrap upper bound | Bootstrap CV | Analytic CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pacific Ocean Perch | 17407.8 | 17409.4 | 12528.9 | 23822.3 | 0.1665 | 0.1649 |
| Yellowtail Rockfish | 355.9 | 355.0 | 87.3 | 914.3 | 0.5692 | 0.5718 |
| Yellowmouth Rockfish | 1194.3 | 1188.4 | 362.2 | 2347.1 | 0.4230 | 0.4305 |
| Arrowtooth Flounder | 4253.2 | 4260.2 | 3254.6 | 5480.1 | 0.1323 | 0.1343 |
| Silvergray Rockfish | 640.4 | 639.2 | 371.3 | 1040.9 | 0.2619 | 0.2672 |
| Dover Sole | 1533.5 | 1531.1 | 1162.3 | 1975.6 | 0.1359 | 0.1346 |
| Lingcod | 384.6 | 381.9 | 236.7 | 621.6 | 0.2499 | 0.2534 |
| Redstripe Rockfish | 1893.5 | 1909.3 | 212.2 | 5942.9 | 0.7743 | 0.7788 |
| Canary Rockfish | 337.8 | 336.6 | 121.6 | 714.0 | 0.4462 | 0.4414 |
| Rock Sole | 681.1 | 676.9 | 366.2 | 1089.4 | 0.2679 | 0.2685 |
| Pacific Cod | 352.8 | 353.3 | 195.2 | 582.6 | 0.2775 | 0.2769 |
| Petrale Sole | 105.6 | 105.4 | 63.0 | 165.8 | 0.2446 | 0.2445 |
| Redbanded Rockfish | 552.2 | 553.2 | 365.0 | 825.5 | 0.2080 | 0.2099 |
| Yelloweye Rockfish | 42.8 | 43.1 | 13.7 | 86.8 | 0.4322 | 0.4330 |
| Bocaccio | 30.6 | 30.7 | 9.4 | 60.6 | 0.4251 | 0.4356 |
| Sandpaper Skate | 11.6 | 11.5 | 1.4 | 29.6 | 0.6016 | 0.5994 |
| Big Skate | 618.2 | 621.9 | 91.7 | 1868.3 | 0.7123 | 0.7153 |
| Wolf Eel | 10.0 | 10.1 | 0.0 | 28.2 | 0.6890 | 0.6949 |
| Spiny Dogfish | 2012.2 | 2023.1 | 384.8 | 5012.1 | 0.5675 | 0.5669 |
| Sablefish | 662.9 | 663.9 | 442.7 | 969.8 | 0.2024 | 0.2059 |
| Greenstripe Rockfish | 93.1 | 93.4 | 43.3 | 180.0 | 0.3593 | 0.3584 |
| Rougheye Rockfish | 814.0 | 819.7 | 122.6 | 1760.8 | 0.5192 | 0.5155 |
| Shortraker Rockfish | 35.8 | 35.7 | 0.0 | 103.6 | 0.7159 | 0.7161 |
| Rex Sole | 1813.3 | 1811.2 | 1437.1 | 2264.2 | 0.1174 | 0.1200 |
| Spotted Ratfish | 419.0 | 419.9 | 263.4 | 685.7 | 0.2505 | 0.2499 |
| Shortspine Thornyhead | 512.5 | 513.3 | 378.4 | 686.3 | 0.1523 | 0.1510 |
| Pacific Hake | 1000.4 | 1001.2 | 706.7 | 1352.0 | 0.1623 | 0.1645 |
| Walleye Pollock | 58.4 | 59.0 | 29.3 | 97.3 | 0.2937 | 0.2990 |
| Longnose Skate | 237.0 | 237.5 | 166.5 | 316.3 | 0.1609 | 0.1614 |
| Flathead Sole | 438.4 | 437.5 | 291.3 | 658.3 | 0.2108 | 0.2084 |
| Slender Sole | 57.5 | 57.6 | 38.5 | 83.3 | 0.1944 | 0.1924 |
| English Sole | 774.9 | 773.3 | 469.5 | 1173.3 | 0.2290 | 0.2290 |
| Pacific Halibut | 594.7 | 599.4 | 297.2 | 1028.2 | 0.3113 | 0.3166 |
| Blackbelly Eelpout | 49.7 | 49.8 | 23.7 | 94.7 | 0.3520 | 0.3519 |
| Sharpchin Rockfish | 50.6 | 50.4 | 14.1 | 110.1 | 0.4753 | 0.4810 |
| Rosethorn Rockfish | 38.8 | 38.9 | 13.9 | 79.6 | 0.4236 | 0.4319 |
| Eulachon | 22.9 | 22.8 | 7.0 | 49.7 | 0.4624 | 0.4647 |
| Pacific Sanddab | 1155.4 | 1150.2 | 562.8 | 1946.0 | 0.3049 | 0.3030 |
| Splitnose Rockfish | 470.9 | 472.1 | 6.5 | 1257.9 | 0.6217 | 0.6115 |
| Darkblotched Rockfish | 96.4 | 96.4 | 19.7 | 270.7 | 0.6720 | 0.6851 |
| Blackfin Sculpin | 2.7 | 2.6 | 0.1 | 9.2 | 0.8565 | 0.8379 |
| Threadfin Sculpin | 7.0 | 7.0 | 2.4 | 14.9 | 0.4449 | 0.4470 |
| Widow Rockfish | 109.2 | 110.7 | 5.5 | 353.6 | 0.7945 | 0.8016 |
| Curlfin Sole | 13.7 | 13.7 | 5.8 | 23.9 | 0.3328 | 0.3299 |

Table 13. Analytic and bootstrap results for 44 species from the northern aerial stratum of the 2003 QCSd survey.

Analytic results are presented for the biomass (Eq. 1) and the CV (Eq. 4). Bootstrap results are for 5,000 replicate samples taken with replacement. Bias corrected $95 \%$ confidence intervals are presented and the bootstrap CV is calculated relative to the bootstrap mean biomass.

| Species | Biomass (t) | $\begin{array}{r} \hline \text { Bootstrap } \\ \text { mean } \\ \text { biomass }(\mathbf{t}) \\ \hline \end{array}$ | Bootstrap lower bound | Bootstrap upper bound | Bootstrap CV | Analytic $\mathrm{CV}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pacific Ocean Perch | 6751.1 | 6765.7 | 4544.9 | 9754.5 | 0.1934 | 0.1906 |
| Yellowtail Rockfish | 628.8 | 636.0 | 93.9 | 1657.8 | 0.6049 | 0.6066 |
| Yellowmouth Rockfish | 527.0 | 520.6 | 140.4 | 1313.0 | 0.5403 | 0.5448 |
| Arrowtooth Flounder | 1815.9 | 1819.2 | 1277.4 | 2771.2 | 0.1996 | 0.1964 |
| Silvergray Rockfish | 1966.1 | 1964.7 | 1325.4 | 2843.4 | 0.1937 | 0.1944 |
| Dover Sole | 309.0 | 308.9 | 235.3 | 407.8 | 0.1427 | 0.1409 |
| Lingcod | 550.9 | 543.7 | 104.2 | 1592.6 | 0.6847 | 0.6726 |
| Redstripe Rockfish | 940.3 | 943.1 | 274.6 | 2138.9 | 0.4914 | 0.4887 |
| Canary Rockfish | 998.4 | 994.1 | 300.9 | 2108.1 | 0.4556 | 0.4608 |
| Rock Sole | 57.8 | 57.1 | 3.3 | 169.6 | 0.7605 | 0.7644 |
| Pacific Cod | 525.8 | 526.0 | 334.5 | 781.4 | 0.2127 | 0.2132 |
| Petrale Sole | 209.6 | 209.8 | 110.4 | 338.4 | 0.2635 | 0.2626 |
| Redbanded Rockfish | 543.3 | 542.4 | 329.4 | 882.8 | 0.2527 | 0.2522 |
| Yelloweye Rockfish | 217.0 | 215.9 | 64.6 | 487.3 | 0.4808 | 0.4734 |
| Bocaccio | 105.1 | 105.5 | 7.7 | 339.9 | 0.8335 | 0.8372 |
| Sandpaper Skate | 12.6 | 12.6 | 7.4 | 19.2 | 0.2392 | 0.2461 |
| Big Skate | 26.4 | 26.0 | 0.0 | 105.7 | 1.0006 | 1.0000 |
| Wolf Eel | 0.0 | 0.0 | - | - | - | 0.0000 |
| Spiny Dogfish | 793.9 | 794.6 | 538.5 | 1139.0 | 0.1910 | 0.1900 |
| Sablefish | 486.1 | 485.8 | 380.4 | 598.7 | 0.1117 | 0.1090 |
| Greenstripe Rockfish | 30.1 | 30.2 | 16.8 | 48.6 | 0.2669 | 0.2675 |
| Rougheye Rockfish | 165.0 | 164.3 | 111.0 | 236.0 | 0.1923 | 0.1919 |
| Shortraker Rockfish | 35.4 | 35.1 | 7.4 | 82.4 | 0.5283 | 0.5259 |
| Rex Sole | 623.0 | 621.3 | 444.7 | 866.1 | 0.1737 | 0.1741 |
| Spotted Ratfish | 3204.9 | 3220.6 | 220.0 | 11817.8 | 0.9140 | 0.9020 |
| Shortspine Thornyhead | 521.6 | 521.6 | 378.0 | 700.3 | 0.1578 | 0.1551 |
| Pacific Hake | 187.3 | 187.5 | 110.7 | 311.3 | 0.2684 | 0.2715 |
| Walleye Pollock | 210.1 | 209.3 | 137.1 | 318.1 | 0.2159 | 0.2163 |
| Longnose Skate | 268.4 | 268.5 | 139.9 | 510.6 | 0.3555 | 0.3549 |
| Flathead Sole | 125.9 | 127.2 | 40.2 | 244.7 | 0.4123 | 0.4126 |
| Slender Sole | 29.4 | 29.0 | 6.8 | 82.5 | 0.6872 | 0.6801 |
| English Sole | 377.6 | 377.1 | 89.9 | 883.1 | 0.5357 | 0.5278 |
| Pacific Halibut | 258.1 | 257.1 | 119.1 | 436.0 | 0.3084 | 0.3059 |
| Blackbelly Eelpout | 1.8 | 1.8 | 0.6 | 3.9 | 0.4570 | 0.4475 |
| Sharpchin Rockfish | 703.3 | 701.4 | 230.4 | 1755.5 | 0.5224 | 0.5198 |
| Rosethorn Rockfish | 50.1 | 50.0 | 25.3 | 86.0 | 0.2982 | 0.3041 |
| Eulachon | 12.1 | 12.0 | 3.9 | 25.7 | 0.4438 | 0.4453 |
| Pacific Sanddab | 30.2 | 30.0 | 4.1 | 85.0 | 0.6712 | 0.6520 |
| Splitnose Rockfish | 2482.1 | 2500.3 | 84.3 | 8719.3 | 0.9152 | 0.9267 |
| Darkblotched Rockfish | 41.8 | 42.2 | 9.8 | 88.7 | 0.4721 | 0.4712 |
| Blackfin Sculpin | 7.1 | 7.1 | 4.0 | 10.9 | 0.2479 | 0.2474 |
| Threadfin Sculpin | 2.4 | 2.4 | 0.2 | 6.7 | 0.7068 | 0.7030 |
| Widow Rockfish | 73.0 | 74.6 | 2.8 | 232.0 | 0.8226 | 0.8236 |
| Curlfin Sole | 0.7 | 0.7 | 0.0 | 1.8 | 0.6948 | 0.7063 |

