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#### Abstract

Annual landings of red sea urchin (Strongylocentrotus franciscanus) increased rapidly in the early 1980s for the south coast of British Columbia (B.C.) and in the late 1980s for the north coast, but subsequently were reduced and stabilised by quotas. Coastwide landings were $4,815.4 \mathrm{t}$ valued at approximately $\$ 8.4 \mathrm{M}$ (Cdn.), with 110 licenses issued during the 2000/2001 fishing season. Bed areas were obtained by digitising locations on charts indicated in harvest logbooks during 1997-2000. Analyses of surveys, during 1994-2001, provided estimates of mean density and weights of red sea urchin biomass within beds. Natural mortality rates of red sea urchins were assumed, for fishery management purposes, to be between 0.075 and 0.15 . The minimum size limit was reduced from 100 mm to 90 mm test diameter (TD) for the 2000-2001 fishing season. Of the total $(\mathrm{N}=22,739)$ measured in a survey of red sea urchins in processing plants, harvested throughout the 2000-2001 fishing season, the majority ( $95.0 \%$ ) was in the 95140 mm TD size group and only a few ( $4.3 \%$ ) were $<95 \mathrm{~mm}$ TD. Biomass and quota options were estimated for a variety of size limits (e.g., $\geq 90, \geq 95, \geq 100,95-140 \mathrm{~mm}$ TD) for red sea urchins in B.C. Further surveys for red sea urchin density are required in areas of B.C. where surveys are $>5$ years old, and that have been heavily fished and or invaded by sea otters. Reliable biomass estimates and accurate estimates of bed areas, natural mortality and recruitment rates for red sea urchins in most areas of B.C. are required to manage this fishery effectively.


## RÉSUMÉ

En Colombie-Britannique, les débarquements annuels d'oursin rouge (Strongylocentrotus franciscanus) ont rapidement augmenté au début des années 1980 sur la côte sud et à la fin de cette décennie sur la côte nord, mais ont par la suite été réduits et stabilisés par le biais de quotas. Les 110 permis délivrés pour la saison de pêche 2000-2001 ont donné des débarquements à l'échelle du littoral de $4815,4 \mathrm{t}$, d'une valeur d'environ $8,4 \mathrm{M} \$ \mathrm{CAN}$. La superficie des gisements a été établie par numérisation sur des cartes des endroits indiqués dans les journaux debord couvrant la période 1997-2000. Les analyses des données de relevés effectués de 1994 à 2001 ont permis d'obtenir des estimations de la densité et du poids moyens des oursins pour chaque gisement. Pour fins de gestion des pêches, on a supposé que le taux de mortalité naturelle de l'espèce se situait entre 0,075 et 0,15 . Le dimaètre minimum du test des prises a été réduit pour la saison de pêche 20002001, pour passer de 100 mm à 90 mm . La plupart des 22739 oursins, soit $95,0 \%$, mesurés lors d'un relevé des usines de transformation de l'oursin rouge récolté tout au long de cette saison de pêche étaient dans la tranche de taille $95-140 \mathrm{~mm}$, seuls quelquesuns (4,3\%) ayant un diamètre du test inférieur à 95 mm . Les options de biomasse et de quotas ont été estimés d'après une gamme de diamètres du test (p. ex. $\mu 90 \mathrm{~mm}$, $\geq 95 \mathrm{~mm}, \geq 100 \mathrm{~mm}, 95-140 \mathrm{~mm}$ ). D'autres relevés visant à établir la densité sont requis dans les secteurs de la province où des relevés n'ont pas effectués depuis plus de cinq ans et qui ont été soit fortement exploités ou envahis par la loutre de mer. Des estimations fiables de la biomasse et des estimations précises de la superficie des gisements, de la mortalité naturelle et des taux de recrutement de l'oursin rouge sont requises pour la plupart des secteurs de la province afin de gérer cette pêche de façon efficace.

## INTRODUCTION

A commercial dive fishery for the red sea urchin (Strongylocentrotus franciscanus ) started during the 1970s in British Columbia (B.C.) (Table 1). Annual landings started to increase rapidly in the early 1980s for the south coast and the late 1980s for the north coast of B.C., but subsequently were reduced and stabilised by quotas (Table 1). Coastwide landings were $4,815.4 \mathrm{t}$ valued at approximately $\$ 8.4$ million (Cdn.) with 110 licenses issued during the 2000-2001 fishing season. The history of the management of this fishery is summarised in Campbell and Harbo (1991), Heizer et al. (1997), Rogers and Neifer (1999), Campbell et al. (1999a), and Rogers and Parker (2000). A number of papers review various aspects of red sea urchin biology (Bernard and Miller 1973; Mottet 1976; Breen 1980; Sloan et al. 1987; Tegner 1989; Campbell and Harbo 1991; Botsford et al. 1993, 1994; Lai and Bradbury 1998; Ebert 1998). Surveys to estimate standing stock of red sea urchins in B.C. during 1976-99 have been published (Breen et al. 1976, 1978; Adkins et al. 1981; Sloan et al. 1987; Jamieson et al. 1998a, b, c, d; Bureau et al. $2000 \mathrm{a}, \mathrm{b}, \mathrm{c}, \mathrm{d}$ ). Based on the results of some of these surveys Campbell (1998), and Campbell et al. (1999 a, b, 2000) provided quota estimates for the red sea urchin fishery in recent years. The density estimated from unpublished surveys conducted during 2000 and 2001 also are included in the present paper.

In this paper, the B.C. coast was divided into two main regions, the 'North Coast' and the 'South Coast'. In addition, fishery statistics in B.C. were subdivided into Pacific Fishery Management (PFM) areas and PFM subareas for management and economic purposes (detailed charts are not shown in this paper).

Managers have requested biomass estimates based on the most recent surveys and that quotas, where possible, be estimated by PFM subarea so they can be applied to the 2002/2003 red sea urchin fishery in B.C. Managers have also requested an evaluation of the size range of urchins harvested in the commercial fishery, since the legal minimum size was reduced to $\geq 90 \mathrm{~mm}$ test diameter (TD), during the 2000-2001 fishing season, from that of $\geq 100 \mathrm{~mm}$ TD in previous years. Improvements in Global Positioning Systems (GPS), the careful recording of fishing locations in log books, digitising data, GIS software and computer capacity have provided opportunities to accurately determine bed areas that were fished during 1997-2000. Consequently, more detailed density data have been required for the last two years in an attempt to match the fished bed areas for estimating red sea urchin biomass. Since previous surveys included transects randomly placed along a coastline, the implications of using only transect samples that occurred in fished bed areas needed to be determined.

The purpose of this paper was to: (1) summarize historical catch trends from the sales slips and harvest logbooks; (2) examine the size range of red sea urchins harvested in the 2000-2001 commercial fishery; (3) compare bed areas determined prior to 1997 with those determined during 1997-2000; (4) determine the effect of using only density data from transects matching the bed areas; (5) summarize the density surveys conducted to
date and estimate biomass of red sea urchins; (6) determine annual quotas based on several size limit options for the red sea urchin fishery in B.C.

## METHODS

## Catch and Effort

Catch and effort data were summarised from sales slips and from harvest logbooks that fishers completed during each day of fishing. Information from sales slips included total weight (pounds) and value (dollars) landed, vessel registration number (VRN), date and days fished. Information from the harvest logbooks included location of bed (with diagram), date, landed weight and minutes of diving. Prior to 1997, the harvest logbooks were not completed by all vessels so the data were used as a sample (CPUE, kilograms per hour) only where both total catch ( kg ) and effort (minutes) per region were reported per diver for each area per day. During 1997 and thereafter, validation was combined with the harvest logbook program so the harvest logbooks were completed by all vessels; subsequently the sales slips were not completed by all vessels. Total average annual CPUE were calculated from mean daily vessel CPUE values.

## Commercial Size Frequency

A survey to determine the sizes of commercially fished red sea urchins was conducted each month from 4 November, 2000, to 27 April, 2001, when most of the landings of red sea urchins occurred during the 2000-2001 fishery. The test diameters of red sea urchins were measured from 7 processing plants and 3 ports. A total of 50 urchins were randomly collected and measured from each tote or bag that was randomly chosen at the location (processing plant or port). A minimum number of 150 to a maximum number of 2000 urchins were measured from a plant or port per month, which would represent as diverse a number of fishing vessels and PFM areas as possible throughout B.C.

## Bed Areas

In the past, historic commercial bed areas of red sea urchins were indicated on charts or diagrams provided by fishers with their harvest logbooks throughout B.C. during 19821996. An inefficient two-step method was used to transcribe the data from the harvesters' charts to a master set of nautical charts from $0-9.1 \mathrm{~m}(0-5 \mathrm{fm})$ below chart datum, and subsequently digitised and areas estimated (based on sea surface area) using proprietary, raster-based software (Compugrid). Prior to 1997, the logbook program was independent from the validation program; compliance with requirements for timely submission of harvest logs and charts was less enforceable and consequently not all fished locations may have been recorded.

Based on the above concerns and those of Campbell et al. (1999 a, b) suggesting that the use of historic cumulative estimates of fishable red sea urchin bed areas may have
included locations that no longer were fished, we devised new procedures in an attempt to improve accuracy and efficiency of estimating bed areas on an annual basis.

The new methods used in this paper to estimate fished red sea urchin bed areas are as follows. Starting with the 1997 fishing season, harvest log data were more reliable, with harvest weights provided through the dockside validation program and submission of chart information facilitated by the third party service provider, hired by the Pacific Urchin Harvesters Association (PUHA), to administer the validation/harvest log program. Harvest location information was digitised, in the form of polygons, directly from the chart record, provided by the harvester to the electronic spatial file. Area measurements were generated using a Universal Transverse Mercator projection, zone 9, which was also consistent with previous analyses. The analysis utilised vector-based software Arcview 3.2. A "fishing event" approach was applied to location information provided for the 1997 through 1999-2000 fishing seasons. A fishing event was defined as a single record as reported in the harvest logbook, representing one dive by one commercial harvester in a single location. The fishing events were digitised in reference to the same land basemap as used in previous urchin bed area analyses. Subsequently, this fishing event record could be linked via a one-to-one relationship to an entry in the GIS spatial data set. The ability to make the one-to-one linkage of location information to harvest logbook information was a result of the combined validation/harvest log program.

A separate spatial file was developed for each fishing season to allow (a) isolation and review of harvest locations from within a single season, (b) comparison of fishing locations from one season to the next or (c) combination of several seasons of harvest locations to review trends over a longer time period. Procedures were implemented to ensure that, where fishing event polygons were wholly or partially coincident, the events were combined to eliminate double counting the area of the coincident portions. In this paper we calculated the total bed areas, restricted to a depth range of $0-9.1 \mathrm{~m}(0-5 \mathrm{fm})$ below chart datum, based on three fishing seasons (1997-2000) for each PFM subarea. Any fishing events falling in areas officially closed to the fishery were excluded from the bed area totals.

Data for bed areas fished in 2000-2001 were not complete so a full analysis was not included in this paper. However, we calculated and included bed areas from new PFM subareas where landings had been reported in logbooks for 2000-2001, but which had not reported landings during 1997-2000. Bed areas from most of PFM area 14, which had been closed to fishing for several years and reopened for the 2000-2001 fishing season, were estimated from a recent population survey (Bureau et al. 2000 b ) using only transects that had $\geq 1$ urchin per $\mathrm{m}^{2}$ (all sizes) to indicate fishable locations. Bed areas from PFM subareas $7.001 \mathrm{~N}, 7.002 \mathrm{~N}, 7.031 \mathrm{~S}$ were estimated from a red sea urchin survey conducted during 2001 (unpublished data) using only transects that had $\geq 1$ urchin per $\mathrm{m}^{2}$ (all sizes) to indicate fishable locations.

Estimation of these red urchin bed areas must be treated with caution since the beds were not measured empirically in the field, and the proportion of the suitable substrate types
(e.g., boulders or flat bedrock area more suitable than sand or mud) are unknown and may differ from one area to another.

## DEnsity and Biomass

Densities of red sea urchins were generally estimated within $1 \mathrm{~m}^{2}$ quadrats along randomly chosen transects. Details of survey methodology varied between surveys (Breen et al. 1976, 1978; Adkins et al. 1981; Sloan et al. 1987; Jamieson et al. 1998a, 1998b, 1998c, 1998d; Bureau et al. 2000 a, b, c, d). Test diameters (TD, in mm) of urchins were measured on all the surveys except that by Adkins et al. (1981). Density estimates from Adkins et al. (1981) could be biased since counts were made only at sites where there were more than 1 red sea urchin $/ \mathrm{m}^{2}$. Surveys of sea urchin density were also conducted during 1995 to 2001, using the methodology described by Jamieson and Schwarz (1998). The survey data prior to 1994 had a different survey methodology and could not be used in the analyses in this paper. We analysed the data from the 1994-2001 surveys using only those transects that occurred within the bed boundaries fished during 1997-2001.

We used similar methods of Campbell et al. (1999b) for estimating mean densities. The methods consider the numbers, sizes, weights and biomass of red sea urchins within each transect length and weights this according to the total transect lengths sampled within the urchin beds of each PFM subarea or area. N.B. (a) the number of quadrats sampled for urchin density was usually half the potential number of quadrats (i.e., alternate quadrats were sampled) along a transect, and (b) the number of quadrats sampled for urchin sizes could be lower or equal to the quadrats sampled for density along a transect. In some areas high urchin abundance made measuring each urchin logistically unfeasible.

The estimated mean density, $\bar{d}$ (number / $\mathrm{m}^{2}$ ) or biomass density ( $\mathrm{g} / \mathrm{m}^{2}$ ), of urchins across a number of transects surveyed in a PFM area or subarea was calculated as

$$
\begin{equation*}
\bar{d}=\frac{\sum_{t}\left(d_{t} * L_{t}\right)}{\sum_{t} L_{t}} \tag{1}
\end{equation*}
$$

The standard error of the mean density, $\mathrm{s}_{\mathrm{d}}$, was calculated as

$$
\begin{equation*}
s_{d}=\sqrt{1-\frac{n}{T}} * \sqrt{\frac{\sum_{t}\left(d_{t} * L_{t}-\bar{d} * L_{t}\right)^{2}}{n *(n-1) * \bar{L}^{2}}} \tag{2}
\end{equation*}
$$

where n is the number of transects, $\mathrm{d}_{\mathrm{t}}=\mathrm{N}_{\mathrm{t}} / \mathrm{S}_{\mathrm{t}}$ is the density at transect $\mathrm{t}, \mathrm{S}_{\mathrm{t}}$ is the number of quadrats surveyed for density estimates in transect $\mathrm{t}, N_{t}$ is the number of red sea urchins counted for density estimates in transect $t, L_{t}$ is the length of transect $t$ (or area in square metres since each transect was one metre wide), $\bar{L}=\frac{\sum_{t} L_{t}}{n}$ is the mean transect length (or mean area in square metres), and T is the total potential number of transects that could possibly be sampled in the beds. The expression $\sqrt{ }(1-n / T)$ was assumed to be approximately one, because the sample size n was substantially smaller than T , as only a small fraction of the potential number of transects in the beds were surveyed.

To estimate the mean densities (number $/ \mathrm{m}^{2}$ ) or biomass density $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ for a specific size group ( $J$ ) the value $\mathrm{d}_{\mathrm{t}}$ was substituted with densities $\left(P_{t, J}\right)$ or biomass $\left(B_{t, J}\right)$, in equations 1 and 2 .

The mean density $P_{t, J}$ (number/ $\mathrm{m}^{2}$ )of red sea urchins of size group $J$ in transect t was calculated as

$$
\begin{equation*}
P_{t, J=} \frac{\left(N_{t} * \frac{\sum_{i \in J} m_{t, i}}{M_{t}}\right)}{S_{t}} \tag{3}
\end{equation*}
$$

The mean biomass density $B_{t, J}\left(\mathrm{~g} / \mathrm{m}^{2}\right)$ of red sea urchins of size group $J$ in transect t was calculated as

$$
\begin{equation*}
B_{t, J}=\frac{\left(\frac{N_{t}}{M_{t}} * \sum_{i \in J}\left(m_{t, i} * w_{i}\right)\right)}{S_{t}} \tag{4}
\end{equation*}
$$

where $J$ is a subset of possible i values representing a range of test diameters $\left(\mathrm{TD}_{\mathrm{i}}\right)$ (e.g. $\geq$ $90, \geq 95, \geq 100,90-130,90-140,95-130,90-140$ or 100-140, mm TD $), M_{t}=\sum_{i} m_{t, i}$ the total number of red sea urchins measured for size in transect $\mathrm{t}, m_{\mathrm{t}, \mathrm{i}}$ is the number of red sea urchins in size range $J$ of transect t , and $w_{i}=0.0012659 * \mathrm{TD}_{\mathrm{i}}{ }^{2.7068}$ is a relationship between mean wet weight ( g ) and size i (TD in mm) for red sea urchins (Campbell 1998).

Where there were no mean density estimates for a PFM subarea, an overall mean density for all years surveyed within the whole PFMA was used. Where there were no density data for a whole PFMA then mean values of data from the nearest or adjacent PFMA
were used. The accuracy of these extrapolations is unknown and will require further comparative field surveys.

Standard errors were not calculated for densities that had < 5 transects. However, because an approximate lower $90 \%$ confidence interval of mean biomass density (L90CIB) was required for each PFM subarea, the following method was used. The standard errors were assumed to be proportional to the means and sample size. Thus, a linear regression, $\mathrm{s}_{\mathrm{d}}=\mathrm{b}^{*} \bar{d} / /_{\mathrm{s}}$, was used to describe the relation between the standard error $\left(\mathrm{s}_{\mathrm{d}}\right)$ and the mean biomass density $(\bar{d})$ divided by the square root of the total transect length (s) used for the biomass estimates, with b being a constant estimated using the least squares method. Data analysed from each red sea urchin size limit class group in all PFM subareas and years $(\mathrm{N}=25)$ resulted in $b$ values of $6.56\left(\mathrm{R}^{2}=0.89\right)$ for $\geq 90 \mathrm{~mm}$ $\mathrm{TD} ; 6.65\left(\mathrm{R}^{2}=0.89\right)$ for $\geq 95 \mathrm{~mm} \mathrm{TD} ; 6.92\left(\mathrm{R}^{2}=0.90\right)$ for $\geq 100 \mathrm{~mm} \mathrm{TD} ; 5.88\left(\mathrm{R}^{2}=\right.$ $0.87)$ for $90-130 \mathrm{~mm} \mathrm{TD} ; 6.48\left(\mathrm{R}^{2}=0.87\right)$ for $90-140 \mathrm{~mm} \mathrm{TD} ; 5.90\left(\mathrm{R}^{2}=0.88\right)$ for 95 -130 mm TD; $6.46\left(\mathrm{R}^{2}=0.87\right)$ for $95-140 \mathrm{~mm} \mathrm{TD} ; 6.74\left(\mathrm{R}^{2}=0.87\right)$ for $100-140 \mathrm{~mm} \mathrm{TD}$. To calculate the approximate lower $90 \%$ confidence interval (L90CIB) of the mean biomass density $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ for each size limit class of red sea urchin and PFM subarea, we assumed $\mathrm{L} 90 \mathrm{CIB}=\bar{d}-1.64^{*} \mathrm{~s}_{\mathrm{d}}$, and by substituting $\mathrm{s}_{\mathrm{d}}$ (with the formula above), the equation $\mathrm{L} 90 \mathrm{CIB}=\bar{d} *(1.0-1.64 * \mathrm{~b} / \sqrt{ } \mathrm{s})$ was used to calculate the approximate lower $90 \%$ confidence interval, where 1.64 is the z -value corresponding to the 0.05 quantile. Negative values of L90CIB for PFM subareas were assigned zeros.

Total current biomass of red sea urchins, for each size group, for each PFM subarea was calculated as

$$
\begin{equation*}
\mathrm{B}_{\mathrm{c}}=\mathrm{A}^{*} \bar{d} \tag{5}
\end{equation*}
$$

where $\mathrm{B}_{\mathrm{c}}$ is the current total biomass $(\mathrm{g})$; $A$ is the commercial urchin bed areas (presented as hectares, but converted to $\mathrm{m}^{2}$ for biomass calculations) estimated from digitised charts and summed for each PFM subarea; $\bar{d}$ is the estimated mean biomass density $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ of red sea urchins in a size group (e.g., $\geq 90 \mathrm{~mm}$ or $95-130 \mathrm{~mm} \mathrm{TD}$ ). The values for $B_{c}$ were subsequently converted to tonnes for presentation. Since many of the surveys were conducted >4 years ago and there could be considerable uncertainty in the mean biomass density estimates in some areas we have adopted the Woodby (1992) method by also including the approximate lower $90 \%$ confidence interval (L90CIB) as a reasonable alternative for the $\bar{d}$ values in the $B_{c}$ calculations.

## Natural Mortality

There are no published estimates of instantaneous natural mortality rate (M) for red sea urchins from northern B.C. Breen (1984) estimated that M ranged from 0.016 to 0.22 for red urchins from 3 sites in southern B.C. and considered a value between 0.1-0.2 to be acceptable. Woodby (1992) estimated $M=0.16$ for red sea urchins from the Sitka, Alaska area. Botsford et al. (1993) estimated $\mathrm{M}=0.14$ for a population of red sea urchins in

California. Lai and Bradbury (1998) estimated M to be about 0.16 for red sea urchins from Washington. Based on published values Campbell (1998) assumed M to be 0.15 in calculating quotas for the 1995 red sea urchin fishery in B.C. However, all these authors considered growth rates of red sea urchins to be faster (e.g., 4-6 years to reach 100 mm TD) than that reported by Ebert (1998) who found tagged sea urchins from Washington and Oregon to take about 10 years to reach 100 mm TD and 50 years to reach 140 mm TD . Ebert (1998) calculated the mean instantaneous total mortality rate ( Z year ${ }^{-1}$ ) of red sea urchins, from a total of twelve samples collected from six locations in Oregon and Washington, to be 0.052 (min. 0.016, max 0.133 , lower $95 \%$ confidence interval (CI) 0.028 , upper CI 0.076); equivalent to a mean annual survival rate of $0.949\left(\mathrm{e}^{-\mathrm{Z}}\right)$. The average mortality values reported by Ebert (1998) were generally below those previously reported in the literature. Clearly M will vary between areas and between size classes for red sea urchins in B.C. Although a similar tagging program on red sea urchins in some areas of B.C. has been conducted, further experimental work is required to estimate growth of urchins < 2 years ( $5-30 \mathrm{~mm}$ TD). Initial analyses of the growth and mortality data indicated Z (instantaneous natural mortality rate) varied between sites from 0.03 to 0.16 (Campbell et al. unpublished data). For the purposes of the present paper, a range of M values from 0.075 to 0.150 (with an approximate mean $M$ of 0.10 ) was considered for red sea urchins in B.C.

## RECRUITMENT

Sloan et al. (1987) estimated recruitment to be highly variable between areas and to average about $9.5 \%$ of the total number of sea urchins in the size frequencies per area. Little is known about the stock and recruitment relations of red sea urchins in B.C. (Campbell et al. 1999a).

## Quota Estimation

A conservative management approach is used to estimate quotas $(\mathrm{Q})$ for the red sea urchin fishery in B.C. A modified surplus production model is used to estimate a maximum sustainable yield (MSY) from a stock that is in the early stages of fishing (Schaefer 1954; Gulland 1971). The model assumes that the MSY occurs when the maximum sustainable fishing mortality is equal to M .

$$
\begin{equation*}
\mathrm{Q}=\mathrm{X} * \mathrm{M} * \mathrm{~B}_{\mathrm{c}} \tag{6}
\end{equation*}
$$

where $\mathrm{B}_{\mathrm{c}}$ is the current biomass, M is the instantaneous natural mortality rate and $X=\mathrm{a}$ correction factor to insure that a sustainable fishing mortality rate is well below the calculated MSY. We chose the value of $X=0.20$ in this paper as a reasonably conservative safeguard to account for errors in estimating the lower current biomass values (Caddy 1986; Garcia et al. 1989). The correction factor should provide for a conservative harvest per year in a developing fishery where little is known about the productivity of the population. Since equation 6 is derived from a Graham-Schaefer production model, recruitment is assumed to be unaltered by these low fishing levels.

Although this approximation was developed for an unexploited virgin stock $\left(B_{o}\right)$ we assumed that $\mathrm{B}_{\mathrm{c}}=\mathrm{B}_{0}$.

Caution is required in the interpretation of these calculations for the quota because there are so many assumptions in the parameters used in the oversimplified model. Since many of the surveys were conducted $>4$ years ago and there could be uncertainty in the mean biomass density estimates in some areas, we have adopted the Woodby (1992) method by also including the approximate lower $90 \%$ confidence intervals as an alternative for the mean $B_{c}$ calculations.

## RESULTS AND DISCUSSION

## Catch and Effort

The number of licences issued peaked at 240 in 1989 and subsequently declined and stabilised at 110 during 1997-2001 (Table 1). The number of fishing vessels reporting landings peaked at 102 in 1992 and subsequently declined to 52 during 2000-2001 (Table 1). This recent decline is due to more licences being used (stacked) by fewer vessels. Coastwide landings peaked in 1992 (Tables 1, 2). Seasonal changes in landings for different PFM areas in B.C. are shown in Table 3. Quotas have generally restricted landings in the South Coast since 1985 and in the North Coast since 1993 (Table 1). There was no overall general trend in annual CPUE (kilograms per diver hour) for the red sea urchin fishery in B.C. between 1983-96 (Table 1). The general lack in CPUE trends suggests that either the fishery is at an early stage of development or CPUE data for red sea urchins may not be used to indicate fishery trends in B.C. Fishermen have increased search time for high quality urchins in response to recent changes in market demands and the implementation of an individual quota scheme. Also, fishermen may be maintaining high CPUE values, through serial depletion, by moving to unexploited sea urchin beds within a PFM area suggesting that CPUE would not decline until most legal-sized sea urchins were removed from most of the areas in the PFM area. There is a need to reexamine the distribution of effort and variability of CPUE data on a smaller spatial scale (e.g., by bed) than the PFM subarea level to determine whether CPUE is an appropriate index of red sea urchin abundance. Although updating the data base on matching bed locations and landings for the fishery during 1997-2000 has been completed, additional years of data probably are required to examine spatial changes in CPUE over many years.

## Commercial Size Frequencies

The summary statistics of test diameters (in mm TD) of commercially fished red sea urchins ( $\mathrm{N}=22,739$ ), measured during the 2000-2001 fishery, were mean $=111.4$, median $=110$, minimum $=40$, maximum $=196($ Fig 1A). A small percentage $(0.85 \%)$ of urchins was $<90 \mathrm{~mm}$ TD, $3.49 \%$ in the $90-94 \mathrm{~mm}$ TD size group, whereas most urchins ( $94.14 \%$ ) occurred for the size groups $90-130 \mathrm{~mm} \mathrm{TD}, 98.52 \%$ for $90-140 \mathrm{~mm}$

TD, $96.48 \%$ for $\geq 95 \mathrm{~mm} \mathrm{TD}, 90.62 \%$ for $95-130 \mathrm{~mm} \mathrm{TD}, 95.00 \%$ for $95-140 \mathrm{~mm}$ TD, $87.38 \%$ for $\geq 100 \mathrm{~mm}$ TD, $85.90 \%$ for $100-140 \mathrm{~mm} \mathrm{TD}, 1.70 \%$ for $\geq 140 \mathrm{~mm}$ TD.

In contrast, the summary statistics, in mm TD, of red sea urchins ( $\mathrm{N}=66,951$ ) measured in all population surveys during 1994-2001, were mean $=85.6$, median $=90.0$, minimum $=2$, maximum $=195($ Fig 1B $)$. Examining only legal-sized urchins, $\geq 90 \mathrm{~mm} \mathrm{TD}(\mathrm{N}=$ $34,098), 12.23 \%$ was in the $90-94 \mathrm{~mm}$ TD size group, $84.84 \%$ for $90-130 \mathrm{~mm}$ TD, $93.51 \%$ for $90-140 \mathrm{~mm} \mathrm{TD}, 87.77 \%$ for $\geq 95 \mathrm{~mm} \mathrm{TD}, 72.61 \%$ for $95-130 \mathrm{~mm} \mathrm{TD}$, $81.28 \%$ for $95-140 \mathrm{~mm} \mathrm{TD}, 75.21 \%$ for $\geq 100 \mathrm{~mm} \mathrm{TD}, 68.72 \%$ for $100-140 \mathrm{~mm} \mathrm{TD}$, $7.61 \%$ for $\geq 140 \mathrm{~mm}$ TD.

Clearly, fishermen targeted the majority ( $95.00 \%$ ) of red sea urchins within the 95-140 mm TD group during 2000-2001 compared to $81.28 \%$ of this size group measured in the total wild legal urchins $\geq 90 \mathrm{~mm}$ TD (Fig. 1, 2). Although median size was about 110 mm TD for both sampling methods of legal urchins $\geq 90 \mathrm{~mm} \mathrm{TD}$ (Fig. 2), fishers tended to select urchins > 95 mm TD and < 130 mm TD compared to what was probably available in the wild. Fishermen were optimising the quality and quantity of urchin gonads within the quota available. Also many fishermen tended to avoid urchins < 100 mm TD (only $12.62 \%$ of commercial samples) to deliberately reduce the chances of being involved with a lower than minimum legal size enforcement problem.

Caution is advised in interpreting these size frequencies in several respects. The size frequencies were grouped together from many areas in B.C. without weighting them according to total wild population abundance or total landings in a fishing season. An initial examination of weighting by landings indicated there was little difference between weighted and unweighted size frequencies of the commercial samples. The size frequencies of wild populations were obtained over a seven year period and some samples obtained > 3 years ago may not reflect accurately the relative abundance of sizes $>90 \mathrm{~mm}$ TD in some fished areas.

Future annual surveys to monitor the sizes of commercially fished red sea urchins should be continued so that a time series of size frequencies can be maintained and examined for changes in sizes within PFM areas and between years.

## Bed Areas

The total estimated bed area was $29,161.8$ ha for the North Coast and $6,459.1$ ha for the South Coast as of 1997-2000, which are $58.08 \%$ and $66.80 \%$, respectively, of the 19821996 cumulative estimates (Table 4). No consistent changes in estimating bed areas between the new and old methods were observed; there were major increases, decreases, totally new beds identified or no beds fished in some PFM subareas.

The exact reasons for these changes are unknown, but may be due to reduced and or increased fishing events between the two periods. Many of the beds identified in the early days of the fishery may not have been actual beds but areas where fishermen
erroneously suspected sustainable urchin numbers could have been present. Some beds may have been abandoned over the course of the fishery due to the possibility of serial depletion. There have been several PFM sub areas that were closed based on fishermen's requests for closure to allow local red sea urchin populations to recover. Some areas may be avoided by fishermen due to the presence of sea otters (major predators of red sea urchins) which have expanded their range in recent years, especially on the west coast of Vancouver Island (PFM areas 25, 26, 27) and some locations in the North Coast (PFM area 7). Some beds may no longer be commercially viable because sea otters have depleted the red sea urchin stocks.

Bed area estimates probably provide the most uncertainty of all the estimates used to calculate red sea urchin biomass. Estimating bed areas was crude, especially as each location may have different substrate surface areas. In the past, the harvest logbooks provided an historical cumulative estimate of fishable sea urchin areas, but may have included areas that no longer had viable red sea urchin populations. Recent logbook entries on locations fished are becoming accurate, due to GPS and the validation process. However, fishing activity in one season may not indicate the extent of some beds. The most recent bed area estimates for the 2000-2001 fishing season have yet to be analysed; there also are a few harvest logs showing fished beds that have to be received from some fishermen and digitised. Another uncertainty is the number of most recent years (e.g., rolling accumulation of the last 3-5 years) that are required to be grouped together to provide a reasonable indication of currently fished beds.

In some areas there may be large numbers of red sea urchins deeper than 9.1 m (D. Bureau, unpublished data; W. Bradbury, Washington State Fisheries and Wildlife, personal communication). Estimating bed areas in the reasonably shallow depth range ( 0 -9.1 m ) probably provides conservative bed area values to the estimation of red sea urchin biomass. There may be areas still unexplored, especially in the North Coast, that may contain unfished "virgin" populations that have not been included in the bed area estimates.

The new method will continue to be improved. The most recent complete season of data were captured in a conservative manner using the depth range information described in the harvest log records and available bathymetric contour data. To address the often small scale and crude nature of the charts provided by commercial harvesters, we recommend that the range of fishing depths reported in the harvest logs continue to be used as a guide when digitizing fishing locations. This has been facilitated through the recent availability of more detailed bathymetric base data, first used for fishing location data capture of the 2000-2001 fishing season data. For future analyses it may be beneficial to reference the older seasons of data to the newer, more detailed land and bathymetric base data. This may remove the need to clip the fishing location data using a restrictive depth strata or allow for analysis of fishing events at varying depths.

The records of log book bed area entries should continue to be made on an annual basis. Clear identification of beds in relation to the amount of red sea urchins removed from
each bed needs to be recorded carefully by fishers and on-ground monitor to allow detailed stock analyses on a bed by bed and/or PFM subarea basis in both the North and South Coasts. Although there has been an on-ground monitor in the North Coast, to date there have not been any on-ground monitor for the South Coast.

## DENSITY AND BIOMASS

In general, there was considerable variation in the size structure and densities of local red sea urchin populations. Estimated mean densities (number $/ \mathrm{m}^{2}$ ) and mean biomass density $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ for each size group of red sea urchins varied between PFM subareas and years (Tables 5, 6). Generally, mean density and biomass density were highest for the widest size limit, $\geq 90 \mathrm{~mm}$ TD, and density lowest for 100-140 mm TD, biomass density lowest for the narrowest size limit $95-130 \mathrm{~mm}$ TD of the size groups examined for each PFM areas (Tables 7).

We examined the implications of using only the "clipped" (including only transects lying within bed boundaries) compared to the unclipped (all transects included inside and outside beds) data to determine red sea urchin densities for each PFM subarea. Means, standard errors and $90 \%$ confidence intervals for biomass densities were calculated for both the clipped and unclipped data. A t-test comparing clipped and unclipped mean densities indicated that there were no significant ( $\mathrm{p}<0.05$ ) differences for each PFM subarea where there was survey data. The results were similar between each size group examined separately. There were more PFM areas ( $64.52 \%$ ) that had higher or the same mean densities for the clipped data compared to the unclipped data (e.g., Fig. 3A). Clipping had the result of generally increasing the mean densities, reducing the number of sample sizes or transects, therefore generally increasing the standard error of the mean densities (Fig. 3B), and consequently generally lowering the lower $90 \%$ confidence bound intervals (Fig. 3C).

As the fishery progresses, the average density and mean weight of the size group of urchins being exploited may decrease. Further surveys for red sea urchin density are required in areas of B.C. where surveys are > 5 years old, that have been heavily fished and or invaded by sea otters.

## Biomass and Quota

Red sea urchin biomass $\left(B_{c}\right)$ differed considerably between PFM subareas (Table 8). Total biomass of red sea urchins in the $90-140 \mathrm{~mm}$ TD size group was estimated at 203,221 t for the North Coast and 29,026 t for the South Coast (Table 8). These biomass and quota estimates must be treated with caution, especially when considering how inaccurate the bed estimates of viable red sea urchin populations may be.

If the size limit of $\geq 90 \mathrm{~mm}$ TD is to remain the same as the 2000-2001 fishing season, and considering fishermen tend to select for better gonad quality, the more appropriate quota estimate should be made from $B_{c}$ calculated from the $90-140 \mathrm{~mm}$ TD size group
(Table 8, 9). Choice of a conservative quota probably should be based on M values $\leq$ 0.10 (fishing rates of $\leq 2 \%$ ) which would suggest that the overall B.C. coastwide $2002 / 2003$ quota could be $4,679.4 \mathrm{t}$, based on the mean densities, or $2,876.5 \mathrm{t}$, based on the approximate lower $90 \%$ confidence interval estimate (Table 9). Alternatively, if fishermen wish to increase the minimum size to $\geq 95 \mathrm{~mm}$ TD, a conservative quota of 4350.6 t could be based on the $95-140 \mathrm{~mm}$ TD size group or the approximate lower $90 \%$ confidence interval estimate of 2262.3 t (Table9).

Although we examined two alternative formulations, proposed by Garcia et al (1989) (i.e., equations 7a and 8a, based on the Schaefer (1954) and Fox (1970) production models ), to estimate Q and MSY by including current yield $\left(\mathrm{Y}_{\mathrm{c}}\right)$ in an exploited population, both formulae were unstable when attempting to obtain MSY estimates below $Y_{c}$. Garcia et al (1989) also indicated that these formulae are unstable under other conditions. Die and Caddy (1997) question whether any simple approximation method alone, in obtaining sustainable yield indicators from biomass estimates, can provide a safe yield target. They advocated use of low conservative targets for fishing mortality and several biological reference points (BRPs) as a precautionary approach. Lai and Bradbury (1998), through simulation of red sea urchin populations in Washington, suggested that target harvest rates should be well below biological references points such as $\mathrm{F}_{\max }=0.48$ (fishing mortality at which $\mathrm{Y} / \mathrm{R}$ is maximised) and $\mathrm{F}_{0.1}=0.19$ (at which slope of $\mathrm{Y} / \mathrm{R}$ curve is $10 \%$ of the slope at origin) calculated from a yield per recruit model. Campbell et al. 1999b (Table 3 in appendix B), using a yield per recruit analysis, found $\mathrm{F}_{0.1}=0.11$ and 0.16 for a size limit of $\geq 90 \mathrm{~mm}$ TD and $90-140 \mathrm{~mm} \mathrm{TD}$, respectively.

If stock assessment and management is needed on a bed by bed and up to date basis, landings and bed areas will be required in a timely way (within one year). A two (or longer) year periodic rotation of fishing grounds would provide for easier monitoring of the fishing fleet and landings and allow timely analysis of up to date landings data. Three or six year rotation of fishing grounds would also allow recovery of the harvestable stock through recruitment and growth. Botsford et al. (1993) and Lai and Bradbury (1998) consider periodic harvest schedules (rotation), although not increasing cumulative yield, to be biologically beneficial, reduce variability of yield, risk, and probably management and enforcement costs.

The influence of different size limits and fishing mortality rates were reported by Campbell et al. (1999 see Appendices A \& B) for a red sea urchin population in B.C. The general results of these analyses are summarised as follows. Minimum size limits of $\geq 70$ and $\geq 90 \mathrm{~mm}$ TD are clearly less precautionary than that of $\geq 100 \mathrm{~mm}$ TD. Adding a lower maximum size to a minimum size reduces yield but increases the reproductive potential and consequently reduces the potential for collapse of red sea urchin populations. Low fishing mortality rates (e.g., $\leq 0.02$ ) are precautionary and may reduce the influence of size limits. However, having both minimum and maximum size limits and a low fishing mortality provide for a precautionary approach to exploiting red sea urchin populations when determining quotas from uncertain natural mortality and biomass estimations for particular PFM areas.

The commercial quality of red sea urchin roe is based on a combination of criteria such as color, texture, size and taste. Generally, good quality roe from red sea urchins is considered to have a yellow/gold consistent color, firm texture (growing or premature gonads), sweet taste and of medium size which is mainly found in the approximately 90 120 mm TD size range. Poor quality roe is generally considered to have variable yellow/brown color, soft texture and or oozing gametes (from mature, post spawn or spent gonads). Red sea urchins $>130 \mathrm{~mm}$ TD generally have large gonads and inconsistent color quality which are less desirable by the industry than smaller mature individuals. There is an annual reproductive cycle with the timing of the spawning season varying within the period from February to September depending on local environmental conditions such as food availability and temperature (Bernard 1977). Gonads increase in size usually from September to January (Kramer and Nordin 1975). The effect of food quality and availability to support growth and reproduction are important limiting factors on urchin stocks. Although red sea urchins are omnivorous grazers, kelps, such as Nereocystis leutkeana, provide optimal growth and gonad quality (Vadas 1977; Bureau et al. 1997; Morris and Campbell 1996). In areas of low supply or quality of food, individuals may relocate nutrients causing poor quality gonads thereby reducing reproduction potential of local red sea urchin populations. The influence of adult age, especially in large old red sea urchins, on low egg production, senescence or of poor quality and low survival of larval offspring is unknown.

Breen (1984) argued that a size limit could be used to protect sufficient reproductive potential in attempt to insure that recruitment did not fall below replacement. A special argument for a maximum size limit can be made for red sea urchins as adult urchins may provide a spine cover for juveniles that is necessary for good survival (Tegner and Dayton 1977). Minimum and maximum size limits would leave enough large adults to protect the settlement of juveniles, such as used in Washington State (Lai and Bradbury 1998). Breen (1984) suggested that using a size limit as the only control mechanism to prevent recruitment overfishing would not be achieved with the $\geq 100 \mathrm{~mm}$ TD size limit. A larger size limit would be required to prevent recruitment overfishing but would be larger than the industry's upper limit for market quality. Breen (1983) recommended that a better way of preventing recruitment overfishing was "controlling effort or catch in such a way as to protect local stocks from over harvesting".

The influence of market demand for high quality gonads may cause fishermen to concentrate fishing in a small size range (e.g., $95-130 \mathrm{~mm} \mathrm{TD}$ ) of red sea urchins even though the quota may be based on a legal size limit of $\geq 90 \mathrm{~mm}$ TD or $95-140 \mathrm{~mm}$ TD. This would result in a higher fishing mortality on the small size range than was originally intended by the managers. Clearly, setting a quota on the appropriate size range that will be harvested in an area will ensure less deviation from the intended fishing mortality on local red sea urchin populations. Continued monitoring of harvested red sea urchins at ports and or at commercial processing plants will provide the means to track temporal changes and area differences in size frequency and provide a tool to assess the appropriate size range to set quotas.

There are a number of complex density dependent compensatory and depensatory mechanisms that red sea urchins may elicit in growth, mortality, increased survival of juveniles due to protection by adult spine canopies, spawning success, roe quality and egg viability. A stock at low density levels may impact on the population negatively, e.g. "the Allee affect" (Allee 1931), reducing the reproductive success of the animals. This is particularly important in organisms that are broadcast spawners. Levitan et al. (1992) found that fertilization success was a function of the number of red sea urchins, distance apart, position in the cluster, flow direction and velocity of the current, etc. For animals that show this type of fertilization, this is a factor that should be considered in the management system in providing various forms of harvest refugia (e.g. size restrictions, catch limits through quotas or rotating spatial harvests and spatial closures) (Botsford et al. 1993; Quinn et al. 1993; Pfister and Bradbury 1996; Levitan and Sewell 1998). Ebert (1998) found potentially little Allee effects on growth of red sea urchins in Oregon and Washington using simulation techniques. He recommended not to automatically assume Allee effects are important in local areas without some evaluation, such as incorporating sensitivity analyses in dynamic modelling methods.

In addition to resource monitoring through fishery independent surveys, additional biological information is clearly required on age estimation and on the variation of growth, mortality and recruitment rates for production modelling of red sea urchin populations in different areas of B.C.

## RECOMMENDATIONS

1. If the size limit of $\geq 90 \mathrm{~mm} \mathrm{TD}$ is to be maintained, we recommend that a precautionary approach could include the inclusion of a maximum size of 140mm TD and a low fishing mortality of $\leq 0.02(\mathrm{M} \leq 0.10)$ when calculating quotas.
2. Continue improving the process of estimating bed areas holding viable populations of red sea urchins. Bed areas fished on an annual basis need to be recorded, validated, digitized and updated.
3. Continue monitoring harvested red sea urchins at ports and or at commercial processing plants which will provide temporal changes and area differences in size frequency, mean weights, and gonad quality of commercial-sized individuals, and provide a tool to assess the appropriate size range to set quotas in the future.
4. Surveys for the abundance estimates of red sea urchins in alternative areas in B.C. should be conducted on an annual basis. Areas with no surveys for over 5 years and areas that have been heavily fished should be considered a priority.
5. Further research is required to understand age, and growth, mortality and recruitment rates of red sea urchins in B.C. to assist with production modeling.

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## FIGURE LEGENDS

Fig. 1. Size frequencies of red sea urchins measured (A) at processing plants during the 2000-2001 fishing season, and (B) on population surveys during 1994-2001 throughout B.C.

Fig. 2. Cumulative size frequencies, expressed as a proportion of the total red sea urchins ( $\geq 90 \mathrm{~mm} \mathrm{TD}$ ) sampled from the processing plants ( $\mathrm{N}=22,546$ ) during the 2000-2001 fishing season, and from population surveys during $(\mathrm{N}=34,098)$ during 1994-2001 throughout B.C.

Fig. 3. Comparison of (A) means, (B) standard error of mean, and (C) lower $90 \%$ confidence bounds of $95-130 \mathrm{~mm}$ TD red sea urchin density (biomass, $\mathrm{g} / \mathrm{m}^{2}$ ) from unclipped (all transects) and clipped (only transects lying within bed boundaries) survey data by PFM subarea in B.C.

Table 1. Annual red sea urchin landings (tonnes), value and effort throughout British Columbia, during 1978-2001, as reported on fish slips and harvest logs. N.B. Annual landings and effort during 1978-96 are on a calendar basis (January to December), but thereafter are based on a fishing season (July to June).

| Year | $\begin{array}{\|l} \text { Licences } \\ \text { Issued } \end{array}$ | Vessels <br> with <br> Landings | South <br> Coast <br> Quota ${ }^{\text {a }}$ <br> (t) | North <br> Coast <br> Quota <br> (t) | HARVEST LOG DATA |  |  | FISH SLIP DATA ${ }^{\text {d }}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Coastwide <br> Landings (t) | Total Diver Hours | Mean <br> CPUE <br> (kg/diver $\mathrm{hr})^{\mathrm{c}}$ | Coastwide <br> Landings <br> (t) |  | Landed <br> Value <br> (\$000) |  | \$/ Kg | Total Vessel Fishing Days | Mean CPUE (t/vessel day) |
| 1978 | C | 4 |  |  | - | - | - | 75.0 |  | 16.0 |  | 0.21 | 54 | 1.4 |
| 1979 | C | 29 |  |  | - | - | - | 317.0 |  | 76.0 |  | 0.24 | 298 | 1.1 |
| 1980 | C | 18 |  |  | - | - | - | 333.0 |  | 84.0 |  | 0.25 | 331 | 1.0 |
| 1981 | C | 18 | 136 |  | - | - | - | 116.0 |  | 34.0 |  | 0.29 | 127 | 0.9 |
| 1982 | C | 21 |  |  | 45.4 | 75.6 | - | 160.0 |  | 56.0 |  | 0.35 | 195 | 0.8 |
| 1983 | Z 64 | 36 |  |  | 720.2 | 1,428.4 | 504 | 982.0 |  | 357.8 |  | 0.36 | 757 | 1.3 |
| 1984 | Z 85 | 47 |  |  | 1,377.0 | 3,781.6 | 364 | 1,764.4 |  | 712.2 |  | 0.40 | 1,058 | 1.7 |
| 1985 | Z 86 | 46 | 1,803 |  | 1,204.4 | 2,881.6 | 418 | 1,815.4 |  | 763.5 |  | 0.42 | 1,126 | 1.6 |
| 1986 | Z 103 | 67 | 1,500 |  | 1,582.0 | 3,397.0 | 466 | 2,066.6 |  | 1,010.8 |  | 0.49 | 1,534 | 1.3 |
| 1987 | Z 184 | 97 | 1,633 |  | 1,435.6 | 3,429.4 | 419 | 2,223.4 |  | 1,275.6 |  | 0.57 | 1,737 | 1.3 |
| 1988 | Z 184 | 84 | 1,678 |  | 1,763.8 | 5,056.7 | 349 | 2,115.4 |  | 1,238.2 |  | 0.59 | 1,239 | 1.7 |
| 1989 | Z 240 | 98 | 1,644 |  | 2,004.8 | 5,409.4 | 371 | 2,658.1 |  | 1,631.5 |  | 0.61 | 1,542 | 1.7 |
| 1990 | Z 188 | 88 | 1,668 |  | 2,439.7 | 7,478.7 | 326 | 3,158.1 |  | 1,953.0 |  | 0.62 | 2,651 | 1.2 |
| 1991 | Z 102 | 76 | 1,531 |  | 6,427.4 | 16,356.0 | 393 | 6,945.2 |  | 4,187.1 |  | 0.60 | 3,862 | 1.8 |
| 1992 | Z 108 | 102 | 1,554 |  | 12,479.9 | 31,170.0 | 400 | 12,981.8 |  | 8,661.5 |  | 0.67 | 6,222 | 2.1 |
| 1993 | Z 107 | 95 | 1,401 | 5,443 | 6,106.4 | 17,201.5 | 355 | 6,388.0 |  | 5,372.7 |  | 0.84 | 3,364 | 1.9 |
| 1994 | Z 110 | 95 | 1,543 | 5,897 | 5,959.8 | 18,942.0 | 315 | 5,828.7 |  | 8,066.4 |  | 1.38 | 3,978 | 1.5 |
| 1995 | Z 108 | 88 | 1,387 | 5,455 | 6,806.9 | 21,397.7 | 318 | 6,584.7 |  | 11,349.8 |  | 1.72 | 4,167 | 1.6 |
| 1996 | Z 109 | 77 | 1,265 | 5,360 | 6,466.4 | 18,180.3 | 356 | 5,282.5 |  | 10,045.8 |  | 1.90 | 3,536 | 1.5 |
| 1997/98 b | Z 110 | 90 | 1,702 | 8,150 | 8,738.2 | 30,227.4 | 289 | 8,450.8 | d | 14,477.1 | e | 1.66 | 5,161 | 1.6 |
| 1998/99 | Z 110 | 68 | 968 | 4,635 | 5,182.9 | 16,554.4 | 313 | 5,076.7 | d | 8,029.5 | e | 1.55 | 3,239 | 1.6 |
| 1999/00 | Z 110 | 61 | 967.5 | 4,634 | 5,282.6 | 16,664.8 | 317 | 4,896.1 | d | 8,450.4 | e | 1.60 | 3,027 | 1.6 |
| 2000/01 | Z 110 | 52 | 843.9 | 4,042 | 4,815.4 | 14,041.0 | 343 | 3,781.8 | d | 8,361.0 | e | 1.74 | 2,113 | 1.8 |

${ }^{\text {a }}$ South coast quota includes exploratory areas; North Coast quota new in 1993.
${ }^{\mathrm{b}}$ Change in licencing from calender year to market-driven year. 1997/98 season ran January 1/97 to June 30/98.

$$
\text { After } 1997 \text { licencing year continued to be July to June }
$$

${ }^{\text {c }}$ CPUE from harvest log data.
${ }^{\text {d }}$ Fish Slip Data preliminary for 1996-2001, landings data not validated since 1997.
${ }^{e}$ Approximate landed value estimated by multiplying Fish Slip $\$ / \mathrm{kg}$ by landings from the harvest log data.

Table 2a. Red sea urchin annual landings (tonnes) by Pacific Fishery Management areas in the North Coast, during 1984-2001, as reported on fish slips and harvest logs. 1994-2001 catch data corrected to use validation \& harvest logs only.

| North Coast PFM Areas |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $1^{\text {a }}$ | 2E | $2 \mathrm{~W}^{\text {b }}$ | 3 | 4 | $5^{\text {c }}$ | $6^{\text {d }}$ | 7 | 8 | 9 | 10 | Total Landings |
| 1984 | 2.2 |  |  |  |  |  |  |  |  |  |  | 2.2 |
| 1986 |  |  |  |  |  |  |  |  |  |  | 12.0 | 12.0 |
| 1987 |  |  |  |  | 23.0 |  |  | 179.0 | 91.0 |  |  | 293.0 |
| 1988 |  |  |  |  | 73.0 | 11.0 | 7.3 | 314.0 | 32.0 |  |  | 437.3 |
| 1989 | 0.2 | 223.0 |  | 1.6 | 116.0 | 1.3 | 168.0 | 217.0 | 65.0 |  | 180.0 | 972.1 |
| 1990 |  | 26.6 | 10.7 | 24.5 | 156.8 | 265.3 | 67.1 | 1,040.1 |  |  |  | 1,591.1 |
| 1991 |  | 333.1 | 2.7 | 143.3 | 1,026.7 | 2,577.3 | 77.7 | 774.7 | 114.6 | 24.5 | 304.9 | 5,379.5 |
| 1992 |  | 1,111.0 |  | 1.0 |  | 3,294.0 | 4,063.0 | 2,763.0 | 140.0 | 114.0 | 38.0 | 11,524.0 |
| 1993 | 97.0 | 189.0 | 88.9 | 127.2 | 1,008.0 | 463.0 | 2,103.0 | 1,012.0 | 43.4 |  | 215.3 | 5,346.8 |
| 1994 | 221.0 | 402.2 | 167.4 | 173.0 | 687.0 | 1,056.0 | 1,244.0 | 861.0 | 57.0 | 46.0 | 164.0 | 5,078.6 |
| 1995 | 258.0 | 440.2 | 256.3 | 48.0 | 940.0 | 1,280.0 | 1,053.0 | 1,076.0 | 111.0 | 49.0 | 224.0 | 5,735.5 |
| 1996 | 259.0 | 365.0 | 241.8 | 66.0 | 851.0 | 1,156.0 | 1,213.0 | 833.0 | 122.0 | 10.0 | 248.0 | 5,364.8 |
| 1997/1998 ${ }^{\text {e }}$ | 582.0 | 718.7 | 311.1 | 62.0 | 1,076.0 | 1,107.0 | 2,175.0 | 870.0 | 112.0 | 41.0 | 152.0 | 7,206.8 |
| 1998/1999 | 276.2 | 294.0 | 144.1 | 38.8 | 595.1 | 741.8 | 1,373.6 | 526.2 | 87.6 | 36.8 | 191.4 | 4,305.6 |
| 1999/2000 | 296.1 | 288.9 | 222.3 | 32.4 | 602.6 | 746.8 | 1430.3 | 462.6 | 88.1 | 38.6 | 189.5 | 4,398.2 |
| 2000/2001 | 269.5 | 273.8 | 197.3 | 136.8 | 400.8 | 690.4 | 1385.6 | 361.9 | 78.3 | 44.2 | 155.2 | 3,993.7 |
| 1984 to |  |  |  |  |  |  |  |  |  |  |  |  |
| June 2001 | 2,261.2 | 4,665.5 | 1,642.6 | 854.5 | 7,556.0 | 13,390.0 | 16,360.5 | 11,290.5 | 1,141.9 | 404.0 | 2,074.3 | 61,641.2 |

${ }^{a}$ Includes landings from Area 101
${ }^{\text {b }}$ Includes landings from Area 142
${ }^{\text {c }}$ Includes landings from Area 105
${ }^{\text {d }}$ Includes landings from Area 106
${ }^{\mathrm{e}}$ Change in licencing from calender year to market-driven year. 1997/98 season ran January 1/97 to June 30/98. After 1997 licencing year continued to be July to June

Table 2b. Red sea urchin annual landings (tonnes) by Pacific Fishery Management areas in the South Coast, during 1971-2001, as reported on fish slips and harvest logs. 1994 to 2001 catch data corrected to use Validation \& Harvest logs only.

| South Coast PFM Areas |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | East Coast Vancouver Island |  |  |  |  |  |  |  |  |  |  | West Coast Vancouver Island |  |  |  |  |  |  | Total Annual Landings |
| Year | $11^{\text {d }}$ | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 28 | 29 | 20 | 21 | $23^{\text {e }}$ | $24^{\text {f }}$ | 25 | 26 | 27 |  |
| 1971 to $1973{ }^{\text {a }}$ |  |  |  |  |  |  |  |  | 110.0 |  |  |  |  |  | 254.0 |  |  |  | 364.0 |
| 1974 to $1977^{\text {a }}$ |  | 1.4 |  | * | * | 1.4 |  |  | 66.0 |  |  | * |  | 1.3 |  |  |  |  | 70.1 |
| 1978 |  |  | * |  |  |  |  |  | 46.0 |  |  | 29.0 |  |  |  |  |  |  | 75.0 |
| 1979 |  |  | * | 78.0 |  |  | 57.0 | 133.0 | 45.0 |  |  | 1.8 | 0.9 | 2.5 |  |  |  |  | 318.2 |
| 1980 |  |  |  | 18.0 |  |  | 162.0 | 54.0 | 97.0 |  |  | 1.8 |  |  |  |  |  |  | 332.8 |
| 1981 |  |  | 20.0 | 4.0 | * |  | 5.3 | 47.0 | 22.0 |  |  |  |  |  | 17.0 |  |  |  | 115.3 |
| 1982 |  | 2.5 |  | 46.0 |  |  | 0.8 | 11.0 | 94.0 |  |  |  |  |  | 5.0 |  |  |  | 158.5 |
| $1983{ }^{\text {b }}$ | 7.8 | 99.0 | 264.0 | 260.0 | * | * | 59.0 | 38.0 | 112.0 |  |  | 24.0 |  | 22.0 | 38.0 |  | 62.0 |  | 985.8 |
| 1984 | 0.3 | 437.0 | 777.3 | 172.0 |  |  | 33.0 | 67.4 | 76.3 |  | 5.7 | 69.1 |  | 17.3 | 103.0 |  | 3.9 |  | 1,762.3 |
| 1985 |  | 354.0 | 492.0 | 167.0 | 106.0 | 5.9 | 29.0 | 48.0 | 77.0 |  | 47.0 | 30.0 |  | 96.0 | 158.0 | 145.0 | 15.0 | 45.0 | 1,814.9 |
| 1986 | 27.0 | 548.0 | 376.0 | 178.0 | 56.0 | 4.4 | 57.0 | 129.0 | 105.0 |  | 2.0 | 40.0 |  | 154.0 | 285.0 |  | 2.5 | 91.0 | 2,054.9 |
| $1987{ }^{\text {c }}$ | 6.9 | 420.0 | 491.0 | 193.0 | 32.4 |  | 71.0 | 71.0 | 123.0 | 17.0 | 7.8 | 17.0 |  | 63.0 | 199.0 | 95.0 | 8.3 | 12.0 | 1,827.4 |
| 1988 | 2.6 | 534.0 | 480.0 | 78.0 | 21.0 | 2.3 | * | 22.0 | 78.0 |  |  | 74.0 |  | 13.0 | 250.0 | 66.0 |  | 58.0 | 1,678.9 |
| 1989 |  | 569.0 | 493.0 | 122.0 | 6.7 |  | 9.0 | 64.0 | 57.0 |  | 1.6 | 15.0 |  |  | 223.0 | 39.0 |  | 86.0 | 1,685.3 |
| 1990 | 84.8 | 437.6 | 428.4 | 56.6 | 1.2 | 0.6 | 43.0 | 46.5 | 58.6 | 0.3 | 1.8 | 7.9 |  | 59.7 | 215.1 | 56.8 |  | 68.1 | 1,567.0 |
| 1991 | 36.4 | 358.7 | 370.7 |  | 8.6 |  | 26.6 | 94.8 | 27.2 |  | 14.1 | 31.2 | 2.7 | 58.4 | 185.1 | 115.8 |  | 121.1 | 1,451.4 |
| 1992 | 8.0 | 531.0 | 320.0 |  |  |  | 103.0 | 36.0 | 86.0 |  | 4.0 | 56.0 | 9.0 | 31.0 | 200.0 | 10.0 |  | 65.0 | 1,459.0 |
| 1993 | 55.5 | 329.0 | 184.0 |  |  |  | 21.0 | 104.7 | 17.3 |  |  | 14.5 |  | 40.4 | 92.0 | 7.0 | 2.0 | 50.0 | 917.4 |
| 1994 | 17.0 | 348.0 | 168.0 |  |  |  | 4.0 | 59.0 | 14.0 |  | 1.0 | 3.0 |  | 54.0 | 111.0 | 50.0 |  | 49.0 | 878.0 |
| 1995 | 34.0 | 364.0 | 175.0 |  |  |  | 28.0 | 69.0 | 15.0 |  | 9.0 | 20.0 |  | 57.0 | 199.0 |  |  | 98.0 | 1,068.0 |
| 1996 | 38.0 | 344.0 | 238.0 |  |  |  | 25.0 | 112.0 | 7.0 |  | 10.0 | 33.0 |  | 46.0 | 122.0 |  |  | 70.0 | 1,045.0 |
| 1997/1998 ${ }^{\text {g }}$ | 61.0 | 594.0 | 426.0 |  | 2.3 |  | 26.0 | 67.0 | 40.0 |  | 16.0 | 32.0 |  | 54.0 | 132.0 |  |  | 85.0 | 1,535.3 |
| 1998/1999 | 49.7 | 285.2 | 196.4 |  |  |  | 26.9 | 85.3 | 21.1 |  | 2.9 | 22.5 |  | 31.5 | 107.6 |  |  | 48.2 | 877.3 |
| 1999/2000 | 81.1 | 270.4 | 204.3 |  |  |  | 26.6 | 63.0 | 42.7 |  | 4.3 | 22.6 |  | 29.1 | 113.1 |  |  | 27.3 | 884.4 |
| 2000/2001 | 57.3 | 287.3 | 178.7 | 23.0 |  |  | 22.7 | 22.4 | 32.1 |  | 3.3 | 22.0 |  | 39.3 | 99.1 |  |  | 34.6 | 821.7 |
| 1971 to |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| June 2001 | 567.4 | 7,114.1 | 6,282.8 | 1,395.6 | 234.2 | 14.6 | 835.1 | 1,444.1 | 1,469.3 | 17.3 | 130.5 | 566.3 | 12.6 | 869.5 | 3,108.0 | 584.6 | 93.7 | 1,008.3 | 25,747.9 |

[^0]Table 3a. Monthly landings (tonnes) by Pacific Fishery Management areas in the North Coast, during 20002001, as reported on validation and harvest logs.

| Month- <br> Year | North Coast PFM Areas |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Queen Charlotte Islands |  |  | North Coast Mainland |  |  |  |  |  |  |  | Monthly Totals |
|  | $1^{\text {a }}$ | 2E | $2 W^{\text {b }}$ | 3 | 4 | $5^{\text {c }}$ | $6{ }^{\text {d }}$ | 7 | 8 | 9 | 10 |  |
| Jul-00 |  |  |  |  |  |  |  |  |  |  |  | 0.0 |
| Aug-00 |  |  |  |  |  |  |  |  |  | 17.0 | 20.0 | 37.0 |
| Sep-00 |  |  |  |  |  | 80.2 | 138.3 |  |  |  | 9.7 | 228.2 |
| Oct-00 |  |  |  | 16.2 | 252.4 | 297.9 |  |  |  |  |  | 566.5 |
| Nov-00 |  |  |  | 34.5 | 90.7 | 267.9 | 365.1 |  |  |  |  | 758.2 |
| Dec-00 |  |  |  |  |  |  | 656.8 | 220.4 |  |  |  | 877.2 |
| Jan-01 |  |  |  |  |  |  | 141.7 | 77.4 | * | 27.1 | 82.6 | 328.8 |
| Feb-01 |  | 273.8 | 37.9 | 86.1 | 57.6 | 23.5 |  |  |  |  |  | 478.9 |
| Mar-01 | 269.5 |  | 39.9 |  |  |  |  |  |  |  |  | 309.4 |
| Apr-01 |  |  | 119.5 |  |  | 20.9 | 83.7 | 51.1 |  |  | 37.9 | 313.1 |
| May-01 |  |  |  |  |  |  |  | 13.0 | * |  | 5.1 | 18.1 |
| Jun-01 |  |  |  |  |  |  |  |  |  |  |  | 0.0 |
| Total: | 269.5 | 273.8 | 197.3 | 136.8 | 400.8 | 690.4 | 1385.6 | 361.9 | 78.3 | 44.2 | 155.2 | 3993.7 |

* Data not provided for reasons of confidentiality where less than 3 vessels report landings.

PFM area totals not affected. Monthly totals reflect only reported values.
${ }^{\text {a }}$ Includes landings from Area 101.
${ }^{0}$ Includes landings from Area 142.
${ }^{\text {c }}$ Includes landings from Area 105.
${ }^{a}$ Includes landings from Area 106.

Table 3b. Monthly landings (tonnes) by Pacific Fishery Management areas in the South Coast, during 2000-2001, as reported on validation and harvest logs.

| Month- <br> Year | South Coast PFM Areas |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | East Coast Vancouver Island |  |  |  |  |  |  |  | West Coast Vancouver Island |  |  |  | Monthly Totals |
|  | $11^{\text {a }}$ | 12 | 13 | 14 | 17 | 18 | 19 | 29 | 20 | $23^{\text {b }}$ | $24^{\text {c }}$ | 27 |  |
| Jul-00 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 |
| Aug-00 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 |
| Sep-00 |  |  |  |  |  |  |  |  |  | 39.3 | 99.1 |  | 138.4 |
| Oct-00 |  | 22.2 | 72.6 | 23.0 | 11.9 |  |  |  |  |  |  | 12.5 | 142.1 |
| Nov-00 |  | 106.3 |  |  | * |  |  | 3.3 |  |  |  | , | 109.6 |
| Dec-00 | * | 72.5 | 106.1 |  |  | 22.4 | 32.1 |  | 22.0 |  |  |  | 255.0 |
| Jan-01 | 20.4 | 43.3 |  |  | * |  |  |  |  |  |  |  | 63.7 |
| Feb-01 | 9.7 | * |  |  |  |  |  |  |  |  |  |  | 9.7 |
| Mar-01 | 14.2 | * |  |  |  |  |  |  |  |  |  |  | 14.2 |
| Apr-01 | 7.7 | * |  |  |  |  |  |  |  |  |  |  | 7.7 |
| May-01 | * | * |  |  |  |  |  |  |  |  |  | * | 0.0 |
| Jun-01 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 |
| Total: | 57.3 | 287.3 | 178.7 | 23.0 | 22.7 | 22.4 | 32.1 | 3.3 | 22.0 | 39.3 | 99.1 | 34.6 | 821.7 |

* Data not provided for reasons of confidentiality where less than 3 vessels report landings.

PFM area totals not affected. Monthly totals reflect only reported values.
a Includes landings from Area 111.
${ }^{0}$ Includes landings from Area 123.
c Includes landings from Area 124.

Table 4. Bed areas (ha) for fished red sea urchins by PFM subarea according to digitized data from fishery logs for period 1982-96 and 1997-2000. NB : a = estimated from 2000/2001 harvest logbooks; $\mathrm{s}=$ estimated from recent population surveys (Bureau et al. 2000a, b); $\mathrm{t}=$ new beds areas estimated from a population survey conducted in 2001 (unpublished data).

| PFM <br> Subarea | Bed Areas (ha) during |  | NB | $\begin{gathered} \hline \text { PFM } \\ \text { Subarea } \end{gathered}$ | Bed Areas (ha) during |  | NB | PFM <br> Subarea | Bed Areas (ha) during |  | NB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1982-96 | 1997-2000 |  |  | 1982-96 | 1997-2000 |  |  | 1982-96 | 1997-2000 |  |
| North Coast of BC |  |  |  |  |  |  |  |  |  |  |  |
| 1.001 | 4837.56 | 967.75 |  | 3.004 | 24.36 |  |  | 7.025 | 796.04 | 167.56 |  |
| 1.002 | 405.00 | 721.82 |  | 4.001 | 1074.00 | 1155.21 |  | 7.026 | 52.16 | 2.68 |  |
| 1.003 | 2024.16 | 1103.96 |  | 4.002 | 848.80 | 697.24 |  | 7.027 | 204.08 | 297.20 |  |
| 1.005 | 454.68 | 31.06 |  | 4.003 | 875.16 | 336.32 |  | 7.028 | 19.44 | 52.24 |  |
| 1.007 | 1360.44 | 1093.93 |  | 4.004 | 55.04 | 96.71 |  | 7.031 | 1107.12 | 0.39 |  |
| 2.003 | 339.28 |  |  | 4.005 | 85.48 | 155.61 |  | 7.031 S |  | 60.00 | t |
| 2.006 | 298.00 |  |  | 4.009 | 659.36 | 422.41 |  | 7.032 | 308.24 | 126.38 |  |
| 2.007 | 469.36 |  |  | 4.012 |  | 201.82 |  | 8.001 | 27.08 | 126.83 |  |
| 2.008 | 331.88 | 329.77 |  | 4.013 | 738.16 | 903.50 |  | 8.002 | 87.48 | 122.39 |  |
| 2.009 |  | 3.62 |  | 5.004 |  | 1.40 |  | 8.003 | 5.48 | 7.06 |  |
| 2.010 | 2.24 | 57.93 |  | 5.009 | 342.48 | 227.78 |  | 8.004 | 156.20 | 143.39 |  |
| 2.011 | 973.44 | 612.58 |  | 5.010 | 1506.36 | 481.23 |  | 8.016 | 64.64 | 51.42 |  |
| 2.012 | 212.84 | 249.58 |  | 5.011 | 556.96 | 420.01 |  | 9.001 | 138.60 | 104.83 |  |
| 2.013 | 26.24 | 0.79 |  | 5.012 | 395.32 | 338.27 |  | 9.002 | 232.20 | 62.44 |  |
| 2.014 | 284.48 | 562.55 |  | 5.013 | 835.60 | 276.32 |  | 9.003 |  | 0.40 | a |
| 2.015 | 128.68 | 283.88 |  | 5.014 | 114.56 | 72.97 |  | 9.010 | 26.28 | 6.22 |  |
| 2.017 | 341.80 | 299.85 |  | 5.016 | 525.84 | 6.02 |  | 9.011 | 5.56 | 9.35 |  |
| 2.018 | 291.12 | 698.26 |  | 5.017 | 996.16 | 450.53 |  | 9.012 | 103.08 | 33.15 |  |
| 2.019 | 34.36 | 92.27 |  | 5.018 |  | 8.36 |  | 10.001 | 266.12 | 107.11 |  |
| 2.031 | 419.60 | 357.48 |  | 5.019 | 3.72 |  |  | 10.002 | 485.80 | 238.41 |  |
| 2.036 | 95.16 |  |  | 5.020 | 2174.24 | 1865.76 |  | 10.003 | 214.84 | 185.97 |  |
| 2.037 | 28.96 |  |  | 5.021 | 667.08 | 191.63 |  | 10.004 | 74.04 | 134.37 |  |
| 2.049 | 193.36 | 20.15 |  | 5.022 | 2169.88 | 1132.48 |  | 10.005 |  | 3.60 | a |
| 2.050 | 146.24 | 18.87 |  | 6.005 | 65.72 | 4.08 |  | 10.006 |  | 0.88 |  |
| 2.051 | 27.20 |  |  | 6.006 |  | 2.72 |  | 10.008 |  | 0.72 |  |
| 2.055 |  | 1.54 |  | 6.009 | 3230.72 | 1970.07 |  | 10.011 |  | 2.71 |  |
| 2.059 |  |  |  | 6.010 | 1671.72 | 1175.59 |  | 10.012 |  | 12.40 |  |
| 2.064 | 29.36 |  |  | 6.011 | 99.68 | 91.14 |  | 101.001 |  | 33.46 |  |
| 2.065 | 27.04 | 2.26 |  | 6.012 | 203.48 | 24.25 |  | 101.002 |  | 91.14 |  |
| 2.066 | 5.80 |  |  | 6.013 | 2545.40 | 1442.52 |  | 101.006 |  | 41.77 |  |
| 2.067 | 61.84 |  |  | 6.014 | 299.88 | 251.46 |  | 101.007 |  | 41.96 |  |
| 2.068 | 561.88 | 387.95 |  | 6.015 | 520.08 | 202.22 |  | 102.002 |  | 0.36 |  |
| 2.069 |  | 24.75 |  | 6.016 | 903.76 | 330.83 |  | 105.001 | 50.32 | 215.36 |  |
| 2.070 |  | 2.10 | a | 6.017 | 510.28 | 202.35 |  | 105.002 | 6.56 | 18.24 |  |
| 2.071 |  | 76.72 |  | 6.018 | 115.40 |  |  | 106.001 |  | 10.21 |  |
| 2.073 |  | 0.05 |  | 6.019 | 109.32 | 109.20 |  | 106.002 | 1827.88 | 870.92 |  |
| 2.074 |  | 1.40 | a | 6.020 | 29.80 |  |  | 142.001 |  | 10.06 |  |
| 2.075 |  | 133.84 |  | 6.025 | 28.80 |  |  | 142.002 |  | 93.04 |  |
| 2.078 |  | 11.93 |  | 6.026 |  | 4.31 |  | North Coast |  |  |  |
| 2.079 |  | 4.91 |  | 7.001 | 55.16 | 32.47 |  | Total | 50205.20 | 29308.63 |  |
| 2.080 |  | 9.61 |  | 7.001 N |  | 33.80 | t |  |  |  |  |
| 2.082 |  | 28.84 |  | 7.002 |  | 3.10 | a |  |  |  |  |
| 2.086 |  | 1.83 |  | 7.002 N |  | 53.50 | t |  |  |  |  |
| 2.087 | 32.60 | 118.84 |  | 7.003 | 439.20 | 311.73 |  |  |  |  |  |
| 2.088 | 61.88 | 36.58 |  | 7.004 | 208.72 | 87.57 |  |  |  |  |  |
| 2.089 |  | 0.40 |  | 7.005 | 56.52 | 7.54 |  |  |  |  |  |
| 2.092 |  | 18.89 |  | 7.006 | 142.40 |  |  |  |  |  |  |
| 2.093 |  | 26.91 |  | 7.008 |  | 26.01 |  |  |  |  |  |
| 2.094 | 47.60 |  |  | 7.009 | 713.12 | 388.32 |  |  |  |  |  |
| 2.095 | 68.84 | 0.56 |  | 7.012 | 148.84 | 167.84 |  |  |  |  |  |
| 2.096 | 96.56 | 17.51 |  | 7.015 |  | 0.40 | a |  |  |  |  |
| 2.097 |  | 35.88 |  | 7.017 | 3.48 |  |  |  |  |  |  |
| 2.098 | 98.84 | 23.97 |  | 7.018 | 1294.32 | 231.16 |  |  |  |  |  |
| 2.099 |  | 24.19 |  | 7.019 | 94.60 | 71.70 |  |  |  |  |  |
| 2.100 | 11.80 | 13.36 |  | 7.020 | 51.76 | 35.17 |  |  |  |  |  |
| 3.001 | 724.16 | 458.91 |  | 7.021 | 40.08 | 27.57 |  |  |  |  |  |
| 3.002 | 33.00 | 72.13 |  | 7.023 | 51.56 | 31.29 |  |  |  |  |  |
| 3.003 | 73.68 | 18.54 |  | 7.024 | 2.64 | 0.00 |  |  |  |  |  |
|  |  |  |  |  | 28 |  |  |  |  |  |  |

Table 4. (continued)

| PFM | Bed Areas (ha) during |  |
| :---: | :---: | :---: |
|  |  |  |
| Subarea | $1982-96$ | $1997-2000$ |
|  |  |  |


| PFM | Bed Areas (ha) during |  |
| :---: | :--- | :--- |
|  |  |  |
| Subarea | $1982-96$ | $1997-2000$ |

South Coast of BC

| 11.001 | 135.88 | 29.92 | 13.033 | 53.20 | 45.76 |  | 25.006 | 70.32 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11.002 | 142.64 | 135.83 | 13.035 | 71.92 | 6.91 |  | 25.007 | 193.44 |  |
| 12.001 | 61.60 | 20.71 | 13.036 | 46.08 |  |  | 25.013 | 297.40 |  |
| 12.002 | 96.52 | 4.09 | 13.039 | 62.60 | 24.36 |  | 25.015 | 41.44 |  |
| 12.003 | 180.60 | 91.58 | 13.040 | 23.32 |  |  | 26.001 | 12.12 |  |
| 12.004 | 2.28 | 0.37 | 13.041 | 18.68 | 2.10 | a | 26.006 | 25.68 |  |
| 12.005 | 143.20 | 56.68 | 14.005 |  | 16.03 | S | 27.001 | 29.44 | 63.84 |
| 12.006 | 146.12 | 20.16 | 14.007 | 79.92 | 20.70 | S | 27.002 | 268.44 | 1.64 |
| 12.007 | 73.48 | 35.38 | 14.008 | 22.00 | 23.80 | s | 27.003 | 42.12 |  |
| 12.008 | 93.00 | 401.45 | 14.009 | 205.36 | 123.04 | S | 27.005 | 21.76 |  |
| 12.009 |  | 5.23 | 14.010 | 226.00 |  |  | 27.007 | 46.16 |  |
| 12.010 |  | 0.34 | 14.011 | 75.92 | 59.31 | s | 27.009 | 59.08 | 43.85 |
| 12.011 | 239.40 | 134.33 | 14.012 | 55.80 | 56.46 | s | 28.001 | 6.48 |  |
| 12.012 | 172.60 | 89.27 | 14.013 | 182.28 | 12.73 | a | 29.002 | 1.12 |  |
| 12.013 | 117.32 | 31.57 | 15.001 | 6.68 |  |  | 29.003 | 3.84 |  |
| 12.014 | 425.30 | 121.14 | 15.002 | 25.80 |  |  | 29.004 | 23.80 | 23.20 |
| 12.015 | 27.56 | 36.29 | 15.004 | 4.44 |  |  | 29.005 | 35.48 | 89.40 |
| 12.016 | 303.72 | 112.00 | 17.001 | 1.16 |  |  | 111.000 | 205.40 | 100.44 |
| 12.017 | 39.08 | 45.93 | 17.002 | 65.76 | 21.47 |  | 123.003 |  | 15.28 |
| 12.018 | 350.04 | 282.29 | 17.003 | 69.72 | 95.33 |  | 124.003 | 2.76 | 45.16 |
| 12.019 | 65.72 | 57.51 | 17.008 | 30.08 | 15.10 |  | 125.001 | 54.92 |  |
| 12.020 | 1.40 | 3.58 | 17.010 | 40.04 | 43.40 |  | 127.003 |  | 45.45 |
| 12.021 | 17.64 | 15.08 | 17.012 | 4.04 |  |  | South Coast | 9669.18 | 6459.09 |
| 12.026 | 3.96 |  | 17.017 | 8.12 | 7.86 |  | Total |  |  |
| 12.039 | 59.84 | 15.36 | 18.001 | 89.84 | 20.72 |  | BC Total | 59874.38 | 35767.72 |
| 12.041 | 86.76 | 29.8 | 18.00 | 131.60 | 77 |  |  |  |  |

Table 5. Estimated mean density (number $/ \mathrm{m}^{2}$ ) of red sea urchins in British Columbia by PFM subarea, obtained from population surveys during 1994-2001. Only those transects which intersect within the beds fished during 1997-2000 are included. $\mathrm{SE}=$ standard error not included if fewer than 5 transects were used.

| PFM <br> Subarea | Survey | Year | Total Transect |  | Urchin density (number $/ \mathrm{m}^{2}$ ) for each size group (mm TD) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\geq 90$ |  | $\geq 95$ |  | $\geq 100$ |  | 90-130 |  | 90-140 |  | 95-130 |  | 95-140 |  | 100-140 |  |
|  |  |  | Number | Length (m) | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |
| North Coast of BC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1.002 | Langara | 1994 | 10 | 1267 | 3.62 | 0.85 | 3.29 | 0.74 | 2.95 | 0.68 | 3.27 | 0.81 | 3.49 | 0.82 | 2.94 | 0.70 | 3.16 | 0.72 | 2.81 | 0.66 |
| 1.003 | Langara | 1994 | 8 | 1046 | 1.86 | 0.56 | 1.78 | 0.54 | 1.76 | 0.53 | 1.16 | 0.42 | 1.58 | 0.58 | 1.08 | 0.40 | 1.50 | 0.56 | 1.48 | 0.55 |
| 1.007 | Langara | 1994 | 8 | 2209 | 0.77 | 0.25 | 0.75 | 0.24 | 0.70 | 0.21 | 0.55 | 0.19 | 0.68 | 0.23 | 0.52 | 0.18 | 0.66 | 0.22 | 0.61 | 0.19 |
| 2.018 | Rennell Sound | 1995 | 8 | 678 | 1.06 | 0.33 | 0.92 | 0.28 | 0.79 | 0.25 | 0.97 | 0.29 | 1.02 | 0.31 | 0.84 | 0.25 | 0.88 | 0.26 | 0.75 | 0.23 |
| 2.031 | Rennell Sound | 1995 | 6 | 541 | 1.89 | 0.82 | 1.66 | 0.77 | 1.57 | 0.77 | 1.13 | 0.36 | 1.57 | 0.68 | 0.89 | 0.27 | 1.34 | 0.62 | 1.24 | 0.62 |
| 2.068 | Rennell Sound | 1995 | 2 | 216 | 2.20 |  | 1.55 |  | 1.14 |  | 2.20 |  | 2.20 |  | 1.55 |  | 1.55 |  | 1.14 |  |
| 2.071 | Rennell Sound | 1995 | 1 | 53 | 3.50 |  | 2.62 |  | 1.89 |  | 3.50 |  | 3.50 |  | 2.62 |  | 2.62 |  | 1.89 |  |
| 2.074 | Rennell Sound | 1995 | 1 | 201 | 1.81 |  | 1.44 |  | 1.15 |  | 1.81 |  | 1.81 |  | 1.44 |  | 1.44 |  | 1.15 |  |
| 2.075 | Rennell Sound | 1995 | 1 | 138 | 4.44 |  | 4.22 |  | 4.00 |  | 4.22 |  | 4.44 |  | 4.00 |  | 4.22 |  | 4.00 |  |
| 2.078 | Rennell Sound | 1995 | 1 | 30 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 4.002 | Stephens Is. | 1995 | 16 | 1041 | 2.14 | 0.55 | 1.87 | 0.50 | 1.54 | 0.40 | 1.66 | 0.43 | 1.85 | 0.48 | 1.39 | 0.36 | 1.58 | 0.41 | 1.26 | 0.32 |
| 4.004 | Stephens Is. | 1995 | 2 | 155 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 4.009 | Stephens Is. | 1995 | 3 | 326 | 1.39 |  | 1.22 |  | 1.04 |  | 1.30 |  | 1.36 |  | 1.14 |  | 1.20 |  | 1.02 |  |
| 5.011 | Banks Is. | 1997 | 3 | 466 | 0.53 |  | 0.48 |  | 0.45 |  | 0.45 |  | 0.49 |  | 0.39 |  | 0.44 |  | 0.41 |  |
| 5.013 | Banks Is. | 1997 | 3 | 396 | 0.41 |  | 0.36 |  | 0.34 |  | 0.28 |  | 0.33 |  | 0.24 |  | 0.29 |  | 0.26 |  |
| 5.020 | Banks Is. | 1997 | 14 | 1586 | 1.53 | 0.33 | 1.36 | 0.29 | 1.23 | 0.26 | 1.22 | 0.29 | 1.40 | 0.32 | 1.05 | 0.26 | 1.23 | 0.28 | 1.10 | 0.25 |
| 5.021 | Banks Is. | 1997 | 8 | 1274 | 1.17 | 0.31 | 1.03 | 0.27 | 0.88 | 0.24 | 0.96 | 0.26 | 1.05 | 0.28 | 0.82 | 0.22 | 0.90 | 0.24 | 0.76 | 0.20 |
| 6.010 | Campania | 1994 | 22 | 4186 | 1.67 | 0.22 | 1.57 | 0.21 | 1.40 | 0.18 | 1.37 | 0.18 | 1.56 | 0.20 | 1.27 | 0.16 | 1.45 | 0.18 | 1.28 | 0.15 |
| 6.012 | Campania | 1994 | 1 | 105 | 0.16 |  | 0.16 |  | 0.16 |  | 0.08 |  | 0.08 |  | 0.08 |  | 0.08 |  | 0.08 |  |
| 6.014 | Laredo | 2000 | 26 | 1649 | 0.50 | 0.17 | 0.43 | 0.14 | 0.36 | 0.12 | 0.45 | 0.15 | 0.48 | 0.16 | 0.37 | 0.12 | 0.41 | 0.13 | 0.34 | 0.11 |
| 6.015 | Laredo | 2000 | 28 | 1298 | 0.75 | 0.22 | 0.64 | 0.19 | 0.57 | 0.17 | 0.67 | 0.20 | 0.72 | 0.21 | 0.57 | 0.17 | 0.62 | 0.18 | 0.54 | 0.16 |
| 6.016 | Price Is. | 1995 | 7 | 409 | 1.26 | 0.14 | 1.12 | 0.13 | 0.89 | 0.13 | 1.24 | 0.14 | 1.24 | 0.14 | 1.10 | 0.14 | 1.11 | 0.14 | 0.88 | 0.13 |
| 6.017 | Price Is. | 1995 | 3 | 216 | 1.82 |  | 1.70 |  | 1.34 |  | 1.62 |  | 1.79 |  | 1.49 |  | 1.67 |  | 1.31 |  |
| $7.001 \mathrm{~N}^{\text {a }}$ | Price Is. | 2001 | 5 | 119 | 2.77 | 0.28 | 2.07 | 0.24 | 1.40 | 0.26 | 2.65 | 0.27 | 2.70 | 0.29 | 1.96 | 0.21 | 2.01 | 0.24 | 1.33 | 0.24 |
| $7.002 \mathrm{~N}^{\text {a }}$ | Price Is. | 2001 | 9 | 247 | 4.36 | 0.66 | 3.54 | 0.52 | 2.54 | 0.52 | 4.25 | 0.59 | 4.33 | 0.64 | 3.44 | 0.46 | 3.51 | 0.50 | 2.52 | 0.50 |
| 7.003 | Price Is. | 2001 | 8 | 224 | 3.31 | 0.85 | 2.83 | 0.71 | 2.40 | 0.62 | 3.15 | 0.81 | 3.28 | 0.84 | 2.67 | 0.67 | 2.81 | 0.70 | 2.37 | 0.62 |
| 7.018 | Heiltsuk | 1994 | 4 | 187 | 2.89 |  | 2.24 |  | 1.84 |  | 2.76 |  | 2.89 |  | 2.11 |  | 2.24 |  | 1.84 |  |
| 7.018 | Heiltsuk | 1995 | 5 | 305 | 1.47 | 0.59 | 1.24 | 0.46 | 0.97 | 0.35 | 1.33 | 0.57 | 1.45 | 0.60 | 1.10 | 0.44 | 1.22 | 0.47 | 0.95 | 0.35 |
| 7.018 | Heiltsuk | 1996 | 8 | 587 | 1.60 | 0.22 | 1.52 | 0.23 | 1.20 | 0.29 | 1.55 | 0.19 | 1.60 | 0.22 | 1.46 | 0.20 | 1.52 | 0.23 | 1.20 | 0.29 |
| 7.018 | Heiltsuk | 1997 | 4 | 263 | 2.21 |  | 1.91 |  | 1.59 |  | 2.09 |  | 2.20 |  | 1.78 |  | 1.90 |  | 1.58 |  |
| $7.031 \mathrm{~S}^{\text {a }}$ | Price Is. | 2001 | 10 | 306 | 2.58 | 0.48 | 2.05 | 0.40 | 1.33 | 0.29 | 2.56 | 0.49 | 2.58 | 0.48 | 2.03 | 0.41 | 2.05 | 0.40 | 1.33 | 0.29 |
| 8.004 | FitzHugh | 2001 | 1 | 21 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 8.016 | FitzHugh | 2001 | 3 | 115 | 1.37 |  | 1.33 |  | 1.18 |  | 1.32 |  | 1.37 |  | 1.28 |  | 1.33 |  | 1.18 |  |
| 106.002 | Campania | 1994 | 7 | 544 | 3.18 | 0.73 | 2.78 | 0.69 | 2.45 | 0.64 | 2.89 | 0.64 | 3.12 | 0.70 | 2.49 | 0.60 | 2.72 | 0.67 | 2.38 | 0.61 |


| PFM <br> Subarea | Survey | Year | Total Transect |  | Urchin density (number $/ \mathrm{m}^{2}$ ) for each size group (mm TD) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\geq 90$ |  | $\geq 95$ |  | $\geq 100$ |  | 90-130 |  | 90-140 |  | 95-130 |  | 95-140 |  | 100-140 |  |
|  |  |  | Number | Length (m) | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |
| South Coast of BC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 11.002 | Cape Sutil | 1996 | 2 | 123 | 0.77 |  | 0.67 |  | 0.67 |  | 0.21 |  | 0.44 |  | 0.11 |  | 0.34 |  | 0.34 |  |
| 12.001 | Kelsey Bay | 1999 | 3 | 91 | 2.04 |  | 1.89 |  | 1.78 |  | 1.80 |  | 1.97 |  | 1.65 |  | 1.82 |  | 1.72 |  |
| 12.003 | Queen Charl. Strait | 1994 | 2 | 42 | 1.12 |  | 1.12 |  | 1.05 |  | 0.98 |  | 0.98 |  | 0.98 |  | 0.98 |  | 0.91 |  |
| 12.003 | Robson Bight | 2001 | 9 | 315 | 0.92 | 0.25 | 0.89 | 0.24 | 0.85 | 0.23 | 0.55 | 0.19 | 0.70 | 0.20 | 0.52 | 0.17 | 0.67 | 0.19 | 0.64 | 0.18 |
| 12.005 | Queen Charl. Strait | 1994 | 2 | 106 | 0.32 |  | 0.32 |  | 0.29 |  | 0.11 |  | 0.22 |  | 0.11 |  | 0.22 |  | 0.18 |  |
| 12.006 | Queen Charl. Strait | 1994 | 1 | 35 | 0.96 |  | 0.84 |  | 0.84 |  | 0.72 |  | 0.84 |  | 0.60 |  | 0.72 |  | 0.72 |  |
| 12.008 | Queen Charl. Strait | 1994 | 1 | 53 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 12.009 | Queen Charl. Strait | 1994 | 2 | 170 | 2.92 |  | 2.62 |  | 2.09 |  | 2.57 |  | 2.75 |  | 2.26 |  | 2.44 |  | 1.91 |  |
| 12.011 | Queen Charl. Strait | 1994 | 4 | 150 | 3.96 |  | 3.72 |  | 3.34 |  | 3.16 |  | 3.66 |  | 2.92 |  | 3.42 |  | 3.04 |  |
| 12.013 | Deserters | 2000 | 9 | 381 | 1.52 | 0.35 | 1.38 | 0.33 | 1.15 | 0.30 | 1.32 | 0.29 | 1.45 | 0.32 | 1.17 | 0.26 | 1.30 | 0.30 | 1.07 | 0.27 |
| 12.015 | Queen Charl. Strait | 1994 | 1 | 34 | 1.35 |  | 1.35 |  | 1.35 |  | 1.35 |  | 1.35 |  | 1.35 |  | 1.35 |  | 1.35 |  |
| 12.016 | Queen Charl. Strait | 1994 | 5 | 205 | 2.20 | 1.19 | 1.99 | 1.00 | 1.78 | 0.96 | 1.37 | 0.82 | 1.76 | 1.06 | 1.16 | 0.64 | 1.55 | 0.88 | 1.34 | 0.82 |
| 12.018 | Queen Charl. Strait | 1994 | 5 | 362 | 0.72 | 0.41 | 0.71 | 0.40 | 0.67 | 0.39 | 0.46 | 0.26 | 0.64 | 0.37 | 0.45 | 0.25 | 0.63 | 0.36 | 0.59 | 0.35 |
| 12.019 | Queen Charl. Strait | 1994 | 1 | 79 | 1.06 |  | 0.92 |  | 0.87 |  | 0.63 |  | 0.92 |  | 0.48 |  | 0.77 |  | 0.72 |  |
| 12.020 | Queen Charl. Strait | 1994 | 1 | 59 | 2.80 |  | 2.80 |  | 2.80 |  | 0.89 |  | 1.77 |  | 0.89 |  | 1.77 |  | 1.77 |  |
| 12.021 | Queen Charl. Strait | 1994 | 1 | 28 | 2.13 |  | 2.02 |  | 1.68 |  | 1.79 |  | 2.13 |  | 1.68 |  | 2.02 |  | 1.68 |  |
| 12.021 | Robson Bight | 2001 | 3 | 93 | 1.34 |  | 1.32 |  | 1.25 |  | 0.66 |  | 0.92 |  | 0.64 |  | 0.89 |  | 0.83 |  |
| 12.039 | Queen Charl. Strait | 1995 | 1 | 31 | 0.90 |  | 0.83 |  | 0.83 |  | 0.14 |  | 0.51 |  | 0.07 |  | 0.43 |  | 0.43 |  |
| $13.030^{\text {a }}$ | Kelsey Bay | 1999 | 2 | 50 | 1.08 |  | 1.04 |  | 0.96 |  | 0.89 |  | 1.00 |  | 0.85 |  | 0.96 |  | 0.89 |  |
| 13.031 | Kelsey Bay | 1999 | 1 | 41 | 0.95 |  | 0.81 |  | 0.67 |  | 0.86 |  | 0.95 |  | 0.71 |  | 0.81 |  | 0.67 |  |
| 13.032 | Kelsey Bay | 1999 | 1 | 145 | 1.16 |  | 1.12 |  | 1.08 |  | 0.99 |  | 1.11 |  | 0.95 |  | 1.07 |  | 1.03 |  |
| 13.033 | Kelsey Bay | 1999 | 5 | 363 | 1.96 | 0.40 | 1.77 | 0.38 | 1.55 | 0.34 | 1.78 | 0.34 | 1.94 | 0.40 | 1.59 | 0.32 | 1.75 | 0.37 | 1.52 | 0.33 |
| $14.005^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 202 | 0.40 |  | 0.31 |  | 0.23 |  | 0.39 |  | 0.39 |  | 0.30 |  | 0.30 |  | 0.23 |  |
| $14.007^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 244 | 0.82 |  | 0.74 |  | 0.58 |  | 0.75 |  | 0.81 |  | 0.67 |  | 0.73 |  | 0.57 |  |
| $14.008^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 202 | 0.98 |  | 0.95 |  | 0.89 |  | 0.74 |  | 0.92 |  | 0.71 |  | 0.89 |  | 0.83 |  |
| $14.009^{\text {a }}$ | Comox, Denman Is. | 1999 | 4 | 866 | 0.69 |  | 0.61 |  | 0.49 |  | 0.63 |  | 0.68 |  | 0.55 |  | 0.60 |  | 0.48 |  |
| $14.011^{\text {a }}$ | Comox, Denman Is. | 1999 | 5 | 503 | 0.30 | 0.19 | 0.30 | 0.18 | 0.28 | 0.17 | 0.20 | 0.13 | 0.27 | 0.17 | 0.20 | 0.13 | 0.26 | 0.17 | 0.24 | 0.16 |
| $14.012^{\text {a }}$ | Comox, Denman Is. | 1999 | 1 | 101 | 0.69 |  | 0.65 |  | 0.57 |  | 0.65 |  | 0.69 |  | 0.61 |  | 0.65 |  | 0.57 |  |
| 18.001 | Gulf Is. | 1998 | 1 | 105 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 18.002 | Gulf Is. | 1998 | 3 | 165 | 0.20 |  | 0.20 |  | 0.20 |  | 0.11 |  | 0.13 |  | 0.11 |  | 0.13 |  | 0.13 |  |
| 18.003 | Gulf Is. | 1998 | 1 | 45 | 0.22 |  | 0.22 |  | 0.22 |  | 0.04 |  | 0.13 |  | 0.04 |  | 0.13 |  | 0.13 |  |
| 18.004 | Gulf Is., Cowichan | 1999 | 4 | 392 | 0.15 |  | 0.15 |  | 0.15 |  | 0.02 |  | 0.06 |  | 0.02 |  | 0.06 |  | 0.06 |  |
| 18.005 | Gulf Is., Cowichan | 1999 | 2 | 142 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 18.006 | Gulf Is., Cowichan | 1999 | 9 | 1081 | 0.38 | 0.13 | 0.37 | 0.12 | 0.36 | 0.12 | 0.15 | 0.07 | 0.22 | 0.08 | 0.15 | 0.06 | 0.22 | 0.08 | 0.20 | 0.07 |
| 18.011 | Gulf Is. | 1998 | 1 | 53 | 0.86 |  | 0.86 |  | 0.80 |  | 0.68 |  | 0.80 |  | 0.68 |  | 0.80 |  | 0.74 |  |
| 24.002 | Tofino | 2000 | 1 | 55 | 0.46 |  | 0.25 |  | 0.21 |  | 0.46 |  | 0.46 |  | 0.25 |  | 0.25 |  | 0.21 |  |
| 24.006 | Tofino | 2000 | 14 | 1574 | 0.82 | 0.27 | 0.72 | 0.22 | 0.53 | 0.17 | 0.78 | 0.26 | 0.80 | 0.26 | 0.67 | 0.22 | 0.70 | 0.22 | 0.51 | 0.16 |
| 24.008 | Tofino | 2000 | 5 | 523 | 2.05 | 0.84 | 1.62 | 0.63 | 1.35 | 0.51 | 1.85 | 0.79 | 1.95 | 0.81 | 1.41 | 0.57 | 1.52 | 0.60 | 1.24 | 0.48 |
| 27.001 | Cape Sutil | 1996 | 1 | 100 | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  | 0.00 |  |
| 111.000 | Cape Sutil | 1996 | 2 | 201 | 2.67 |  | 2.59 |  | 2.27 |  | 2.45 |  | 2.63 |  | 2.36 |  | 2.54 |  | 2.22 |  |
| 124.003 | Tofino | 2000 | 3 | 149 | 2.01 |  | 1.39 |  | 0.97 |  | 2.01 |  | 2.01 |  | 1.39 |  | 1.39 |  | 0.97 |  |

[^1]Table 6. Estimated mean biomass density $\left(\mathrm{g} / \mathrm{m}^{2}\right)$ of red sea urchins in British Columbia by PFM subarea, obtained from population surveys during 1994-2001. Only those transects which intersect within the beds fished during 1997-2000 are included. SE = standard error not included if fewer than 5 transects were used.

| PFM <br> Subarea | Survey | Year | Total Transect |  | Urchin biomass (g/m ${ }^{2}$ for each size group (mm TD) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\geq 90$ |  | $\geq 95$ |  | $\geq 100$ |  | 90-130 |  | 90-140 |  | 95-130 |  | 95-140 |  | 100-140 |  |
|  |  |  | Number | Length (m) | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |
| North Coast of BC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1.002 | Langara | 1994 | 10 | 1267 | 1726.4 | 405.3 | 1638.5 | 377.9 | 1536.1 | 360.9 | 1441.3 | 369.7 | 1598.7 | 380.4 | 1353.4 | 341.1 | 1510.8 | 352.5 | 1408.4 | 334.9 |
| 1.003 | Langara | 1994 | 8 | 1046 | 1149.0 | 306.6 | 1126.7 | 303.0 | 1121.7 | 301.3 | 558.7 | 207.2 | 871.8 | 325.5 | 536.5 | 202.1 | 849.5 | 320.8 | 844.5 | 318.7 |
| 1.007 | Langara | 1994 | 8 | 2209 | 440.7 | 139.8 | 433.6 | 137.4 | 418.5 | 130.7 | 256.3 | 86.6 | 355.6 | 114.3 | 249.2 | 83.8 | 348.5 | 111.5 | 333.4 | 103.2 |
| 2.018 | Rennell Sound | 1995 | 8 | 678 | 453.4 | 148.1 | 418.4 | 136.9 | 379.7 | 128.6 | 385.1 | 117.6 | 418.0 | 129.3 | 350.1 | 106.6 | 383.0 | 118.1 | 344.3 | 109.1 |
| 2.031 | Rennell Sound | 1995 | 6 | 541 | 1113.9 | 563.0 | 1053.3 | 554.1 | 1025.8 | 556.8 | 474.3 | 154.1 | 805.5 | 422.9 | 413.7 | 136.2 | 744.9 | 411.2 | 717.4 | 413.0 |
| 2.068 | Rennell Sound | 1995 | 2 | 216 | 779.8 |  | 607.0 |  | 487.5 |  | 779.8 |  | 779.8 |  | 607.0 |  | 607.0 |  | 487.5 |  |
| 2.071 | Rennell Sound | 1995 | 1 | 53 | 1224.4 |  | 998.7 |  | 788.2 |  | 1224.4 |  | 1224.4 |  | 998.7 |  | 998.7 |  | 788.2 |  |
| 2.074 | Rennell Sound | 1995 | 1 | 201 | 645.1 |  | 548.7 |  | 463.9 |  | 645.1 |  | 645.1 |  | 548.7 |  | 548.7 |  | 463.9 |  |
| 2.075 | Rennell Sound | 1995 | 1 | 138 | 1889.6 |  | 1833.2 |  | 1766.0 |  | 1725.4 |  | 1889.6 |  | 1668.9 |  | 1833.2 |  | 1766.0 |  |
| 2.078 | Rennell Sound | 1995 | 1 | 30 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 4.002 | Stephens Is. | 1995 | 16 | 1041 | 1124.6 | 310.2 | 1053.7 | 304.0 | 956.2 | 274.0 | 678.2 | 177.2 | 820.8 | 208.7 | 607.4 | 158.7 | 749.9 | 193.2 | 652.4 | 165.0 |
| 4.004 | Stephens Is. | 1995 | 2 | 155 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 4.009 | Stephens Is. | 1995 | 3 | 326 | 612.9 |  | 571.2 |  | 518.3 |  | 548.4 |  | 593.5 |  | 506.7 |  | 551.8 |  | 498.9 |  |
| 5.011 | Banks Is. | 1997 | 3 | 466 | 251.2 |  | 236.1 |  | 227.3 |  | 179.9 |  | 216.6 |  | 164.8 |  | 201.6 |  | 192.7 |  |
| 5.013 | Banks Is. | 1997 | 3 | 396 | 230.3 |  | 219.0 |  | 212.1 |  | 119.1 |  | 152.9 |  | 107.8 |  | 141.6 |  | 134.7 |  |
| 5.020 | Banks Is. | 1997 | 14 | 1586 | 772.1 | 152.0 | 727.0 | 142.1 | 687.8 | 132.9 | 512.6 | 115.0 | 649.5 | 138.5 | 467.6 | 105.8 | 604.4 | 128.4 | 565.1 | 118.7 |
| 5.021 | Banks Is. | 1997 | 8 | 1274 | 573.7 | 150.4 | 536.0 | 141.9 | 492.1 | 135.0 | 397.5 | 106.6 | 458.3 | 121.8 | 359.9 | 96.8 | 420.7 | 111.5 | 376.8 | 102.5 |
| 6.010 | Campania | 1994 | 22 | 4186 | 848.0 | 110.9 | 820.5 | 107.7 | 770.7 | 100.5 | 596.7 | 70.1 | 738.1 | 87.4 | 569.2 | 66.7 | 710.6 | 83.8 | 660.8 | 77.8 |
| 6.012 | Campania | 1994 | 1 | 105 | 112.3 |  | 112.3 |  | 112.3 |  | 39.2 |  | 39.2 |  | 39.2 |  | 39.2 |  | 39.2 |  |
| 6.014 | Laredo | 2000 | 26 | 1649 | 222.2 | 72.3 | 202.9 | 66.1 | 181.8 | 60.0 | 176.3 | 58.8 | 204.0 | 67.6 | 157.0 | 52.1 | 184.7 | 61.1 | 163.6 | 54.9 |
| 6.015 | Laredo | 2000 | 28 | 1298 | 331.0 | 95.2 | 304.3 | 87.9 | 280.3 | 82.0 | 271.3 | 80.7 | 305.2 | 89.4 | 244.6 | 73.2 | 278.5 | 82.0 | 254.6 | 76.2 |
| 6.016 | Price Is. | 1995 | 7 | 409 | 502.9 | 51.8 | 466.0 | 52.3 | 397.7 | 52.0 | 486.6 | 56.3 | 491.4 | 55.7 | 449.7 | 56.7 | 454.5 | 55.8 | 386.1 | 54.1 |
| 6.017 | Price Is. | 1995 | 3 | 216 | 798.8 |  | 766.4 |  | 658.4 |  | 638.8 |  | 767.7 |  | 606.4 |  | 735.3 |  | 627.3 |  |
| $7.001 \mathrm{~N}^{\text {a }}$ | Price Is. | 2001 | 5 | 119 | 1022.8 | 132.0 | 842.0 | 137.0 | 637.0 | 152.8 | 929.5 | 101.5 | 965.8 | 125.5 | 748.8 | 97.1 | 785.1 | 123.8 | 580.0 | 131.4 |
| $7.002 \mathrm{~N}^{\text {a }}$ | Price Is. | 2001 | 9 | 247 | 1615.9 | 293.7 | 1403.4 | 261.6 | 1104.7 | 262.1 | 1536.7 | 238.3 | 1594.4 | 274.8 | 1324.2 | 205.5 | 1381.8 | 242.6 | 1083.2 | 243.6 |
| 7.003 | Price Is. | 2001 | 8 | 224 | 1349.0 | 352.9 | 1224.7 | 316.0 | 1094.1 | 293.4 | 1224.4 | 322.3 | 1325.1 | 347.9 | 1100.2 | 285.1 | 1200.8 | 310.8 | 1070.2 | 288.2 |
| 7.018 | Heiltsuk | 1994 | 4 | 187 | 1164.7 |  | 993.9 |  | 871.9 |  | 1059.7 |  | 1164.7 |  | 888.9 |  | 993.9 |  | 871.9 |  |
| 7.018 | Heiltsuk | 1995 | 5 | 305 | 621.5 | 232.7 | 559.6 | 200.2 | 480.9 | 165.7 | 510.9 | 216.6 | 602.2 | 236.9 | 449.0 | 181.2 | 540.3 | 202.9 | 461.6 | 168.2 |
| 7.018 | Heiltsuk | 1996 | 8 | 587 | 675.7 | 143.2 | 652.9 | 146.7 | 556.3 | 172.8 | 635.3 | 110.1 | 675.7 | 143.2 | 612.5 | 113.0 | 652.9 | 146.7 | 556.3 | 172.8 |
| 7.018 | Heiltsuk | 1997 | 4 | 263 | 895.0 |  | 814.8 |  | 717.2 |  | 806.0 |  | 888.7 |  | 725.8 |  | 808.4 |  | 710.9 |  |
| $7.031 \mathrm{~S}^{\mathrm{a}}$ | Price Is. | 2001 | 10 | 306 | 908.1 | 175.4 | 770.1 | 156.8 | 555.9 | 123.6 | 894.5 | 177.5 | 908.1 | 175.4 | 756.5 | 158.3 | 770.1 | 156.8 | 555.9 | 123.6 |
| 8.004 | FitzHugh | 2001 | 1 | 21 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 8.016 | FitzHugh | 2001 | 3 | 115 | 620.0 |  | 611.7 |  | 566.5 |  | 583.5 |  | 620.0 |  | 575.2 |  | 611.7 |  | 566.5 |  |
| 106.002 | Campania | 1994 | 7 | 544 | 1402.5 | 359.3 | 1298.7 | 353.6 | 1195.8 | 339.3 | 1149.2 | 268.3 | 1324.3 | 326.1 | 1045.4 | 259.1 | 1220.5 | 319.6 | 1117.7 | 305.1 |

Table 6. (continued)

| PFM <br> Subarea | Survey | Year | Total Transect |  | $\geq 90$ |  | $\geq 95$ |  | Urchin biomass ( $\mathrm{g} / \mathrm{m}^{2}$ ) for each size group (mm TD) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\geq 100$ | 90-130 |  | 90-140 |  | 95-130 |  | 95-140 |  | 100-140 |  |
|  |  |  | Number | Length (m) |  |  | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |
| South Coast of BC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 11.002 | Cape Sutil | 1996 | 2 | 123 | 564.0 |  |  |  | 539.4 |  | 539.4 |  | 81.7 |  | 259.3 |  | 57.2 |  | 234.8 |  | 234.8 |  |
| 12.001 | Kelsey Bay | 1999 | 3 | 91 | 975.4 |  | 936.1 |  | 904.8 |  | 796.2 |  | 921.0 |  | 756.8 |  | 881.6 |  | 850.4 |  |
| 12.003 | Queen Charl. Strait | 1994 | 2 | 42 | 589.4 |  | 589.4 |  | 569.5 |  | 422.1 |  | 422.1 |  | 422.1 |  | 422.1 |  | 402.2 |  |
| 12.003 | Robson Bight | 2001 | 9 | 315 | 594.1 | 179.5 | 586.1 | 177.4 | 576.8 | 177.0 | 251.7 | 84.6 | 366.6 | 101.9 | 243.7 | 80.7 | 358.6 | 98.6 | 349.3 | 95.9 |
| 12.005 | Queen Charl. Strait | 1994 | 2 | 106 | 246.3 |  | 246.3 |  | 235.1 |  | 56.1 |  | 138.3 |  | 56.1 |  | 138.3 |  | 127.2 |  |
| 12.006 | Queen Charl. Strait | 1994 | 1 | 35 | 576.0 |  | 545.5 |  | 545.5 |  | 387.5 |  | 476.3 |  | 357.0 |  | 445.7 |  | 445.7 |  |
| 12.008 | Queen Charl. Strait | 1994 | 1 | 53 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 12.009 | Queen Charl. Strait | 1994 | 2 | 170 | 1532.2 |  | 1453.6 |  | 1297.9 |  | 1113.7 |  | 1253.6 |  | 1035.1 |  | 1175.0 |  | 1019.3 |  |
| 12.011 | Queen Charl. Strait | 1994 | 4 | 150 | 2088.4 |  | 2027.6 |  | 1915.0 |  | 1425.6 |  | 1796.7 |  | 1364.9 |  | 1736.0 |  | 1623.3 |  |
| 12.013 | Deserters | 2000 | 9 | 381 | 726.3 | 189.2 | 687.8 | 186.4 | 618.4 | 180.1 | 559.6 | 129.0 | 654.0 | 158.9 | 521.1 | 124.3 | 615.5 | 155.1 | 546.1 | 147.3 |
| 12.015 | Queen Charl. Strait | 1994 | 1 | 34 | 691.8 |  | 691.8 |  | 691.8 |  | 691.8 |  | 691.8 |  | 691.8 |  | 691.8 |  | 691.8 |  |
| 12.016 | Queen Charl. Strait | 1994 | 5 | 205 | 1352.0 | 719.4 | 1298.1 | 673.0 | 1236.9 | 662.0 | 604.2 | 368.0 | 899.5 | 548.6 | 550.3 | 321.0 | 845.6 | 502.1 | 784.4 | 485.2 |
| 12.018 | Queen Charl. Strait | 1994 | 5 | 362 | 434.8 | 253.0 | 430.7 | 250.5 | 419.3 | 246.8 | 226.9 | 130.6 | 361.1 | 215.6 | 222.8 | 128.2 | 357.0 | 213.2 | 345.6 | 209.6 |
| 12.019 | Queen Charl. Strait | 1994 | 1 | 79 | 637.6 |  | 598.2 |  | 584.0 |  | 273.9 |  | 498.7 |  | 234.5 |  | 459.4 |  | 445.2 |  |
| 12.020 | Queen Charl. Strait | 1994 | 1 | 59 | 2021.8 |  | 2021.8 |  | 2021.8 |  | 423.9 |  | 1087.4 |  | 423.9 |  | 1087.4 |  | 1087.4 |  |
| 12.021 | Queen Charl. Strait | 1994 | 1 | 28 | 1021.6 |  | 991.4 |  | 888.0 |  | 782.9 |  | 1021.6 |  | 752.7 |  | 991.4 |  | 888.0 |  |
| 12.021 | Robson Bight | 2001 | 3 | 93 | 931.3 |  | 926.0 |  | 906.5 |  | 325.0 |  | 512.9 |  | 319.7 |  | 507.6 |  | 488.1 |  |
| 12.039 | Queen Charl. Strait | 1995 | 1 | 31 | 729.5 |  | 710.6 |  | 710.6 |  | 57.9 |  | 334.5 |  | 39.0 |  | 315.5 |  | 315.5 |  |
| $13.030^{\text {a }}$ | Kelsey Bay | 1999 | 2 | 50 | 551.2 |  | 541.1 |  | 516.7 |  | 397.9 |  | 486.8 |  | 387.8 |  | 476.6 |  | 452.2 |  |
| 13.031 | Kelsey Bay | 1999 | 1 | 41 | 439.0 |  | 401.3 |  | 358.5 |  | 366.4 |  | 439.0 |  | 328.6 |  | 401.3 |  | 358.5 |  |
| 13.032 | Kelsey Bay | 1999 | 1 | 145 | 624.8 |  | 614.3 |  | 601.8 |  | 485.0 |  | 574.6 |  | 474.4 |  | 564.0 |  | 551.6 |  |
| 13.033 | Kelsey Bay | 1999 | 5 | 363 | 885.9 | 198.4 | 835.4 | 193.1 | 765.9 | 179.6 | 748.7 | 148.3 | 862.6 | 190.6 | 698.2 | 142.3 | 812.0 | 185.0 | 742.5 | 171.6 |
| $14.005^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 202 | 160.5 |  | 136.5 |  | 112.9 |  | 152.2 |  | 152.2 |  | 128.2 |  | 128.2 |  | 104.6 |  |
| $14.007^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 244 | 362.5 |  | 341.2 |  | 292.8 |  | 307.0 |  | 354.7 |  | 285.8 |  | 333.5 |  | 285.0 |  |
| $14.008^{\text {a }}$ | Comox, Denman Is. | 1999 | 2 | 202 | 545.0 |  | 537.1 |  | 519.2 |  | 348.0 |  | 491.0 |  | 340.1 |  | 483.0 |  | 465.2 |  |
| $14.009^{\text {a }}$ | Comox, Denman Is. | 1999 | 4 | 866 | 296.1 |  | 275.3 |  | 240.7 |  | 252.0 |  | 287.5 |  | 231.2 |  | 266.7 |  | 232.1 |  |
| $14.011^{\text {a }}$ | Comox, Denman Is. | 1999 | 5 | 503 | 186.3 | 114.6 | 184.2 | 113.3 | 178.3 | 109.7 | 100.6 | 66.2 | 147.6 | 94.9 | 98.5 | 64.4 | 145.6 | 93.3 | 139.6 | 90.6 |
| $14.012^{\text {a }}$ | Comox, Denman Is. | 1999 | 1 | 101 | 301.4 |  | 291.7 |  | 268.0 |  | 272.7 |  | 301.4 |  | 263.0 |  | 291.7 |  | 268.0 |  |
| 18.001 | Gulf Is. | 1998 | 1 | 105 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 18.002 | Gulf Is. | 1998 | 3 | 165 | 143.3 |  | 143.3 |  | 143.3 |  | 60.5 |  | 78.2 |  | 60.5 |  | 78.2 |  | 78.2 |  |
| 18.003 | Gulf Is. | 1998 | 1 | 45 | 190.2 |  | 190.2 |  | 190.2 |  | 22.3 |  | 93.3 |  | 22.3 |  | 93.3 |  | 93.3 |  |
| 18.004 | Gulf Is., Cowichan | 1999 | 4 | 392 | 131.2 |  | 131.2 |  | 131.2 |  | 9.4 |  | 39.3 |  | 9.4 |  | 39.3 |  | 39.3 |  |
| 18.005 | Gulf Is., Cowichan | 1999 | 2 | 142 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 18.006 | Gulf Is., Cowichan | 1999 | 9 | 1081 | 279.3 | 95.5 | 276.9 | 94.8 | 273.1 | 93.8 | 74.9 | 30.9 | 127.4 | 42.7 | 72.5 | 30.2 | 125.0 | 42.0 | 121.1 | 39.8 |
| 18.011 | Gulf Is. | 1998 | 1 | 53 | 530.3 |  | 530.3 |  | 512.2 |  | 370.4 |  | 462.9 |  | 370.4 |  | 462.9 |  | 444.7 |  |
| 24.002 | Tofino | 2000 | 1 | 55 | 151.6 |  | 96.3 |  | 84.9 |  | 151.6 |  | 151.6 |  | 96.3 |  | 96.3 |  | 84.9 |  |
| 24.006 | Tofino | 2000 | 14 | 1574 | 328.8 | 102.1 | 302.3 | 91.8 | 247.4 | 77.4 | 294.0 | 96.0 | 310.0 | 97.3 | 267.6 | 85.2 | 283.5 | 86.8 | 228.6 | 71.4 |
| 24.008 | Tofino | 2000 | 5 | 523 | 868.0 | 334.6 | 756.0 | 285.3 | 673.1 | 254.4 | 690.7 | 283.3 | 767.5 | 299.9 | 578.7 | 229.5 | 655.6 | 249.8 | 572.6 | 217.0 |
| 27.001 | Cape Sutil | 1996 | 1 | 100 | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 111.000 | Cape Sutil | 1996 | 2 | 201 | 1204.1 |  | 1180.1 |  | 1086.0 |  | 1027.6 |  | 1163.5 |  | 1003.7 |  | 1139.6 |  | 1045.5 |  |
| 124.003 | Tofino | 2000 | 3 | 149 | 669.9 |  | 507.2 |  | 385.4 |  | 669.9 |  | 669.9 |  | 507.2 |  | 507.2 |  | 385.4 |  |

[^2]Table 7. Summary of estimated mean density (number $/ \mathrm{m}^{2}$ ) and biomass ( $\mathrm{g} / \mathrm{m}^{2}$ ) of red sea urchins by PFM area in
British Columbia, all years combined, obtained from population surveys during 1994-2001. Only those transects which
intersect within the beds fished during 1997-2000 are included.

| PFM Area | Year | Total Transect |  | Mean urchin density (per $\mathrm{m}^{2}$ ) for each size group (mm TD) |  |  |  |  |  |  |  | Mean biomass ( $\mathrm{g} / \mathrm{m}^{2}$ ) for each size group (mm TD) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\geq 90$ | $\geq 95$ | $\geq 100$ | 90-130 | 90-140 | 95-130 | 95-140 | 100-140 | $\geq 90$ | $\geq 95$ | $\geq 100$ | 90-130 | 90-140 | 95-130 | 95-140 | 100-140 |
| North Coast of BC: |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1994 | 26 | 4522 | 1.82 | 1.70 | 1.57 | 1.45 | 1.68 | 1.33 | 1.55 | 1.43 | 964.77 | 931.55 | 894.30 | 658.25 | 823.28 | 625.03 | 790.05 | 752.80 |
| 2 | 1995 | 19 | 1656 | 1.82 | 1.56 | 1.38 | 1.52 | 1.70 | 1.25 | 1.43 | 1.26 | 847.90 | 779.30 | 726.53 | 597.31 | 732.68 | 528.71 | 664.08 | 611.31 |
| 4 | 1995 | 21 | 1522 | 1.76 | 1.54 | 1.28 | 1.42 | 1.56 | 1.19 | 1.34 | 1.08 | 900.45 | 843.04 | 765.00 | 581.35 | 688.52 | 523.94 | 631.11 | 553.08 |
| 5 | 1997 | 28 | 3722 | 1.16 | 1.03 | 0.92 | 0.94 | 1.05 | 0.80 | 0.92 | 0.81 | 581.31 | 546.15 | 512.56 | 389.69 | 477.00 | 354.53 | 441.84 | 408.24 |
| 6 | 1994-2000 | 87 | 7863 | 1.24 | 1.14 | 1.00 | 1.04 | 1.16 | 0.95 | 1.06 | 0.93 | 602.32 | 576.37 | 535.00 | 442.83 | 533.31 | 416.88 | 507.36 | 465.99 |
| 7 | 1994-2001 | 37 | 1819 | 2.10 | 1.81 | 1.48 | 1.99 | 2.09 | 1.71 | 1.80 | 1.47 | 861.93 | 787.34 | 685.65 | 779.89 | 851.87 | 705.30 | 777.28 | 675.59 |
| 8 | 2001 | 4 | 136 | 1.16 | 1.13 | 1.00 | 1.11 | 1.16 | 1.09 | 1.13 | 1.00 | 524.23 | 517.23 | 479.06 | 493.42 | 524.23 | 486.41 | 517.23 | 479.06 |
| 106 | 1994 | 7 | 544 | 3.18 | 2.78 | 2.45 | 2.89 | 3.12 | 2.49 | 2.72 | 2.38 | 1402.48 | 1298.66 | 1195.85 | 1149.20 | 1324.31 | 1045.39 | 1220.50 | 1117.68 |
| North Coa | Totals: | 229 | 21784 | 1.55 | 1.39 | 1.23 | 1.30 | 1.45 | 1.14 | 1.29 | 1.13 | 754.64 | 714.28 | 665.11 | 545.99 | 656.18 | 505.63 | 615.82 | 566.65 |
| South Coast of BC: |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 11 | 1996 | 2 | 123 | 0.77 | 0.67 | 0.67 | 0.21 | 0.44 | 0.11 | 0.34 | 0.34 | 563.96 | 539.43 | 539.43 | 81.69 | 259.31 | 57.16 | 234.78 | 234.78 |
| 12 | 1994-2001 | 51 | 2234 | 1.56 | 1.45 | 1.30 | 1.14 | 1.36 | 1.03 | 1.25 | 1.10 | 878.67 | 850.96 | 806.17 | 508.27 | 675.48 | 480.56 | 647.77 | 602.97 |
| 13 | 1999 | 9 | 599 | 1.63 | 1.49 | 1.32 | 1.45 | 1.59 | 1.31 | 1.45 | 1.29 | 764.20 | 727.59 | 677.46 | 629.40 | 732.50 | 592.79 | 695.89 | 645.76 |
| 14 | 1999 | 16 | 2118 | 0.61 | 0.56 | 0.47 | 0.53 | 0.59 | 0.47 | 0.54 | 0.45 | 288.70 | 273.76 | 247.53 | 223.00 | 269.20 | 208.05 | 254.26 | 228.03 |
| 18 | 1998-1999 | 21 | 1983 | 0.28 | 0.28 | 0.27 | 0.12 | 0.17 | 0.11 | 0.16 | 0.16 | 208.63 | 207.32 | 204.72 | 58.15 | 98.23 | 56.84 | 96.92 | 94.32 |
| 24 | 2000 | 20 | 2152 | 1.11 | 0.92 | 0.72 | 1.03 | 1.07 | 0.84 | 0.88 | 0.68 | 455.29 | 407.32 | 346.72 | 386.77 | 417.11 | 338.80 | 369.15 | 308.55 |
| 27 | 1996 | 1 | 100 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 111 | 1996 | 2 | 201 | 2.67 | 2.59 | 2.27 | 2.45 | 2.63 | 2.36 | 2.54 | 2.22 | 1204.07 | 1180.15 | 1086.00 | 1027.61 | 1163.53 | 1003.69 | 1139.61 | 1045.46 |
| 124 | 2000 | 3 | 149 | 2.01 | 1.39 | 0.97 | 2.01 | 2.01 | 1.39 | 1.39 | 0.97 | 669.92 | 507.19 | 385.44 | 669.92 | 669.92 | 507.19 | 507.19 | 385.44 |
| South Coast Totals: |  | 125 | 9659 | 1.00 | 0.90 | 0.77 | 0.81 | 0.91 | 0.70 | 0.81 | 0.68 | 500.76 | 474.53 | 437.44 | 336.35 | 411.63 | 310.12 | 385.40 | 348.31 |
| BC Totals: |  | 354 | 31443 | 1.38 | 1.24 | 1.09 | 1.15 | 1.28 | 1.01 | 1.14 | 0.99 | 676.65 | 640.63 | 595.17 | 481.59 | 581.06 | 445.57 | 545.04 | 499.58 |

Table 8. Quota (tonnes) options for the red sea urchin fishery by PFM subarea, estimated from various natural mortality values applied to current biomass (Bc) calculated from mean and approximate $90 \%$ lower confidence bound (CB) of the clipped biomass values ( $\mathrm{g} / \mathrm{m}^{2}$ ) for commercial red sea urchin sizes 90-140 mm TD, and bed areas fished during 1997-2000. NB: blank = mean values directly from survey data within subarea; a mean values from all surveys (1994-2001) in PFM area (may include several subareas); $b=$ mean values of surveys in PFM area 4 used for subareas of PFM area 3; $c=$ mean values of surveys in PFM area 7 used for subareas of PFM areas 9 and $10 ; \mathrm{d}=$ mean values of surveys in PFM area 1 used for subareas of PFM area 101; $\mathrm{e}=$ mean values of surveys in PFM area 2E used for subareas of PFM area 102; $f=$ mean values of surveys in PFM area 5 used for subareas of PFM area 105; $g=$ mean values of surveys in PFM area 2W used for subareas of PFM area 142; $\mathrm{h}=$ mean values of surveys in PFM area 18 used for subareas of PFM areas 17, 19, 29, and PFM subarea 18.005; $\mathrm{i}=$ mean values of surveys in PFM area 24 used for subareas of PFM areas 20, 23, 27, and $123 ; \mathrm{j}=$ mean values of surveys in PFM area 111 used for subareas of PFM area $127 ; k=$ bed area estimated using survey data; $y=$ Subareas with only one survey transect and/or low (0) mean biomass updated to reflect mean biomass for entire PFM area.

| PFM <br> SubArea | Year | NB | Transect |  | Biomass (g/m) |  | Bed <br> Area (ha) | Bc Biomass (t) |  | Quota 0.2 M Bc (of urchins 90-140 mm TD) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{M}=0.075$ | $\mathrm{M}=0.10$ |  |  |  | $\mathrm{M}=0.15$ |  |
|  |  |  | Number | Length <br> (m) |  |  | Mean | Lower $90 \%$ CB | Mean | Lower $90 \%$ CB | Mean | $\begin{gathered} \text { Lower } \\ 90 \% \mathrm{CB} \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ |
| North Coast of BC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1.001 | 94 | a | 26 | 4522 | 823.3 | 695.2 |  | 967.7 | 7967.2 | 6728.1 | 119.5 | 100.9 | 159.3 | 134.6 | 239.0 | 201.8 |
| 1.002 | 94 |  | 10 | 1267 | 1598.7 | 1128.9 |  | 721.8 | 11539.4 | 6506.6 | 173.1 | 97.6 | 230.8 | 130.1 | 346.2 | 195.2 |
| 1.003 | 94 |  | 8 | 1046 | 871.8 | 589.9 | 1104.0 | 9624.2 | 2816.5 | 144.4 | 42.2 | 192.5 | 56.3 | 288.7 | 84.5 |
| 1.005 | 94 | a | 26 | 4522 | 823.3 | 695.2 | 31.1 | 255.7 | 215.9 | 3.8 | 3.2 | 5.1 | 4.3 | 7.7 | 6.5 |
| 1.007 | 94 |  | 8 | 2209 | 355.6 | 276.4 | 1093.9 | 3889.7 | 1521.5 | 58.3 | 22.8 | 77.8 | 30.4 | 116.7 | 45.6 |
| 2.008 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 329.8 | 2416.2 | 1795.2 | 36.2 | 26.9 | 48.3 | 35.9 | 72.5 | 53.9 |
| 2.009 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 3.6 | 26.5 | 19.7 | 0.4 | 0.3 | 0.5 | 0.4 | 0.8 | 0.6 |
| 2.010 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 57.9 | 424.4 | 315.3 | 6.4 | 4.7 | 8.5 | 6.3 | 12.7 | 9.5 |
| 2.011 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 612.6 | 4488.3 | 3334.8 | 67.3 | 50.0 | 89.8 | 66.7 | 134.6 | 100.0 |
| 2.012 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 249.6 | 1828.6 | 1358.6 | 27.4 | 20.4 | 36.6 | 27.2 | 54.9 | 40.8 |
| 2.013 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 0.8 | 5.8 | 4.3 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.1 |
| 2.014 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 562.6 | 4121.7 | 3062.4 | 61.8 | 45.9 | 82.4 | 61.2 | 123.7 | 91.9 |
| 2.015 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 283.9 | 2079.9 | 1545.4 | 31.2 | 23.2 | 41.6 | 30.9 | 62.4 | 46.4 |
| 2.017 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 299.8 | 2196.9 | 1632.3 | 33.0 | 24.5 | 43.9 | 32.6 | 65.9 | 49.0 |
| 2.018 | 95 |  | 8 | 678 | 418.0 | 250.1 | 698.3 | 2919.0 | 1208.5 | 43.8 | 18.1 | 58.4 | 24.2 | 87.6 | 36.3 |
| 2.019 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 92.3 | 676.0 | 502.3 | 10.1 | 7.5 | 13.5 | 10.0 | 20.3 | 15.1 |
| 2.031 | 95 |  | 6 | 541 | 805.5 | 443.3 | 357.5 | 2879.6 | -166.6 | 43.2 | -2.5 | 57.6 | -3.3 | 86.4 | -5.0 |
| 2.049 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 20.1 | 147.6 | 109.7 | 2.2 | 1.6 | 3.0 | 2.2 | 4.4 | 3.3 |
| 2.050 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 18.9 | 138.3 | 102.7 | 2.1 | 1.5 | 2.8 | 2.1 | 4.1 | 3.1 |
| 2.055 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 1.5 | 11.3 | 8.4 | 0.2 | 0.1 | 0.2 | 0.2 | 0.3 | 0.3 |
| 2.065 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 2.3 | 16.6 | 12.3 | 0.2 | 0.2 | 0.3 | 0.2 | 0.5 | 0.4 |
| 2.068 | 95 |  | 2 | 216 | 779.8 | 224.9 | 388.0 | 3025.1 | 872.4 | 45.4 | 13.1 | 60.5 | 17.4 | 90.8 | 26.2 |
| 2.069 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 24.7 | 181.3 | 134.7 | 2.7 | 2.0 | 3.6 | 2.7 | 5.4 | 4.0 |
| 2.070 | 95 | a | 19 | 1656 | 732.7 | 544.4 | 2.1 | 15.4 | 11.4 | 0.2 | 0.2 | 0.3 | 0.2 | 0.5 | 0.3 |
| 2.071 | 95 |  | 1 | 53 | 1224.4 | 0.0 | 76.7 | 939.4 | 0.0 | 14.1 | 0.0 | 18.8 | 0.0 | 28.2 | 0.0 |
|  |  |  |  |  |  |  |  | 35 |  |  |  |  |  |  |  |

Table 8. (continued)


Table 8. (continued)

| $\begin{gathered} \text { PFM } \\ \text { SubArea } \end{gathered}$ | Year | NB | Transect |  | Biomass (g/m2) |  | Bed <br> Area (ha) | Bc Biomass (t) |  | Quota 0.2 M Bc (of urchins 90-140 mm TD) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{M}=0.075$ | $\mathrm{M}=0.10$ |  |  |  | $\mathrm{M}=0.15$ |  |
|  |  |  | Number | Length <br> (m) |  |  | Mean | $\begin{gathered} \text { Lower } \\ 90 \% \mathrm{CB} \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ |
| 5.016 | 97 | a | 28 | 3722 | 477.0 | 395.2 |  | 6.0 | 28.7 | 23.8 | 0.4 | 0.4 | 0.6 | 0.5 | 0.9 | 0.7 |
| 5.017 | 97 | a | 28 | 3722 | 477.0 | 395.2 |  | 450.5 | 2149.0 | 1780.6 | 32.2 | 26.7 | 43.0 | 35.6 | 64.5 | 53.4 |
| 5.018 | 97 | a | 28 | 3722 | 477.0 | 395.2 | 8.4 | 39.9 | 33.0 | 0.6 | 0.5 | 0.8 | 0.7 | 1.2 | 1.0 |
| 5.020 | 97 |  | 14 | 1586 | 649.5 | 478.9 | 1865.8 | 12117.4 | 7541.2 | 181.8 | 113.1 | 242.3 | 150.8 | 363.5 | 226.2 |
| 5.021 | 97 |  | 8 | 1274 | 458.3 | 324.0 | 191.6 | 878.2 | 436.1 | 13.2 | 6.5 | 17.6 | 8.7 | 26.3 | 13.1 |
| 5.022 | 97 | a | 28 | 3722 | 477.0 | 395.2 | 1132.5 | 5401.9 | 4475.9 | 81.0 | 67.1 | 108.0 | 89.5 | 162.1 | 134.3 |
| 6.005 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 4.1 | 21.8 | 19.2 | 0.3 | 0.3 | 0.4 | 0.4 | 0.7 | 0.6 |
| 6.006 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 2.7 | 14.5 | 12.8 | 0.2 | 0.2 | 0.3 | 0.3 | 0.4 | 0.4 |
| 6.009 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 1970.1 | 10506.5 | 9267.3 | 157.6 | 139.0 | 210.1 | 185.3 | 315.2 | 278.0 |
| 6.010 | 94 |  | 22 | 4186 | 738.1 | 618.8 | 1175.6 | 8677.6 | 6909.3 | 130.2 | 103.6 | 173.6 | 138.2 | 260.3 | 207.3 |
| 6.011 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 91.1 | 486.1 | 428.7 | 7.3 | 6.4 | 9.7 | 8.6 | 14.6 | 12.9 |
| 6.012 | 94-00 | y | 87 | 7863 | 533.3 | 470.4 | 24.3 | 129.3 | 114.1 | 1.9 | 1.7 | 2.6 | 2.3 | 3.9 | 3.4 |
| 6.013 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 1442.5 | 7693.1 | 6785.7 | 115.4 | 101.8 | 153.9 | 135.7 | 230.8 | 203.6 |
| 6.014 | 00 |  | 26 | 1649 | 204.0 | 151.5 | 251.5 | 513.1 | 222.8 | 7.7 | 3.3 | 10.3 | 4.5 | 15.4 | 6.7 |
| 6.015 | 00 |  | 28 | 1298 | 305.2 | 216.6 | 202.2 | 617.2 | 309.2 | 9.3 | 4.6 | 12.3 | 6.2 | 18.5 | 9.3 |
| 6.016 | 95 |  | 7 | 409 | 491.4 | 237.3 | 330.8 | 1625.6 | 1267.7 | 24.4 | 19.0 | 32.5 | 25.4 | 48.8 | 38.0 |
| 6.017 | 95 |  | 3 | 216 | 767.7 | 221.4 | 202.4 | 1553.5 | 448.0 | 23.3 | 6.7 | 31.1 | 9.0 | 46.6 | 13.4 |
| 6.019 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 109.2 | 582.4 | 513.7 | 8.7 | 7.7 | 11.6 | 10.3 | 17.5 | 15.4 |
| 6.026 | 94-00 | a | 87 | 7863 | 533.3 | 470.4 | 4.3 | 23.0 | 20.3 | 0.3 | 0.3 | 0.5 | 0.4 | 0.7 | 0.6 |
| 7.001 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 32.5 | 276.6 | 208.8 | 4.1 | 3.1 | 5.5 | 4.2 | 8.3 | 6.3 |
| 7.002 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 3.5 | 30.1 | 22.7 | 0.5 | 0.3 | 0.6 | 0.5 | 0.9 | 0.7 |
| 7.003 | 01 |  | 8 | 224 | 1325.1 | 399.1 | 311.7 | 4130.6 | 2076.1 | 62.0 | 31.1 | 82.6 | 41.5 | 123.9 | 62.3 |
| 7.004 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 87.6 | 746.0 | 563.0 | 11.2 | 8.4 | 14.9 | 11.3 | 22.4 | 16.9 |
| 7.005 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 7.5 | 64.2 | 48.5 | 1.0 | 0.7 | 1.3 | 1.0 | 1.9 | 1.5 |
| 7.008 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 26.0 | 221.6 | 167.2 | 3.3 | 2.5 | 4.4 | 3.3 | 6.6 | 5.0 |
| 7.009 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 388.3 | 3308.0 | 2496.8 | 49.6 | 37.5 | 66.2 | 49.9 | 99.2 | 74.9 |
| 7.012 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 167.8 | 1429.8 | 1079.2 | 21.4 | 16.2 | 28.6 | 21.6 | 42.9 | 32.4 |
| 7.015 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 0.4 | 3.4 | 2.6 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 |
| 7.018 | 94-97 |  | 21 | 1342 | 768.9 | 549.4 | 231.2 | 1777.3 | 1363.3 | 26.7 | 20.4 | 35.5 | 27.3 | 53.3 | 40.9 |
| 7.019 | 97 |  | 1 | 49 | 1308.2 | 0.0 | 71.7 | 938.0 | 0.0 | 14.1 | 0.0 | 18.8 | 0.0 | 28.1 | 0.0 |
| 7.020 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 35.2 | 299.6 | 226.2 | 4.5 | 3.4 | 6.0 | 4.5 | 9.0 | 6.8 |
| 7.021 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 27.6 | 234.9 | 177.3 | 3.5 | 2.7 | 4.7 | 3.5 | 7.0 | 5.3 |
| 7.023 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 31.3 | 266.6 | 201.2 | 4.0 | 3.0 | 5.3 | 4.0 | 8.0 | 6.0 |
| 7.024 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 7.025 | 94-95 |  | 6 | 187 | 816.6 | 192.1 | 167.6 | 1368.3 | 518.4 | 20.5 | 7.8 | 27.4 | 10.4 | 41.0 | 15.6 |

Table 8. (continued)

| $\begin{gathered} \hline \text { PFM } \\ \text { SubArea } \end{gathered}$ | Year | NB | Transect |  | Biomass (g/m) |  | Bed Area (ha) | Bc Biomass (t) |  | Quota 0.2 M Bc (of urchins 90-140 mm TD) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{M}=0.075$ | $\mathrm{M}=0.10$ |  |  |  | $\mathrm{M}=0.15$ |  |
|  |  |  | Number | Length <br> (m) |  |  | Mean | $\begin{gathered} \text { Lower } \\ 90 \% \mathrm{CB} \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ |
| 7.026 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 |  | 2.7 | 22.8 | 17.2 | 0.3 | 0.3 | 0.5 | 0.3 | 0.7 | 0.5 |
| 7.027 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 |  | 297.2 | 2531.8 | 1911.0 | 38.0 | 28.7 | 50.6 | 38.2 | 76.0 | 57.3 |
| 7.028 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 52.2 | 445.0 | 335.9 | 6.7 | 5.0 | 8.9 | 6.7 | 13.4 | 10.1 |
| 7.031 | 95 |  | 1 | 17 | 242.6 | 0.0 | 0.4 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 7.032 | 94-01 | a | 37 | 1819 | 851.9 | 643.0 | 126.4 | 1076.6 | 812.6 | 16.1 | 12.2 | 21.5 | 16.3 | 32.3 | 24.4 |
| 8.001 | 01 | a | 4 | 136 | 524.2 | 54.1 | 126.8 | 664.9 | 68.6 | 10.0 | 1.0 | 13.3 | 1.4 | 19.9 | 2.1 |
| 8.002 | 01 | a | 4 | 136 | 524.2 | 54.1 | 122.4 | 641.6 | 66.2 | 9.6 | 1.0 | 12.8 | 1.3 | 19.2 | 2.0 |
| 8.003 | 01 | a | 4 | 136 | 524.2 | 54.1 | 7.1 | 37.0 | 3.8 | 0.6 | 0.1 | 0.7 | 0.1 | 1.1 | 0.1 |
| 8.004 | 01 | y | 4 | 136 | 524.2 | 54.1 | 143.4 | 751.7 | 77.6 | 11.3 | 1.2 | 15.0 | 1.6 | 22.6 | 2.3 |
| 8.016 | 01 |  | 3 | 115 | 620.0 | 15.3 | 51.4 | 318.8 | 7.9 | 4.8 | 0.1 | 6.4 | 0.2 | 9.6 | 0.2 |
| 9.001 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 104.8 | 893.0 | 674.0 | 13.4 | 10.1 | 17.9 | 13.5 | 26.8 | 20.2 |
| 9.002 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 62.4 | 531.9 | 401.5 | 8.0 | 6.0 | 10.6 | 8.0 | 16.0 | 12.0 |
| 9.003 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 0.4 | 3.4 | 2.6 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 |
| 9.010 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 6.2 | 53.0 | 40.0 | 0.8 | 0.6 | 1.1 | 0.8 | 1.6 | 1.2 |
| 9.011 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 9.3 | 79.6 | 60.1 | 1.2 | 0.9 | 1.6 | 1.2 | 2.4 | 1.8 |
| 9.012 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 33.1 | 282.4 | 213.1 | 4.2 | 3.2 | 5.6 | 4.3 | 8.5 | 6.4 |
| 10.001 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 107.1 | 912.4 | 688.7 | 13.7 | 10.3 | 18.2 | 13.8 | 27.4 | 20.7 |
| 10.002 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 238.4 | 2031.0 | 1532.9 | 30.5 | 23.0 | 40.6 | 30.7 | 60.9 | 46.0 |
| 10.003 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 186.0 | 1584.2 | 1195.8 | 23.8 | 17.9 | 31.7 | 23.9 | 47.5 | 35.9 |
| 10.004 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 134.4 | 1144.6 | 864.0 | 17.2 | 13.0 | 22.9 | 17.3 | 34.3 | 25.9 |
| 10.005 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 3.2 | 27.4 | 20.7 | 0.4 | 0.3 | 0.5 | 0.4 | 0.8 | 0.6 |
| 10.006 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 0.9 | 7.5 | 5.7 | 0.1 | 0.1 | 0.2 | 0.1 | 0.2 | 0.2 |
| 10.008 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 0.7 | 6.2 | 4.6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.1 |
| 10.011 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 2.7 | 23.1 | 17.4 | 0.3 | 0.3 | 0.5 | 0.3 | 0.7 | 0.5 |
| 10.012 | 94-01 | c | 37 | 1819 | 851.9 | 643.0 | 12.4 | 105.6 | 79.7 | 1.6 | 1.2 | 2.1 | 1.6 | 3.2 | 2.4 |
| 101.001 | 94 | d | 26 | 4522 | 823.3 | 695.2 | 33.5 | 275.5 | 232.6 | 4.1 | 3.5 | 5.5 | 4.7 | 8.3 | 7.0 |
| 101.002 | 94 | d | 26 | 4522 | 823.3 | 695.2 | 91.1 | 750.4 | 633.7 | 11.3 | 9.5 | 15.0 | 12.7 | 22.5 | 19.0 |
| 101.006 | 94 | d | 26 | 4522 | 823.3 | 695.2 | 41.8 | 343.9 | 290.4 | 5.2 | 4.4 | 6.9 | 5.8 | 10.3 | 8.7 |
| 101.007 | 94 | d | 26 | 4522 | 823.3 | 695.2 | 42.0 | 345.4 | 291.7 | 5.2 | 4.4 | 6.9 | 5.8 | 10.4 | 8.8 |
| 102.002 | 95 | e | 8 | 678 | 418.0 | 250.1 | 0.4 | 1.5 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 105.001 | 97 | f | 28 | 3722 | 477.0 | 395.2 | 215.4 | 1027.3 | 851.2 | 15.4 | 12.8 | 20.5 | 17.0 | 30.8 | 25.5 |
| 105.002 | 97 | f | 28 | 3722 | 477.0 | 395.2 | 18.2 | 87.0 | 72.1 | 1.3 | 1.1 | 1.7 | 1.4 | 2.6 | 2.2 |
| 106.001 | 94-00 | a | 7 | 544 | 1324.3 | 730.5 | 10.2 | 135.2 | 74.6 | 2.0 | 1.1 | 2.7 | 1.5 | 4.1 | 2.2 |
| 106.002 | 94 |  | 7 | 544 | 1324.3 | 730.5 | 870.9 | 11533.7 | 6015.2 | 173.0 | 90.2 | 230.7 | 120.3 | 346.0 | 180.5 |
| 142.001 | 95 | g | 11 | 978 | 950.8 | 632.8 | 10.1 | 95.6 | 63.6 | 1.4 | 1.0 | 1.9 | 1.3 | 2.9 | 1.9 |
| 142.002 | 95 | g | 11 | 978 | 950.8 | 632.8 | 93.0 | 884.6 | 588.8 | 13.3 | 8.8 | 17.7 | 11.8 | 26.5 | 17.7 |
|  |  |  |  |  | N. Coast | Total: | 29161.8 | 203221.2 | 130405.3 | 3048.3 | 1956.1 | 4064.4 | 2608.1 | 6096.6 | 3912.2 |

Table 8. (continued)

| $\begin{gathered} \hline \text { PFM } \\ \text { SubArea } \end{gathered}$ | Year | NB | Transect |  | Biomass (g/m2) |  | $\begin{gathered} \text { Bed } \\ \text { Area (ha) } \end{gathered}$ | Bc Biomass (t) |  | Quota 0.2 M Bc (of urchins 90-140 mm TD) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{M}=0.075$ | $\mathrm{M}=0.10$ |  |  |  | $\mathrm{M}=0.15$ |  |
|  |  |  | Number | Length <br> (m) |  |  | Mean | $\begin{gathered} \text { Lower } \\ 90 \% \mathrm{CB} \end{gathered}$ | Mean | Lower $90 \%$ CB | Mean | Lower $90 \%$ CB | Mean | Lower 90\% CB | Mean | Lower $90 \%$ CB |


| South Coast of BC: |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11.001 | 96 | a | 2 | 123 | 259.3 | 14.8 | 29.9 | 77.6 | 4.4 | 1.2 | 0.1 | 1.6 | 0.1 | 2.3 | 0.1 |
| 11.002 | 96 |  | 2 | 123 | 259.3 | 14.8 | 135.8 | 352.2 | 20.1 | 5.3 | 0.3 | 7.0 | 0.4 | 10.6 | 0.6 |
| 12.001 | 99 |  | 3 | 91 | 921.0 | 0.0 | 20.7 | 190.7 | 0.0 | 2.9 | 0.0 | 3.8 | 0.0 | 5.7 | 0.0 |
| 12.002 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 4.1 | 27.6 | 21.5 | 0.4 | 0.3 | 0.6 | 0.4 | 0.8 | 0.6 |
| 12.003 | 94-01 |  | 11 | 357 | 373.1 | 166.6 | 91.6 | 341.7 | 191.7 | 5.1 | 2.9 | 6.8 | 3.8 | 10.3 | 5.8 |
| 12.004 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 0.4 | 2.5 | 1.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| 12.005 | 94 |  | 2 | 106 | 138.3 | 0.0 | 56.7 | 78.4 | 0.0 | 1.2 | 0.0 | 1.6 | 0.0 | 2.4 | 0.0 |
| 12.006 | 94 |  | 1 | 35 | 476.3 | 0.0 | 20.2 | 96.0 | 0.0 | 1.4 | 0.0 | 1.9 | 0.0 | 2.9 | 0.0 |
| 12.007 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 35.4 | 239.0 | 186.1 | 3.6 | 2.8 | 4.8 | 3.7 | 7.2 | 5.6 |
| 12.008 | 94-01 | y | 51 | 2234 | 675.5 | 526.0 | 401.4 | 2711.7 | 2111.7 | 40.7 | 31.7 | 54.2 | 42.2 | 81.4 | 63.3 |
| 12.009 | 94 |  | 2 | 170 | 1253.6 | 248.0 | 5.2 | 65.6 | 13.0 | 1.0 | 0.2 | 1.3 | 0.3 | 2.0 | 0.4 |
| 12.010 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 0.3 | 2.3 | 1.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| 12.011 | 94 |  | 4 | 150 | 1796.7 | 262.4 | 134.3 | 2413.5 | 352.5 | 36.2 | 5.3 | 48.3 | 7.1 | 72.4 | 10.6 |
| 12.012 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 89.3 | 603.0 | 469.6 | 9.0 | 7.0 | 12.1 | 9.4 | 18.1 | 14.1 |
| 12.013 | 00 |  | 9 | 381 | 654.0 | 303.6 | 31.6 | 206.5 | 113.2 | 3.1 | 1.7 | 4.1 | 2.3 | 6.2 | 3.4 |
| 12.014 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 121.1 | 818.3 | 637.2 | 12.3 | 9.6 | 16.4 | 12.7 | 24.5 | 19.1 |
| 12.015 | 94 |  | 1 | 34 | 691.8 | 0.0 | 36.3 | 251.0 | 0.0 | 3.8 | 0.0 | 5.0 | 0.0 | 7.5 | 0.0 |
| 12.016 | 94 |  | 5 | 205 | 899.5 | 242.5 | 112.0 | 1007.5 | -302.3 | 15.1 | -4.5 | 20.2 | -6.0 | 30.2 | -9.1 |
| 12.017 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 45.9 | 310.3 | 241.6 | 4.7 | 3.6 | 6.2 | 4.8 | 9.3 | 7.2 |
| 12.018 | 94 |  | 5 | 362 | 361.1 | 162.6 | 282.3 | 1019.2 | -278.5 | 15.3 | -4.2 | 20.4 | -5.6 | 30.6 | -8.4 |
| 12.019 | 94 |  | 1 | 79 | 498.7 | 0.0 | 57.5 | 286.8 | 0.0 | 4.3 | 0.0 | 5.7 | 0.0 | 8.6 | 0.0 |
| 12.020 | 94 |  | 1 | 59 | 1087.4 | 0.0 | 3.6 | 38.9 | 0.0 | 0.6 | 0.0 | 0.8 | 0.0 | 1.2 | 0.0 |
| 12.021 | 94-01 |  | 4 | 121 | 630.6 | 31.0 | 15.1 | 95.1 | 4.7 | 1.4 | 0.1 | 1.9 | 0.1 | 2.9 | 0.1 |
| 12.039 | 95 |  | 1 | 31 | 334.5 | 0.0 | 15.4 | 51.4 | 0.0 | 0.8 | 0.0 | 1.0 | 0.0 | 1.5 | 0.0 |
| 12.041 | 94-01 | a | 51 | 2234 | 675.5 | 526.0 | 29.9 | 201.7 | 157.1 | 3.0 | 2.4 | 4.0 | 3.1 | 6.1 | 4.7 |
| 13.001 | 99 | a | 9 | 599 | 732.5 | 419.5 | 117.2 | 858.2 | 491.5 | 12.9 | 7.4 | 17.2 | 9.8 | 25.7 | 14.7 |
| 13.002 | 99 | a | 9 | 599 | 732.5 | 419.5 | 169.5 | 1241.7 | 711.1 | 18.6 | 10.7 | 24.8 | 14.2 | 37.3 | 21.3 |
| 13.003 | 99 | a | 9 | 599 | 732.5 | 419.5 | 6.5 | 47.4 | 27.2 | 0.7 | 0.4 | 0.9 | 0.5 | 1.4 | 0.8 |
| 13.006 | 99 | a | 9 | 599 | 732.5 | 419.5 | 5.1 | 37.0 | 21.2 | 0.6 | 0.3 | 0.7 | 0.4 | 1.1 | 0.6 |
| 13.007 | 99 | a | 9 | 599 | 732.5 | 419.5 | 11.0 | 80.4 | 46.1 | 1.2 | 0.7 | 1.6 | 0.9 | 2.4 | 1.4 |
| 13.008 | 99 | a | 9 | 599 | 732.5 | 419.5 | 3.8 | 27.7 | 15.8 | 0.4 | 0.2 | 0.6 | 0.3 | 0.8 | 0.5 |
| 13.009 | 99 | a | 9 | 599 | 732.5 | 419.5 | 31.6 | 231.6 | 132.6 | 3.5 | 2.0 | 4.6 | 2.7 | 6.9 | 4.0 |
| 13.010 | 99 | a | 9 | 599 | 732.5 | 419.5 | 10.6 | 77.3 | 44.3 | 1.2 | 0.7 | 1.5 | 0.9 | 2.3 | 1.3 |
| 13.011 | 99 | a | 9 | 599 | 732.5 | 419.5 | 1.6 | 12.0 | 6.9 | 0.2 | 0.1 | 0.2 | 0.1 | 0.4 | 0.2 |
|  |  |  |  |  |  |  |  | 39 |  |  |  |  |  |  |  |

Table 8. (continued)

| $\begin{gathered} \text { PFM } \\ \text { SubArea } \end{gathered}$ | Year | NB | Transect |  | Biomass (g/m2) |  | Bed <br> Area (ha) | Bc Biomass (t) |  | Quota 0.2 M Bc (of urchins 90-140 mm TD) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{M}=0.075$ | $\mathrm{M}=0.10$ |  |  |  | $\mathrm{M}=0.15$ |  |
|  |  |  | Number | Length <br> (m) |  |  | Mean | $\begin{gathered} \text { Lower } \\ 90 \% \mathrm{CB} \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \mathrm{CB} \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \end{gathered}$ | Mean | $\begin{gathered} \hline \text { Lower } \\ 90 \% \text { CB } \\ \hline \end{gathered}$ |
| 13.012 | 99 | a | 9 | 599 | 732.5 | 419.5 |  | 30.8 | 225.7 | 129.3 | 3.4 | 1.9 | 4.5 | 2.6 | 6.8 | 3.9 |
| 13.014 | 99 | a | 9 | 599 | 732.5 | 419.5 |  | 15.0 | 110.1 | 63.1 | 1.7 | 0.9 | 2.2 | 1.3 | 3.3 | 1.9 |
| 13.018 | 99 | a | 9 | 599 | 732.5 | 419.5 | 3.8 | 28.1 | 16.1 | 0.4 | 0.2 | 0.6 | 0.3 | 0.8 | 0.5 |
| 13.023 | 99 | a | 9 | 599 | 732.5 | 419.5 | 3.1 | 22.6 | 13.0 | 0.3 | 0.2 | 0.5 | 0.3 | 0.7 | 0.4 |
| 13.025 | 99 | a | 9 | 599 | 732.5 | 419.5 | 106.2 | 777.8 | 445.4 | 11.7 | 6.7 | 15.6 | 8.9 | 23.3 | 13.4 |
| 13.026 | 99 | a | 9 | 599 | 732.5 | 419.5 | 10.2 | 75.0 | 42.9 | 1.1 | 0.6 | 1.5 | 0.9 | 2.2 | 1.3 |
| 13.028 | 99 | a | 9 | 599 | 732.5 | 419.5 | 123.6 | 905.2 | 518.4 | 13.6 | 7.8 | 18.1 | 10.4 | 27.2 | 15.6 |
| 13.029 | 99 | a | 9 | 599 | 732.5 | 419.5 | 4.3 | 31.8 | 18.2 | 0.5 | 0.3 | 0.6 | 0.4 | 1.0 | 0.5 |
| 13.030 | 99 |  | 2 | 50 | 486.8 | 0.0 | 1.7 | 8.4 | 0.0 | 0.1 | 0.0 | 0.2 | 0.0 | 0.3 | 0.0 |
| 13.031 | 99 |  | 1 | 41 | 439.0 | 0.0 | 26.5 | 116.3 | 0.0 | 1.7 | 0.0 | 2.3 | 0.0 | 3.5 | 0.0 |
| 13.032 | 99 |  | 1 | 145 | 574.6 | 75.5 | 35.9 | 206.5 | 27.2 | 3.1 | 0.4 | 4.1 | 0.5 | 6.2 | 0.8 |
| 13.033 | 99 |  | 5 | 363 | 862.6 | 389.1 | 45.8 | 394.7 | 208.8 | 5.9 | 3.1 | 7.9 | 4.2 | 11.8 | 6.3 |
| 13.035 | 99 | a | 9 | 599 | 732.5 | 419.5 | 6.9 | 50.6 | 29.0 | 0.8 | 0.4 | 1.0 | 0.6 | 1.5 | 0.9 |
| 13.039 | 99 | a | 9 | 599 | 732.5 | 419.5 | 24.4 | 178.4 | 102.2 | 2.7 | 1.5 | 3.6 | 2.0 | 5.4 | 3.1 |
| 13.041 | 99 | a | 9 | 599 | 732.5 | 419.5 | 2.1 | 15.4 | 8.8 | 0.2 | 0.1 | 0.3 | 0.2 | 0.5 | 0.3 |
| 14.005 | 99 |  | 2 | 202 | 152.2 | 40.2 | 16.0 | 24.4 | 6.4 | 0.4 | 0.1 | 0.5 | 0.1 | 0.7 | 0.2 |
| 14.007 | 99 |  | 2 | 244 | 354.7 | 117.2 | 20.7 | 73.4 | 24.3 | 1.1 | 0.4 | 1.5 | 0.5 | 2.2 | 0.7 |
| 14.008 | 99 |  | 2 | 202 | 491.0 | 129.7 | 23.8 | 116.8 | 30.9 | 1.8 | 0.5 | 2.3 | 0.6 | 3.5 | 0.9 |
| 14.009 | 99 |  | 4 | 866 | 287.5 | 185.3 | 123.0 | 353.8 | 228.0 | 5.3 | 3.4 | 7.1 | 4.6 | 10.6 | 6.8 |
| 14.011 | 99 |  | 5 | 503 | 147.6 | 78.8 | 59.3 | 87.6 | -32.5 | 1.3 | -0.5 | 1.8 | -0.6 | 2.6 | -1.0 |
| 14.012 | 99 |  | 1 | 101 | 301.4 | 0.0 | 56.5 | 170.2 | 0.0 | 2.6 | 0.0 | 3.4 | 0.0 | 5.1 | 0.0 |
| 14.013 | 99 | a | 16 | 2118 | 269.2 | 208.0 | 12.7 | 34.3 | 26.5 | 0.5 | 0.4 | 0.7 | 0.5 | 1.0 | 0.8 |
| 17.002 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 21.5 | 21.1 | 16.1 | 0.3 | 0.2 | 0.4 | 0.3 | 0.6 | 0.5 |
| 17.003 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 95.3 | 93.6 | 71.7 | 1.4 | 1.1 | 1.9 | 1.4 | 2.8 | 2.1 |
| 17.008 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 15.1 | 14.8 | 11.4 | 0.2 | 0.2 | 0.3 | 0.2 | 0.4 | 0.3 |
| 17.010 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 43.4 | 42.6 | 32.6 | 0.6 | 0.5 | 0.9 | 0.7 | 1.3 | 1.0 |
| 17.017 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 7.9 | 7.7 | 5.9 | 0.1 | 0.1 | 0.2 | 0.1 | 0.2 | 0.2 |
| 18.001 | 98-99 | y | 21 | 1983 | 98.2 | 75.2 | 20.7 | 20.4 | 15.6 | 0.3 | 0.2 | 0.4 | 0.3 | 0.6 | 0.5 |
| 18.002 | 98 |  | 3 | 165 | 78.2 | 14.5 | 77.7 | 60.7 | 11.3 | 0.9 | 0.2 | 1.2 | 0.2 | 1.8 | 0.3 |
| 18.003 | 98 |  | 1 | 45 | 93.3 | 0.0 | 31.8 | 29.6 | 0.0 | 0.4 | 0.0 | 0.6 | 0.0 | 0.9 | 0.0 |
| 18.004 | 99 |  | 4 | 392 | 39.3 | 18.6 | 78.5 | 30.9 | 14.6 | 0.5 | 0.2 | 0.6 | 0.3 | 0.9 | 0.4 |
| 18.005 | 98-99 | y | 21 | 1983 | 98.2 | 75.2 | 356.5 | 350.2 | 267.9 | 5.3 | 4.0 | 7.0 | 5.4 | 10.5 | 8.0 |
| 18.006 | 99 |  | 9 | 1081 | 127.4 | 86.9 | 344.6 | 439.0 | 165.6 | 6.6 | 2.5 | 8.8 | 3.3 | 13.2 | 5.0 |
| 18.011 | 98 |  | 1 | 53 | 462.9 | 0.0 | 96.8 | 447.8 | 0.0 | 6.7 | 0.0 | 9.0 | 0.0 | 13.4 | 0.0 |
| 19.003 | 98-99 | h | 21 | 1983 | 98.2 | 75.2 | 111.5 | 109.5 | 83.8 | 1.6 | 1.3 | 2.2 | 1.7 | 3.3 | 2.5 |
| 40 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 8. (continued)


New bed areas estimated from 2001 Price Island Broadbrush Survey:

| 7.001 N | 01 | k | 5 | 119 | 965.8 | 698.3 | 33.8 | 326.4 | 236.0 | 4.9 | 3.5 | 6.5 | 4.7 | 9.8 | 7.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.002 N | 01 | k | 9 | 247 | 1594.4 | 1083.4 | 53.5 | 853.0 | 579.6 | 12.8 | 8.7 | 17.1 | 11.6 | 25.6 | 17.4 |
| 7.031 S | 01 | k | 10 | 306 | 908.1 | 586.7 | 60.0 | 544.9 | 352.0 | 8.2 | 5.3 | 10.9 | 7.0 | 16.3 | 10.6 |


| BC | Total: | 35768.3 | 233971.5 | 143823.4 | 3509.6 | 2157.4 | 4679.4 | 2876.5 | 7019.1 | 4314.7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

Table 9. Total quota (tonnes) options for the red sea urchin fishery by north and south BC, estimated from various natural mortality values applied to current biomass $(\mathrm{Bc})$ calculated from mean and approximate $90 \%$ lower confidence bound (CB) of the clipped biomass values for eight size limits of commercial red sea urchin, and bed areas fished from 1997-2000.

| Size Limit (mm TD) | Region | Quota 0.2 M Bc |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\mathrm{M}=0.075$ |  | $\mathrm{M}=0.10$ |  | $\mathrm{M}=0.15$ |  |
|  |  | Mean | Lower $90 \%$ CB | Mean | Lower $90 \%$ CB | Mean | Lower $90 \%$ CB |
| $\geq 90$ | North Coast | 3553.9 | 2303.4 | 4738.5 | 3071.2 | 7107.7 | 4606.9 |
|  | South Coast | 535.3 | 226.0 | 713.8 | 301.3 | 1070.7 | 451.9 |
|  | BC | 4089.2 | 2529.4 | 5452.2 | 3372.5 | 8178.4 | 5058.8 |
| $\geq 95$ | North Coast | 3333.5 | 2145.4 | 4444.6 | 2860.6 | 6666.9 | 4290.9 |
|  | South Coast | 509.0 | 213.9 | 678.7 | 285.2 | 1018.1 | 427.8 |
|  | BC | 3842.5 | 2359.3 | 5123.3 | 3145.7 | 7685.0 | 4718.6 |
| $\geq 100$ | North Coast | 3082.1 | 1953.9 | 4109.5 | 2605.2 | 6164.2 | 3907.9 |
|  | South Coast | 474.2 | 190.4 | 632.3 | 253.9 | 948.4 | 380.8 |
|  | BC | 3556.3 | 2144.3 | 4741.8 | 2859.1 | 7112.7 | 4288.7 |
| 90-130 | North Coast | 2577.0 | 1699.7 | 3436.0 | 2266.3 | 5154.0 | 3399.5 |
|  | South Coast | 351.2 | 158.9 | 468.3 | 211.9 | 702.4 | 317.9 |
|  | BC | 2928.2 | 1858.7 | 3904.3 | 2478.2 | 5856.4 | 3717.3 |
| 90-140 | North Coast | 3074.2 | 1973.6 | 4098.9 | 2631.5 | 6148.4 | 3947.2 |
|  | South Coast | 435.4 | 183.8 | 580.5 | 245.0 | 870.8 | 367.5 |
|  | BC | 3509.6 | 2157.4 | 4679.4 | 2876.5 | 7019.1 | 4314.7 |
| 95-130 | North Coast | 2356.6 | 1548.3 | 3142.1 | 2064.4 | 4713.2 | 3096.6 |
|  | South Coast | 324.9 | 148.4 | 433.2 | 197.9 | 649.8 | 296.9 |
|  | BC | 2681.5 | 1696.8 | 3575.3 | 2262.3 | 5363.0 | 3393.5 |
| 95-140 | North Coast | 2853.8 | 1818.9 | 3805.1 | 2425.1 | 5707.6 | 3637.7 |
|  | South Coast | 409.1 | 172.1 | 545.5 | 229.5 | 818.2 | 344.2 |
|  | BC | 3262.9 | 1991.0 | 4350.5 | 2654.6 | 6525.8 | 3981.9 |
| 100-140 | North Coast | 2602.4 | 1629.1 | 3469.9 | 2172.2 | 5204.9 | 3258.3 |
|  | South Coast | 374.3 | 149.7 | 499.0 | 199.5 | 748.6 | 299.3 |
|  | BC | 2976.7 | 1778.8 | 3969.0 | 2371.7 | 5953.4 | 3557.6 |



Fig. 1. Size frequencies of red sea urchins measured (A) at processing plants during the 2000-2001 fishing season, and (B) on population surveys during 1994-2001 throughout B.C.


Fig. 2. Cumulative size frequencies, expressed as a proportion of the total red sea urchins ( $\geq 90 \mathrm{~mm} \mathrm{TD}$ ) sampled from the processing plants ( $\mathrm{N}=22,546$ ) during the 2000-2001 fishing season, and from population surveys during ( $\mathrm{N}=34,098$ ) during 1994-2001 throughout B.C.




Fig. 3. Comparison of (A) means, (B) standard error of mean, and (C) lower $90 \%$ confidence bounds of $95-130 \mathrm{~mm}$ TD red sea urchin density (biomass, $\mathrm{g} / \mathrm{m}^{2}$ ) from unclipped (all transects) and clipped (only transects lying within bed boundaries) survey data by PFM subarea in B.C.

## APPENDIX

PSARC INVERTEBRATE SUBCOMMITTEE
Request for Working Paper
Red Urchin Size Implications

## Date Submitted:

Individual or group requesting advice: Resource Management/PUHA (Fisheries Manager/Biologist, Science, SWG, PSARC, Industry, Other stakeholder etc.)

Proposed PSARC Presentation Date: December 2001

Subject of Paper (title if developed): Size frequency of commercially harvested red urchins and potential implications on quota assessments; a summary of the red urchin sampling program

## Stock Assessment Lead Author: Alan Campbell, Dimitri Tzotzos

Fisheries Management Author/Reviewer: Juanita Rogers, Laurie Convey, Rick Harbo

## Rationale for request:

(What is the issue, what will it address, importance, etc.)
Quotas in the red urchin fishery are currently presented to managers for a size range of animals from 100 to 140 mm test diameter. Stock Assessment has had some concerns that this full range of sizes is not fished in the commercial fishery and that the fishery focuses on animals closer to 100 mm (best market value). In this event, the concern would be that the true exploitation rate could equate to something much higher than the recommended rate of 2-3\%.

A port sampling program was initiated in 1999 in the North Coast, in order to start gathering data with which to compare the size classes of urchins from the commercial fishery ("actual") to the size range on which quota calculations ("theoretical") are based.

When the minimum size limit was changed from 100 mm to 90 mm for the 2000/2001 management plan, a more extensive, and plant-based, sampling program was initiated to assess the size and quality of urchins harvested, with the objective to compare the theoretical versus actual harvest sizes and implications for exploitation rates.

## Question(s) to be addressed in the Working Paper: (To be developed by initiator)

Present a summary of all port and plant sampling data undertaken to date.
Is there a significant difference between the theoretical size range of urchins (100 to 140 mm test diameter) and the actual size range fished?

If so, does this difference have an impact on the quotas presented to managers in CSAS 99/201?

What are the future recommendations for this sampling program (e.g. sampling protocols, objectives)?

What are the quota recommendations for upcoming annual management planning process?

## Objective of Working Paper: <br> (To be developed by FM \& StAD for internal papers)

To compare the size range of urchins harvested in the commercial fishery to the size range on which quotas are based $(100-140 \mathrm{~mm})$ and to determine the implications for management of the fishery and the setting of quotas.

If the comparison indicates significant discrepancy between the actual and theoretical size range of harvest, to calculate quotas for the red urchin fishery based on the actual size range fished.

## Stakeholders Affected:

PUHA, First Nations ( $2 \%$ of TAC designated for food use)

## How Advice May Impact the Development of a Fishing Plan:

If StAD feels that the actual range of animals being fished is significantly different from the theoretical, it is likely that a reduction in quota will be required. Impacts on the commercial industry could be significant.

## Timing Issues Related to When Advice is Necessary

If advice is to be incorporated into an upcoming annual fishing plan, it must be presented in the winter PSARC session, in order to be used in the spring management planning process.


[^0]:    * Less than 500 kg
    ${ }^{\text {a }}$ Data for each year cannot be published separately.
    Mandatory log book under Z licence came into effect in 1983.
    ${ }^{\text {c }}$ Sales slips were combined for red and green sea urchins in 1987, were later separated by price criteria, but 320 t remains missing in area table
    ${ }^{4}$ Includes landings from Area 111.
    ${ }^{\mathrm{e}}$ Includes landings from Area 123.
    ${ }^{t}$ Includes landings from Area 124.
    ${ }^{8}$ Change in licencing from calender year to market-driven year. 1997/98 season ran January 1/97 to June 30/98.
    After 1997 licencing year continued to be July to June

[^1]:    ${ }^{\text {a }}$ Survey transects did not fall within clipped 1997-2000 commercial red sea urchin beds. Beds were estimated using survey data.

[^2]:    ${ }^{\text {a }}$ Survey transects did not fall within clipped 1997-2000 commercial red sea urchin beds. Beds were estimated using survey data.

