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Flatfish Stock Assessments for the west coast of Canada for 1997 and recommended yield options for 1998

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#### Abstract

Interim assessments were prepared for important stocks of flatfish caught in the B.C. trawl fishery. In the past CPUE from the commercial trawl fishery has been used as a surrogate for stock abundance for the assessments for Area 5A-B rock sole and Area 3C-D and 5C-E Dover sole. However, the behaviour of the fishing fleet has been dramatically altered by changing trip limits and other regulatory measures in recent years and CPUE can no longer be considered a reliable index of relative abundance. Only summaries of the landing statistics are presented for the stocks listed above. Until more comprehensive assessments for these stocks are undertaken the yield options will remain the same as those recommended for the 1997 fishery. The assessments for Area 5C-D rock sole and English sole stocks are based on catchage analysis and include biological samples collected during the 1996 fishery. Assessments were produced for four species and six stocks.

Petrale sole stocks are at low abundance. Rock sole biomass in Areas 5C-D in 1996 was above the long term average for the last 50 years. The estimate of fishing mortality for the stock in 1996 was below $\mathrm{F}_{0.1}$. The yield recommendations for this stock in 1998 are unchanged from those recommended for 1997. English sole biomass in Areas 5C-D in 1996 was slightly above the long term average for the last 50 years. The estimate of fishing mortality for the stock in 1996 was below $\mathrm{F}_{0.1}$ as well. The yield recommendations for this stock for 1998 remain the same as those recommended for 1997. Landings of Dover sole from Areas 5C-E in 1996 were near the longterm sustainable harvest for that stock while effort in 1996 was close to the optimum corresponding to MSY.


## Résumé

Des évaluations provisoires ont été préparées pour d'importants stocks de poissons plats faisant l'objet de la pêche au chalut en C.-B. Les PUE de la pêche commerciale au chalut étaient antérieurement utilisés comme indices de l'abondance du stock aux fins des évaluations de la fausse limande des zones 5A-B et de la limande-sole des zones 3C-D et $5 C-E$. Les modifications apportées aux limites par sortie et d'autres mesures réglementaires adoptées ces dernières années ont fortement modifié le compırtement de la flottille de pêche de sorte que les PUE ne constituent plus des indices: fiables de l'abondance relative. Les statistiques sur les débarquements des stocks mentionnés plus haut ne sont présentées que sous forme résumée. Tant que des évaluations plus détaillées n'auront pas été réalisées, les options de rendement demeureront les mêmes; que celles recommandées pour la pêche de 1997. Les évaluations des stocks de fausse limande et de carlottin anglais de 5C-D sont fondées sur l'analyse des prises selon l'âge et des échantillons biologiques prélevés pendant la pêche de 1996. Des évaluations ont été faites pour quatre espèces et six stocks.

L'abondance des stocks de plie de Californie est faible. La biomasse de la fausse limande en 5C-D en 1996 était supérieure à la moyenne à la long terme des 50 dernières années au moins. La mortalité par pêche estimée du stock était inférieure au $\mathrm{F}_{0,1}$ en 1996. Les recommandations relatives au rendement de ce stock en 1998 demeurent inchangies par rapport à celles de 1997. La biomasse du carlottin anglais en 5C-D en 1995 était légèrement supérieure à la moyenne à long terme des 50 dernières années at: moins. La mortalité par pêche estimée du stock en 1996 était aussi inférieure au $\mathrm{F}_{0,1}$. Les recommandations relatives au rendement en 1998 demeurent inchangées par rapport à celles de 1997. Les débarquements de limande-sole en provenance de 5C-E en 1996 approchaient la récolte soutenue à long terme pour ce stock tandis que l'effort en 1996 se rapprochait de la valeur optimale correspondant au RMS.

### 1.0 General Introduction

This year interim assessments have been prepared for all flatfish stocks. Catch and effort statistics have been updated for all stocks to include information from the 1996 fishery. The $25 \%$ qualified median CPUE from the commercial fishery has been used as the index for monitoring stock status for petrale sole (Eopsetta jordani), Area 3C-D Dover sole (Microstomus Pacificus), Area 5A-B rock sole (Pleuronectes bilineata) and Area 5C-E Dover sole (Fargo and Kronlund 1997). Catch-age analysis provides of assessment for Hecate Strait rock sole and Hecate Strait English sole (Pleuronectes vetulus).

The groundfish trawl fishery has changed significantly in recent years. All option A vessels were required to carry an observer on every fishing trip beginning in 1996. Transferable vessel quotas were used to regulate the 1997 fishery. Observers recorded information on area, depth, catch and effort for the trip. Port monitors provided estimates of the landings for each trip. Stock assessment analyses of fishery catch-effort data assume that CPUE is proportional to stock abundance. Cases where this assumption is false far outnumber cases where it is true. Changes in management of groundfish fisheries including the use of trip limits and changes in vessel efficiency and fishing patterns over time have nullified the use of the fishery CPUE index for many cases (Richards and Schnute 1986). Hyperstability of the index has been documented in many situations. That is, in cases where vessel trip limits have been invoked by managers or in the case of multispecies fisheries, the CPUE index will not provide a signal of stock decline until the stock has been depleted (Hillborn and Walters 1992, Richards and Schnute 1986). Accordingly, CPUE is not used as the basis of assessment for any of the cases presented here. In the presence of skewed observations, both the mean of ratios and the ratio of means of CPUE can perform badly; they are sensitive to a small number of outliers. The median or $50 \%$ trimmed mean provides a robust alternative to the two former statistics (Fargo and Kronlund 1997). Accordingly, the median statistic has been presented for each case where time series of catch effort data exist. In the analysis herein year to year changes in the CPUE index, were not considered. The rate of decrease or increase in the index over four or five years is discussed in the context of changing management and results of independent analysis of biological data. To aid in evaluation of the index the $90 \%$ bootstrapped confidence intervals for each CPUE estimate were computed. The bootstrapping was implemented in S-plus version 4.0 (Mathsoft 1997). For each case the bootstrap procedure included resampling 1000 new samples, each of the same size as the observed data. Each new sample was drawn with replacement from the observed data. The median statistic was first calculated using the observed data and then re-calculated using each of the new samples to yield a bootstrap distribution. The resulting replicates were used to calculate the bootstrap estimates of bias, mean and standard error for the median statistic.

Whenever possible, analysis of age composition data was used as the basis for assessment. Catch curves of the most current age composition data were used for 'northern' Dover sole and Petrale sole to estimate the total instantaneous mortality rate for these stocks. For Hecate Strait Rock sole and English sole the state space model of

Schnute and Richards was applied to age composition data time series' to reconst.uct the stock history. Yield for these cases was determined for the $25^{\text {th }}$ and $50^{\text {th }}$ percentiles of the $95 \%$ confidence region for the 1996 biomass estimate using a specific target fishing mortality rate. Yield options for 1997-98 are summarised in Table 4.1.

### 1.1 Coastwide

Yield options are not proposed for flatfish species on a coastwide basis.

### 1.2 Strait of Georgia

Yield options are not proposed for flatfish for this region.

### 1.3 West Coast of Vancouver Island (Areas 3C and 3D)

### 1.3.1. Petrale sole

### 1.3.1.1. Introduction

The petrale sole population off the west coast of Vancouver Island is composed of two stocks based on results from tagging experiments conducted in the 1960s (Ketchen and Forrester 1966), (Pedersen 1975a). The southern stock occupies both the Canadian and U.S. portions of Area 3C, while the northern stock occupies Areas 3D5D. Detailed data for the intensive fishery occurring from 1950 to 1970 is rot available because the fishery was carried out largely by U.S. trawlers and no comprehensive data were collected. Petrale sole recruit to the commercial fishery beginning at age four but recruitment is not knife-edged and fish are not fully recruited until age 8. Length 5 f $50 \%$ maturity, $\mathrm{L}_{50}$, is $38.1 \mathrm{~cm}(7 \mathrm{y})$ for males and $44.3 \mathrm{~cm}(8 \mathrm{y})$ for females (Ketchen and Forrester 1966). Assessment of these stocks is hampered by the lack of effort datia series and a lack of age composition data after 1970. The backlog of age composition data is now being processed. Due to PSARC's concern over the long term decline in landings for these stocks, only incidental landings of this species have been permitted since 1991. The stock catch histories and biological data available are used as the beisis for this assessment.

Previous catch-age analyses by Ketchen and Forrester (1966) and Pedersen (1975b) indicated that 'there appears to be no need for regulation of the summer fishery and the effects of winter fishing on the spawning concentrations appeared to be overshadowed by environmentally induced variations in the production of recruits' (Ketchen 1979). More recently Castillo et al. (1994) showed that offstore Ekman transport of eggs and larvae accounted for $55 \%$, and $65 \%$ of the variation in petrale sole year-class strength in PMFC Areas 2B and 3A, respectively. They concluced, as have previous investigators, that density-independent survival variation at the early life stages is high compared to variation in spawning biomass; thus, environment regulates recruitment for this species. However, in view of the low abundance of these: stocks, we continue to recommend that no target fishery be permitted at this time.

### 1.3.1.2 Landing statistics

Landings for the southern stock decreased slightly to 314 t in 1996 from 353 t in 1995 while landings for the northern stock decreased $67 \%$ to 145 t in 1996 from 446 t in 1995, (Tables 4.2-4.3). The time series of landings for this species shows cyclic fluctuations with peaks occurring about once a decade (Figure 4.1). Fluctuations in landings coincide with recruitment cycles for the species (Ketchen and Forrester 1966, Castillo et al. 1994). For both stocks, landings show an overall decline since the start of the fishery. Regulatory measures are partly responsible for the decline in landings from these stocks since 1985. A trip limit of $40,000 \mathrm{lb}$ was in effect for the first quarter from 1985 to 1991. From 1991 to 1995 a trip limit of $10,000 \mathrm{lb}$ was in effect during the first quarter of the year while in 1996 only incidental catches were permitted, and in 1997 there was a coastwide cap of 479 t .

Effort for both stocks has decreased dramatically over the last four years with no target fishery permitted on the spawning stock. Effort for southern stock decreased to 139 h in 1996 from 414 h in 1995 while effort for the northern stock has remained lower than 100 h for the 1994-96 period. Catch per unit effort for the southern stock increased to $0.233 \mathrm{t} / \mathrm{h}$ in 1996 from $0.120 \mathrm{t} / \mathrm{h}$ in 1995 while CPUE for the northern stock increased to $0.271 \mathrm{t} / \mathrm{h}$ in 1996 from 0.126 in 1995.

### 1.3.1.3 Biological data

The age composition of petrale sole females (the number of males sampled was too low to produce meaningful results) sampled from the 1994 fishery (the most recent data available), is presented in Figure 4.2. The maximum age of 35 y is older than any previous age determinations made from otolith surfaces (maximum age $=25 \mathrm{y}$ ). This is likely due to the use of the burnt otolith cross-section technique employed. Prior to this time otolith surfaces were read. Older fish are under aged by this method (Fargo and Kronlund 1997). The instantaneous total mortality rate, z, was estimated as the slope of the linear regression relationship of $\ln$ (numbers) on age. Estimates for $z$ ranged between 0.15 and 0.18 depending on the range of ages used (Table 4.4 and Figure 4.3). Ketchen and Forrester suggested that M for females was around 0.20 . At that time they believed that maximum age for the species was between 20 and 25 years. Recent age composition data suggests that M , for these stocks is between 0.15 and 0.2 (Fargo 1995). This analysis suggests that the fishing mortality rate for petrale sole in recent years has been negligible. We caution that these estimates of $z$ are preliminary and require corroboration with age composition time series which can be used to estimate the instantaneous total mortality rate for cohorts. In addition, the small sample size ( 233 fish) hampers the utility of this type of analysis.

### 1.3.1.4 Stock status

The significant decrease in fishing effort, on these stocks over the last three years could have a positive effect on stock abundance in the future. If the estimates of total mortality are accurate the stock should have started to rebuild. However it is too
early to tell if the low removals over the last three years will produce a measurable change in stock abundance. Further, evidence suggests that environment mitigates stock abundance. The recent El Nino should produce temperature regime that is favourable for stock production. In any case these stocks remain at a low level. Given, their turnover rate (approximately 20-30 years) any significant change in stock abundance will probably not be discernible for at least a decade.

### 1.3.1.5. Recommendations And Yield options

The catch histories for these stocks indicate that yield decreased substantially after large removals by the U.S. fleet between the mid 1940s and mid 1960s. Age composition data are needed to reconstruct the stock history and examire the effect of spawning biomass and environment on recruitment. This analysis is still several years away.

Precautionary yield option: Managers should continue to permit incidental landings only for these stocks in 1997, as in 1996.

### 1.3.2 Area 3CD Dover sole

### 1.3.2.1 Area 3CD Introduction

In the last assessment for Area 3CD Dover sole, Fargo and Krisnlund (1997) noted a significant decline in catch per unit effort (CPUE) in this fishery since it began in the late 1980s. Although this decline was due in part to a regulatory effect it persisted in every index of CPUE estimated regardless of the qualification le'rel. Vessel participation in this fishery has increased threefold since 1988 and the area over which the fishery takes place has expanded from shallower to deeper depths. However there was little evidence that adjusting the CPUE series for differences in the fishery over time would be of benefit. Significant changes in the age structure of the stock: have also occurred between 1981 and 1995 but it was difficult to attribute this to the fishery.

The Dover sole is a right-eyed flounder that inhabits the Pacific coast of North America from California to the Bering Sea (Hart 1973). It occupies mud-bottom and feeds primarily on benthic invertebrates. Abundance has been shown :o decrease with increasing latitude (Westrheim et al. 1992). Significant commercial quantities of this species occur between California and British Columbia. Results of U.S. adult tagging studies indicate that a number of individual stocks exist along the Pacific coast and that there is minimal intermingling of adults among stocks (Westrheim et al. 1992). This suggests that the population off the west coast of Vancouver Island is probably a discrete stock. Dover sole become vulnerable to the commercial trawl fishery at about 5 years of age but are not fully recruited until age 7-8 (Fargo and Workman 1995). Length of $50 \%$ sexual maturity, $\mathrm{L}_{50}$, is $37.1 \mathrm{~cm}(4-5 \mathrm{y})$ for males and 39.5 cm ( 8 y ) for females. Little north-south movement of adults has been observed, although they dc uncertake bathymetric migrations from shallow ( $140-200 \mathrm{~m}$ ) to deep ( $400-800 \mathrm{~m}$ ) water for spawning (Westrheim et al. 1992). Adults spawn over a six month season (December-

May) and spawning is age specific with older fish spawning earlier than younger fish (Hunter et al. 1992). The larvae of this species undergo a prolonged pelagic phase offshore that can last as long as two years. Thus, the larvae of different stocks could intermingle extensively. The growth rate for this species is relatively slow (von Bertalanffy $\mathrm{k}=0.12$ for males and 0.09 for females (Fargo and Workman 1995) and maximum age is 52 years among (Westrheim et al. 1992). The maximum age for Area 3CD Dover sole estimated from biological samples collected to date is 49 years (Fargo and Workman 1995).

### 1.3.2.2 Area 3CD Management history

The Area 3CD Dover sole fishery was unregulated prior to 1992. In 1992, a $20,000 \mathrm{lb}(9 \mathrm{t})$ trip limit was invoked after $70 \%$ of the quota was caught. Since 1992, variable trip limits less than $50,000 \mathrm{lbs}$ ( 23 tons) have been used to manage the fishery within the recommended yield. During the period from 1988 to 1996, trips where less than $50,000 \mathrm{lb}(23 \mathrm{t})$ of Dover sole were landed accounted for $60 \%-95 \%$ of the total landings from this area. Trips greater than $50,000 \mathrm{lb}(23 \mathrm{t})$ were not permitted after 1993 and in 1996 vessel quotas were imposed on the trawl fleet. Commercial catch and effort data and results from research surveys of this resource in 1981 and 1995 provide the basis for assessment of this stock.

### 1.3.2.3 Area 3CD Commercial catch and effort data

Annual catch and effort statistics for Area 3CD Dover sole are presented in Table 4.5 and Figure 4.4. CPUE and effort for 1996 are not directly comparable to the observations for previous years because the estimates are determined from observations made by at-sea observers while estimates for previous years are determined from logbook information recorded by the vessel captains (Kronlund and Fargo 1997). Landings of Area 3CD Dover sole decreased to 1083 t in 1996 from 1630 t in 1995 while effort decreased to 2318 h in 1996 from 5352 h in 1995. CPUE declined to $0.229 \mathrm{t} / \mathrm{h}$ in 1996 from $0.259 \mathrm{t} / \mathrm{h}$ in 1995. The bootstrapped $95 \%$ confidence intervals of median CPUE estimates (Efron 1989) are large for years prior to 1988 due to the limited fishery (few observations) and narrow considerably after 1990. There was a significant decline in CPUE, partly due to vessel trip limits imposed, between 1992 and 1996.

### 1.3.2.4. Stock status

The significant expansion of the fishery on this stock over area and depth and the change in the stock age structure indicated in last year's assessment (Fargo and Kronlund 1997) indicate that the stock is probably fully exploited. Unfortunately, no biological samples were obtained from this stock from the 1996 fishery. Analysis of age composition data over time will provide estimates of mortality rates for the stock, critical information for this assessment. The effect of lower removals and lower fishing effort in 1996 cannot be assessed for several years. Landings should be restricted to that level until a more detailed assessment can be undertaken.

### 1.3.2.5 Recommendations and yield options

Low-risk yield option: A yield of 1000 t , appears sustainable based on the trend in commercial CPUE between 1988 and 1990.

High-risk yield option: Yields above 1500 t observed from 1991 to 1995 are associated with a significant decline in the commercial CPUE index. A yield of 1500 t can be considered as an upper limit of the sustainable range for this stock at the present time.

### 1.4 Queen Charlotte Sound (Areas 5A and 5B)

### 1.4.1 Rock sole

### 1.4.1.1 Introduction

The rock sole (Pleuronectes bilineata) is a minor componen: of the shelf, on-bottom trawl fishery in Queen Charlotte Sound and Hecate Strait. Four discrete stocks have been identified based on results from numerous tagging experiments (Ketchen 1982, Fargo and Westrheim 1987). Landings of rock sole are coincidental with landings of lingcod (Ophiodon elongatus) and Pacific cod (Gadus macrccephalus) in Queen Charlotte Sound. Yield is dependent upon recruitment which has been highly variable over time (Fargo 1995). Rock sole recruit to the fishery at age 4 but recruitment is not knife-edged. Length of $50 \%$ maturity, $\mathrm{L}_{50}$, is $32.4 \mathrm{~cm}(5 \mathrm{y})$ for females and 27.6 $\mathrm{cm}(4 \mathrm{y})$ for males. Managers have used a coastwide trip limit as a catch limitation measure for this species. This interim assessment of the Area 5A and 5B stycks is based on analysis of catch-effort data.

The trawl fishery for rock sole in areas 5 A and 5 B was unregulated prior to 1986. During the period from 1986 to 1992 a $30,000 \mathrm{lb}$ trip limit 'was imposed followed by a $20,000 \mathrm{lb}$ trip limit in 1993. Various trip limits less than $2 \mathrm{C}, 000 \mathrm{lb}$ were used by managers in recent years. Quotas for 1996 were set for each quarter year with a total harvest of 880 t allowed for areas 5A and 5B.

### 1.4.1.2 Area 5A Landing statistics

Landing statistics for rock sole from the 5A trawl fishery are presented in Fig. 4.5 and Table 4.6. These landings include contributions from the U.S. fishery in this area prior to 1978. The CPUE index for this fishery deteriorated beginning in the 1980s because of the effect of regulations on the fishery since the 1980. For this reason it may not be useful as a measure of stock abundance. Interpretation of the catch-effort data for area 5 A rock sole is further complicated by the fact that this stock is a minor component of the Area 5A multispecies trawl fishery. Also, the statistics for 1996 are not directly comparable to those in 1995 due to the mandatory at-sea observer program implemented by managers in 1996. Both landings and fishing effort for this stock decreased significantly in 1996. Landings of rock sole in Area 5A decreased to 87 t in 1996 from 212 t in 1995 while effort decreased to 540 h in 1996 from 939 h in 1995 and CPUE decreased to $0.102 \mathrm{t} / \mathrm{h}$ in 1996 from $0.138 \mathrm{t} / \mathrm{h}$ in 1995. Although it is difficult to interpret because of changing fishing patterns and regulatory measures a decline in CPUE since 1992 is apparent.

### 1.4.1.3 Stock status

Age composition data are not available for this stock but analysis of size composition data from port samples in last year's assessment indicate that recruitment has declined in recent years (Fargo and Kronlund 1997). We expect that yield for this stock will decrease over the next several years. However, the level of fishing effort observed in 1996 has probably reduced the risk of overfishing for this stock.

### 1.4.1.4. Recommendations and Yield Options

There is no change in the yield options recommended for this stock in 1998.

Low risk yield option A yield of 200 t , equivalent to the low-risk yield for lastser year's assessment, appears to be sustainable at this time.

High risk yield option: Yields greater than 400 t , the maximum annual yield observed can be considered as a upper limit of the sustainable yield for this stock..

### 1.4.1.5 Area 5B Landing statistics

Landing statistics for rock sole from the 5B trawl fishery are presented in Fig. 4.6 and Table 4.7. These landings include contributions from the U.S. fishery in this area prior to 1978. The CPUE index for this fishery deteriorated beginning in the 1980s because of the effect of regulations on the fishery since the 1980 . For this reason it may not be useful as a measure of stock abundance. Interpretation of the catch-effort data for area 5 B rock sole is further complicated by the fact that this stock is a minor component
of the Area 5B multispecies trawl fishery. Also, the statistics for 1996 are not directly comparable to those in 1995 because of the mandatory at-sea observer program implemented by managers in 1996. Landings and effort for this stock have derreased steadily since 1991, partly due to lower vessel trip limits applied by managers. Landings in 1996 were 231 t , down from 252 t in 1995. Effort in 1996 was 842 h similar to effort in 1995 while CPUE in 1996 increased to $0.176 \mathrm{t} / \mathrm{h}$ in 1996 from $0.150 \mathrm{t} / \mathrm{h}$ in 1995. Although it is difficult to interpret because of changing fishing patterns and regulatory measures a decline in CPUE is apparent since the early 1990s.

### 1.4.1.6 Stock status

Age composition data are not available for this stock but analysis of size composition data from port samples in last year's assessment indicate that recruitment has declined in recent years (Fargo and Kronlund 1997). We therefore expert that yield for this stock will decline in the near future. The impact of the decline in fishing effort on this stock over the last two years should help reduce the risk of overfishing bu.t this effect cannot be properly assessed until more detailed analysis can be performed in light of the loss of the CPUE index. This will not be possible in the near future.

### 1.4.1.7 Recommendations and yield options

The risk options for 1998 are the same as those recommended in 1997.
Low risk yield option: A yield of 200 t , is sustainable with low risk to the area 5B stock.

High risk yield option: Yields greater than 500 t , equivalent to the historic maximum, constitute a greater risk to the area 5 B stock.

### 1.5.1. Rock sole - Hecate Strait

### 1.5.1.1. Introduction

Stock delineation studies conducted by Ketchen (1982) and Fargo and Westrheim (1987) indicate that there are probably multiple stocks of rock sole in Hecate Strait. However, these stocks are treated as a single unit for this assessment. Past work has suggested that both density-dependent and density-independent factcrs regulate abundance of this species. Two significant determinants of recruitment for this stock are spawning stock size and ocean temperature at the time of spawning. Low recruitment has been associated with low spawning biomass and warm ocean temperatures (Firrester and Thomson 1969, Fargo and McKinnell 1989). Recruitment for these stocks has fluctuated greatly over time with the last significant increase occurring during the late 1980s and early 1990s. Landing statistics have been updated with data from the 1996 observer program and the age composition data series has been updated with 1996 fishery samples.

Age composition data are available for this stock and analysis of these data is the basis for this assessment.

### 1.5.1.2. Landing statistics

Landing statistics for rock sole in Hecate Strait are presented in Table 4.8 and Figure 4.7. Annual catch statistics for the 1945-96 period are calculated directly from data observations. No data records exist prior to 1954 and the index of Forrester and Thomson (1969) was used.

Landings decreased to 670 t in 1996 from 1294 t in 1995 while effort decreased to 2336 h in 1996 from 3538 h . Median CPUE decreased to $0.207 \mathrm{t} / \mathrm{h}$ in 1996 from $0.322 \mathrm{t} / \mathrm{h}$ in 1995. Since the early 1980 s there is little contrast in the commercial CPUE series although stock abundance as determined from catch-age analysis has fluctuated greatly (Fargo and Kronlund 1997). Since 1980 area-specific trip limits have influenced fishing patterns of the fleet and for the most recent years hyperstability is apparent in the CPUE index.

### 1.5.1.3 Catch-age analysis

The age composition time series for this species was updated with data from the 1996 trawl fishery. Samples collected from Minor Area 4 (Major Area 8) were used for this analysis. This was due to differences among areas in age composition data. Samples from Minor area 4 constitute the longest unbroken time series of rock sole age composition for Hecate Strait. The series covers the period 1945-96 and encompasses a range of fish age between 3 and 21 y (Figure 4.8). The range of ages used for catch-age analysis was 4 to $12+$ with the last age group representing fish aged 12 years or older. Three year olds are not fully recruited and fish 12 and older were grouped together due to differences in the ageing technique over time. Otolith surface readings (1945-72) produce underestimates for the ages of older fish compared to otolith burnt cross-sections (1973-96).

The catch-age model of Schnute and Richards (1995) was used for this assessment. The model of Schnute and Richards is essentially similar to other catch-age models (Fournier and Archibald 1982, Methot 1989) but does differ in the specification of the model error structure. Parameters in the model likelihood include standard deviations $\sigma 1, \tau 1$, and $\tau 2$, corresponding to the error in the recruitment, biomass index and proportions at age, respectively. The variance ratio $\rho=\sigma_{1}{ }^{2} /\left(\sigma_{1}{ }^{2}+\tau_{2}{ }^{2}\right)$ must be specified in the likelihood calculation, analogous to emphasis factors in the stock synthesis model of Methot $(1989,1990)$. Details of the model are presented in Appendix A.1. The model was run with the instantaneous rate of natural mortality, M , assumed to be constant at 0.20 and $\rho$ fixed at 0.7. These values are the same as those used in last year's assessment (Fargo and Kronlund 1997). Input data included landed catch, proportions at age in the catch, weight at age and CPUE estimates for rock sole (adults) from the Hecate Strait surveys conducted between 1984 and 1996 (Figure 4.9).

The model residuals were examined to assess the fit to the data. There was no trend in the residuals from the model for any year or age group (Figure 4.10). The model fit was poorest (largest residuals) for the 1960 to 1980 period and for the 4-6, age groups. As well, there were negative residuals for the $12+$ age group for nearly all years. This is likely a result of the change in age determination methods. The model expected a greater proportion of older fish than that indicated in the age proportion data.

Biomass and recruitment trajectories from the model are presented in Figure 4.11. Results indicate that exploitable biomass for this stock increased significantly in the late 1980s, peaked in the early 1990s and is now declining. Recruitment has declined since 1993 (Fargo and Kronlund 1997). The estimate of exploitable biomass in 1996, $\mathrm{B}_{96}$, was 4738 t ( $95 \%$ c.i. $=3221 \mathrm{t}$, 6970 t ) fior the $95 \%$ confidence interval. This compares to $\mathrm{B}_{95}$ of $5963 \mathrm{t}(95 \% \mathrm{c} . \mathrm{i} .=3175 \mathrm{t}-8751 \mathrm{t})$ presented in last year's assessment.

The model estimate of fishing mortality for the stock in 1995 was 0.15 , below the level of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {med }}$ (Fargo and Kronlund 1997). $\mathrm{F}_{0.1}$ is the fishing mortality rate which generates a marginal increase in yield per recruit of $10 \%$ of that from a lightly exploited stock (Gulland and Boerema 1973). This has been determined to bf: the fishing mortality rate at which marginal yield per recruit was $10 \%$ of the slope at $\mathrm{F}=0.0$ (Anthony 1982). For Hecate Strait rock sole this corresponds to $\mathrm{F}=0.22$. Stock-recruit data suggest that this stock has a $50 \%$ probability of maintaining its spawning stock bionass (SSB) with a fishing mortality equivalent to $0.21\left(\mathrm{~F}_{\mathrm{med}}\right)$. This fishing mortality rate was determined from the spawning stock biomass per recruit curve for the stock as the line which bisect the data. The ratio of SSB per recruit determined by this method was then located on an SSB per recruit projection plot for different values of fishing inortality to locate the appropriate value of fishing mortality (Patterson 1992). $\mathrm{F}_{\text {met }}$ has been suggested by the ICES Comprehensive Fishery Evaluation Working Group as the target fishing mortality rate most consistent with precautionary approach to fisheries suggested by the United Nations (FAO Fisheries Technical Paper 350/1).

### 1.5.1.4 Stock status

The results of the catch-age analysis indicate a significant increase i:1 stock biomass accompanied by strong recruitment occurred in the late 1980s. By the mid 1990s there declines in recruitment and biomass had occurred. CPUE frem research trawl surveys conducted in Hecate Strait has declined since the early 1990s as well. In addition, the El Nino event along the B.C. coast in 1997 will produce unfavourable temperature conditions for rock sole year-class production (Fargo and McKinnell 1989). Thus, yield and recruitment for this stock should continue to decline over the next several years.
1.5.1.5 Recommendations and yield options

The target fishing mortality rate $\mathrm{F}_{\text {med }} 0.21$, was used :o estimate sustainable yield for the stock. As a precautionary strategy last year, yield was estimated
using the 25 th and 50 th percentiles of the $95 \%$ confidence region for the terminal biomass estimate. Using the estimate $B_{96}$ the corresponding yield range corresponding to the 25 th and 50th percentiles of the biomass distribution are 766 t and 897 t , respectively. The yield options for 1998 remain unchanged from last year. The low-risk option may be more appropriate at this time than the high-risk option given the decline in stock biomass.

Low risk yield option A yield of 800 t , is the low-risk sustainable option for these stocks.

High risk yield option A yield of 1100 t , is the high-risk sustainable option for these stocks.

### 1.5.2. English sole - Hecate Strait

### 1.5.2.1. Introduction

Stock delineation studies conducted by Ketchen (1956) and Fargo et al. (1984) indicate that a single stock of English sole is resident in Hecate Strait. The stock was probably near the pristine level in the 1940s, declined after large removals in the early 1950s and has remained fairly stable since the late 1960s (Fargo and Kronlund 1997). Both density dependent and density independent factors exert significant influence on recruitment for this stock (Fargo 1994). Spawning stock biomass and Ekman transport during the egg and larval stages influence year-class production for this stock. The stock has produced strong year-classes about once a decade with the latest increase in recruitment occurring in the early 1990s. The age of recruitment is 4 years for both males and females although recruitment is not knife-edged. Length of $50 \%$ maturity, $\mathrm{L}_{50}$, is 25.5 cm ( 3 y ) for males and $35.1 \mathrm{~cm}(4 \mathrm{y})$ for females (Foucher et al. 1989). The contribution of strong year-classes to the fishery usually lasts about 4-5 years. The series of annual landing statistics has been updated with estimates from the 1996 fishery. The age composition data series has been updated with data from samples collected from the 1996 fishery as well. Age composition data are available for this stock and analysis of these data is the basis for this assessment.

### 1.5.2.2. Landing statistics

Annual landing statistics are presented in Table 4.9 and Figure 4.12. Statistics for 1954-96 are calculated directly from data observations. No detailed records exist prior to 1954 and the historical catch index of Ketchen has been used (Fargo 1994). English sole landings decreased to 455 t in 1996 from 1190 t in 1995 while effort decreased to 570 h from 2321 h over the same period. CPUE in 1996 decreased slightly to $0.310 \mathrm{t} / \mathrm{h}$ from $0.320 \mathrm{t} / \mathrm{h}$ in 1995. This stock has been under quota management since the late 1980s and this has influenced the fishing pattern of the fleet. In addition to this a mandatory observer program was initiated by managers for the 1996 trawl fishery and landing statistics for 1996 are not directly comparable to those for 1995.

### 1.5.2.3. Catch-age analysis

The age composition data series for this stock was updated with age determinations made from samples collected during the 1996 trawl fishery. The data series covers the period 1944-96 and includes fish ranged in age from 3 to 23 years. For the catch-age analysis, the full compliment of years was analysed over an age range of 4 to $12+$. Three year olds are not fully recruited while age groups older than 11 were combined because of the bias in age determinations made from otolith surface readings as compared to determinations made from otolith burnt cross-sections (see Secticn 4.i.1.3).

The catch-age model of Schnute and Richards (1995) was used for this assessment. The model of Schnute and Richards is essentially similar to other carch-age models (Fournier and Archibald 1982, Methot 1989) but does differ in the specilication of the model error structure. Parameters in the model likelihood incluile standard deviations $\sigma 1, \tau 1$, and $\tau 2$, corresponding to the error in the recruitment, bicmass index and proportions at age, respectively. The variance ratio $\rho=\sigma_{1}{ }^{2} /\left(\sigma_{1}{ }^{2}+\tau_{2}{ }^{2}\right)$ must je specified in the likelihood calculation, analogous to emphasis factors in the stock syntiesis model of Methot $(1989,1990)$. The model was run with the instantaneous rate: of natural mortality, M, assumed to be constant at 0.20 and $\rho$ fixed at 0.7 . These values are the same as those used in last year's assessment (Fargo and Kronlund 1997). Ditails of the model are presented in Appendix A.1.

Input data for the model included landed catch, proportions at age in the catch (Figure 4.13) weight at age and CPUE estimates for English sole from the Hecate Strait research trawl surveys conducted between 1984 and 1996 (Figure 4.14). For last year's assessment the catch-age model was tuned with the commercial CPUE series rather than the survey CPUE series. We felt it was more appropriate to use the research survey CPUE than the index from the commercial fishery because of problerss with the commercial index (see Section 4.5.2.2).

Model residuals were examined for indications of problems with the model fit (Figure 4.15). There was a negative trend in residuals for the early years (1944 to 1952) with respect to age while no such trend was present for the later years (19158-96). The residuals showed no trend with respect to time for any of the age groups analysed. The model fit was poorest for the younger age groups in the early years and for the older age groups in the later years.

Biomass and recruitment trajectories from the model are presented in Figure 4.16. The estimate of exploitable biomass in 1995, $\mathrm{B}_{95}$, from last year's assessment was $3177 \mathrm{t}(95 \%$ c.i. $=2766 \mathrm{t}-3588 \mathrm{t})$. The estimate of $\mathrm{B}_{96}$, was $2071 \mathrm{t}(95 \%$ c.i. $=2590 \mathrm{t}-3641 \mathrm{t}$ ). Exploitable biomass for this stock has declined since 1993 while the level of recruitment has declined since 1991.

The catch-age model estimate of fishing mortality for the stock in 1996 was 0.16, below the level of $\mathrm{F}_{0.1}(\mathrm{~F}=0.17)$ and $\mathrm{F}_{\text {med }}(\mathrm{F}=0.19)$, (see Section 4.5.1.3 for an explanation of these fishing rates). $\mathrm{F}_{\text {med }}$ has been suggested by the ICES Cornprehensive

Fishery Evaluation Working Group as the target fishing mortality rate most consistent with the UN precautionary approach to fisheries (FAO Fisheries Technical Paper 350/1)

### 1.5.2.4 Stock status

The estimate of fishing mortality, F, for the stock in 1996 from the new catch-age analysis was 0.16 , significantly lower than the estimate of 0.37 for 1995 when over 1000 t were removed associated with a high level of effort ( $\mathrm{F}=0.37$ ). The estimate, $\mathrm{B}_{96}, 3071 \mathrm{t}(95 \%$ c.i. $=2590 \mathrm{t}-3641 \mathrm{t})$ was used to estimate yield. $\mathrm{F}_{\text {med }}(\mathrm{F}=0.19)$ was the target fishing mortality rate that was used to estimate yield. This fishing rate was applied to the 25 th and 50th percentiles of the confidence region for $\mathrm{B}_{96}$ to produce a yield range of $496 \mathrm{t}-591 \mathrm{t}$. This range is similar to that in last year's assessment and the yield options for 1998 remain unchanged from those for 1997.

### 1.5.2.5. Recommendations and yield options

Low risk yield option A yield of $500 \mathrm{t} t$, is the low-risk option.

High risk yield option A yield of 600 t , is the high-risk option.

### 1.5.3 Dover sole

### 1.5.3.1 Introduction

The fishery for Dover sole in Areas 5C-E takes place in northern Hecate Strait at 100 to 160 m depths between May and October, and off the west coast of the Queen Charlotte Islands at 400 to 800 m depths from December to April. The seasonal shift in the fishery is related to the bathymetric spawning migration for the species. The fishery off the west coast of the Queen Charlotte Islands takes place on a spawning population. Dover sole begin to recruit to the fishery at 5 years of age but are not fully recruited until age 7-8. Length of $50 \%$ maturity, $\mathrm{L}_{50}$, is $37.1 \mathrm{~cm}(5-6 \mathrm{y})$ for males and $39.5 \mathrm{~cm}(6-7 \mathrm{y})$ for females. The Dover sole fishery in area 5C-E was unregulated prior to 1981. Beginning in 1981, annual quotas were applied: 300 t from 1981 to $1984,500 \mathrm{t}$ from 1985 to 1990, 1000 t from 1991 to 1994, and 1100 t in 1995 and 1996.

### 1.5.3.2 Area 5CDE Landing statistics

Landing statistics for Dover sole from the Area 5C-E trawl fishery for 1970-96 are presented in Table 4.10 and Figure 4.17. Landings decreased to 1133 t in 1996 from 1587 t in 1995 while effort decreased to 2245 h in 1996 from 4220 h in 1995 and CPUE decreased to $0.308 \mathrm{t} / \mathrm{h}$ in 1996 from $0.320 \mathrm{t} / \mathrm{h}$ in 1995. Landings for 1993 to 1995 were the highest recorded for this fishery. The CPUE series exhibits the characteristic high variability in the early years due to low numbers of observations with fishing up (positive trend in CPUE) occurring between the late 1970s and the late 1980s.

A marked increase in effort beginning in the late 1980s is associated with ar increase in catch and a decrease in CPUE (Fargo and Kronlund 1997). The CPUE estimate for 1996 is the lowest in the time series. Unlike in previous assessments Dover sole are the main target of the trawl fishery in these areas and vessel trip limits have never been applied. The fishery has been subjected to quota management since 1979. The CPUE and effort data from this fishery were analysed for this assessment.

### 1.5.3.3. Surplus production analysis

A dynamic surplus production model based on the Gompertz growth model (Yoshimoto and Clarke 1992) was employed to estimate maximum sustainable yield, MSY, and optimum effort, $\mathrm{E}_{\mathrm{opt}}$, for the stock. The details of the model are presented in Appendix A.2. Input data for the model included median CPUE, and fishing effort series for the period 1970-96. Results from this analysis indicated a maximum sustainable harvest of 1048 t for this stock at the present time. The optimum effort level corresponding to MSY was 2288 h very close to the level of effort in 1996.

### 1.5.3.4. Biological data

Age composition for females from samples collected from the 1996 commercial fishery is presented in Figure 4.18. The instantaneous rate of total mortality, Z , was estimated as the slope in the regression of the $\ln$ (numbers) on age ranged from 0.15 to 0.27 (Figure 4.19, Table 4.11). The instantaneous rate of natural mortality for this stock is estimated to be between 0.10 and 0.15 (Fargo and Westrheim 1985;. Using the midpoint of the range of total mortality ( $\mathrm{Z}=0.22$ ) and natural mortality ( $\mathrm{I} \mathrm{M}=0.12$ ) the fishing mortality rate for this stock in 1996 would be $\mathrm{F}=0.10$, slightly lower than the natural mortality rate for the stock. Thus, the stock is close to being fully exploited at the present time. The estimates of $Z$ from this analysis must be considered preliminary at this time. A time series of age composition data is necessary to estimate the total mortality rate of recent cohorts and corroborate preliminary results of the analysis presented here.

### 1.5.3.5 Area 5CDE Recommendations and yield options

The high-risk yield option was exceeded by 18 to 32 percent during 1993 to 1995 . The decline in CPUE since that time suggests that yields above 12 j 0 t increase the risk of overfishing.

Low risk yield option: A yield of 800 t , equivalent to the MSY estimated using surplus production analysis.

High risk yield option: A yield of 1200 t is suggested as an upper limit for yield for the area SCDE Dover sole stock.

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Table 4.1. Yield options for British Columbia flatfish species/stocks 1997-98

| Species | Area | 1997 |  | 1998 |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Low risk | High risk | Low risk | High risk |
| Petrale sole | Coastwide | incidental catches only |  | incidental catches only |  |
| Dover sole | Area 3C-D | 1000 t | 1500 t | 1000 t | 1500 t |
|  | Area 5C-E | 800 t | 1200 t | 800 t | 1200 t |
| rock sole | Area 5A | 200 t | 400 t | 200 t | 400 t |
|  | Area 5B | 200 t | 500 t | 200 t | 500 t |
|  | Area 5C-D | 800 t | 1100 t | 800 t | 1100 t |
| English sole | Area 5C-D | 500 t | 600 t | 500 t | 600 t |

Table 4.2. Canada-U.S. landings (t) of petrale sole from southwest Vancouver Island, Area 3C, 1945-96.

| Year | Flattery Spit | Area 3C north | Total Area 3C | Total Canadian | Year | Flattery Spit | Area 3C north | Total Area 3C | Total Canadian | CPUE ${ }^{2}$ (th) | Effort ${ }^{6}$ <br> (h) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1942 | - | - | 1561 | - | 1969 | 255 | 142 | 397 | 52 | - | - |
| 1943 | - | - | 2264 | . | 1970 | 80 | 198 | 278 | 142 | - | - |
| 1944 | - | - | 1489 | - | 1971 | 74 | 523 | 597 | 366 | - | - |
| 1945 | - | - | 718 | - | 1972 | 22 | 561 | 583 | 426 | - | - |
| 1946 | - | - | 906 | - | 1973 | 211 | 452 | 663 | 328 | - | - |
| 1947 | - | - | 627 | - | 1974 | 230 | 684 | 914 | 466 | - | - |
| 1948 | - | - | 1321 | - | 1975 | 474 | 465 | 939 | 295 | - | - |
| 1949 | - | - | 1178 | - | 1976 | 304 | 453 | 757 | 172 | - | - |
| 1950 | - | - | 854 | 362 | 1977 | 157 | 311 | 468 | 311 | - | - |
| 1951 | - | - | 794 | 293 | 1978 | 287 | 126 | 413 | 126 | - | - |
| 1952 | - | - | 948 | 419 | 1979 | 256 | 92 | 348 | 92 | - | - |
| 1953 | - | - | 748 | 367 | 1980 | 147 | 115 | 262 | 115 | - | - |
| 1954 | - | - | 664 | 279 | 1981 | 125 | 180 | 305 | 180 | - | - |
| 1955 | - | - | 415 | 142 | 1982 | 45 | 232 | 277 | 232 | - | - |
| 1956 | 40 | 585 | 625 | 173 | 1983 | 179 | 183 | 362 | 183 | - | - |
| 1957 | 9 | 629 | 638 | 200 | 1984 | 237 | 218 | 455 | 218 | - | - |
| 1958 | 19 | 609 | 628 | 144 | 1985 | 122 | 147 | 269 | 147 | - | - |
| 1959 | 33 | 1072 | 1105 | 159 | 1986 | 75 | 197 | 272 | 197 | - | - |
| 1960 | 233 | 974 | 1207 | 174 | 1987 | 113 | 123 | 236 | 123 | 0.392 | 12 |
| 1961 | 375 | 1109 | 1484 | 156 | 1988 | 185 | 183 | 368 | 183 | 0.42 | 102 |
| 1962 | 215 | 850 | 1065 | 135 | 1989 | 191 | 386 | 587 | 386 | 0.352 | 450 |
| 1963 | 90 | 658 | 748 | 66 | 1990 | 134 | 478 | 612 | 478 | 0.316 | 599 |
| 1964 | 71 | 530 | 601 | 141 | 1991 | 106 | 408 | 514 | 408 | 0.217 | 1026 |
| 1965 | 140 | 658 | 798 | 118 | 1992 | 260 | 128 | 388 | 128 | 0.18 | 548 |
| 1966 | 118 | 512 | 630 | 90 | 1993 | 200 | 248 | 448 | 248 | 0.139 | 926 |
| 1967 | 106 | 259 | 365 | 104 | 1994 | 189 | 139 | 328 | 139 | 0.114 | 453 |
| 1968 | 114 | 233 | 347 | 110 | 1995 | 195 | 158 | 353 | 158 | 0.12 | 414 |
|  |  |  |  |  | 1996 | 202 | 112 | 314 | 112 | 0.233 | 139 |

${ }^{2}$ Area 3C north $25 \%$ qualified CPUE (January - March)


Table 4.3. Canada-U.S. landings (t) of petrale sole from Areas 3D, SA-D, 1944-95.

| Year | Area 3D | $\begin{aligned} & \text { Areas } \\ & 5 \mathrm{~A}-\mathrm{B} \end{aligned}$ | Areas $5 C-D$ | Total Canadian | Total | Year | Area $3 D$ | Areas $5 \mathrm{~A}-\mathrm{B}$ | $\begin{gathered} \text { Areas } \\ 5 \mathrm{C}-\mathrm{D} \end{gathered}$ | Total | Total Canadian | CPUE <br> (th) | Effort <br> (h) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1944 | 499 | 303 | - | 802 | - | 1969 | 262 | 114 | 22 | 398 | 101 | - | - |
| 1945 | 270 | 1535 | 193 | 1998 | - | 1970 | 136 | 56 | 22 | 214 | 65 | - | - |
| 1946 | 623 | 1258 | 494 | 2375 | - | 1971 | 127 | 97 | 55 | 280 | 118 | - | - |
| 1947 | 469 | 986 | 769 | 2224 | - | 1972 | 50 | 154 | 33 | 237 | 102 | - | - |
| 1948 | 943 | 920 | 3011 | 4874 | - | 1973 | 197 | 211 | 24 | 432 | 78 | - | - |
| 1949 | 316 | 429 | 1644 | 2390 | - | 1974 | 196 | 283 | 14 | 493 | 85 | - | - |
| 1950 | 694 | 569 | 700 | 1963 | 435 | 1975 | 234 | 156 | 27 | 417 | 99 | - | - |
| 1951 | 305 | 326 | 642 | 1273 | 426 | 1976 | 153 | 132 | 30 | 315 | 118 | - | - |
| 1952 | 265 | 305 | 574 | 1144 | 249 | 1977 | 58 | 73 | 24 | 155 | 155 | - | - |
| 1953 | 235 | 450 | 46 | 731 | 92 | 1978 | 21 | 63 | 13 | 97 | 97 | - | - |
| 1954 | 712 | 234 | 300 | 1237 | 96 | 1979 | 10 | 57 | 39 | 106 | 106 | - | - |
| 1955 | 452 | 462 | 94 | 1008 | 118 | 1980 | 31 | 40 | 33 | 104 | 104 | - | - |
| 1956 | 291 | 528 | 53 | 872 | 68 | 1981 | 15 | 41 | 42 | 98 | 98 | - | - |
| 1957 | 1320 | 333 | 216 | 1869 | 198 | 1982 | 30 | 61 | 16 | 107 | 107 | - | - |
| 1958 | 174 | 227 | 171 | 572 | 205 | 1983 | 29 | 161 | 35 | 225 | 225 | - | - |
| 1959 | 227 | 160 | 216 | 603 | 175 | 1984 | 77 | 79 | 24 | 180 | 180 | - | - |
| 1960 | 93 | 212 | 120 | 425 | 238 | 1985 | 50 | 81 | 22 | 153 | 153 | - | - |
| 1961 | 277 | 171 | 102 | 550 | 192 | 1986 | 24 | 120 | 25 | 169 | 169 | - | - |
| 1962 | 295 | 343 | 165 | 803 | 331 | 1987 | 37 | 165 | 101 | 303 | 303 | - | - |
| 1963 | 202 | 537 | 82 | 821 | 329 | 1988 | 276 | 167 | 133 | 576 | 576 | 0.552 | 233 |
| 1964 | 183 | 421 | 163 | 767 | 359 | 1989 | 178 | 220 | 151 | 549 | 549 | 0.357 | 258 |
| 1965 | 300 | 418 | 202 | 920 | 363 | 1990 | 249 | 148 | 142 | 539 | 539 | 0.383 | 425 |
| 1966 | 264 | 469 | 260 | 993 | 465 | 1991 | 137 | 143 | 85 | 365 | 365 | 0.313 | 217 |
| 1967 | 169 | 485 | 176 | 830 | 350 | 1992 | 133 | 93 | 72 | 298 | 298 | 0.252 | 154 |
| 1968 | 293 | 266 | 137 | 696 | 257 | 1993 | 117 | 105 | 63 | 285 | 285 | 0.252 | 146 |
|  |  |  |  |  |  | 1994 | 53 | 197 | 45 | 295 | 295 | 0.118 | 34 |
|  |  |  |  |  |  | 1995 | 77 | 327 | 42 | 446 | 446 | 0.126 | 8 |
|  |  |  |  |  |  | 1996 | 52 | 68 | 25 | 145 | 145 | 0.271 | 57 |

- Area 3D 25\% qualified CPUE (January - March)
- Area 3D 25\% qualified effort (January - March)

Table 4.4 Estimates of the instantaneous total mortality rate, Z, for west coast Vancouver Island petrale sole.

| Age range | Z | $\mathrm{r}^{2}$ |
| :--- | :---: | :---: |
| $6-17$ | 0.15 | 0.54 |
| $6-20$ | 0.17 | 0.65 |
| $6-26$ | 0.18 | 0.75 |
| $6-35$ | 0.17 | 0.75 |
| $7-35$ | 0.17 | 0.78 |
| $8-35$ | 0.18 | $0.7 i$ |

Table 4.5 Annual landing statistics for the Area 3CD Dover sole trawl fishery, 1980.96.

| Year | Landings <br> $(\mathrm{t})$ | Effort $^{\mathrm{a}}$ <br> $(\mathrm{h})$ | CPUE $^{\mathrm{b}}$ <br> $(\mathrm{t})$ |
| :---: | :---: | :---: | :---: |
| 1980 | 184 | 306 | 0.556 |
| 1981 | 171 | 461 | 0.339 |
| 1982 | 129 | 281 | 0.361 |
| 1983 | 22 | 84 | 0.389 |
| 1984 | 24 | 79 | 0.256 |
| 1985 | 3 | 9 | 0.280 |
| 1986 | 2 | 8 | 0.321 |
| 1987 | 1 | 4 | 0.143 |
| 1988 | 371 | 620 | 0.426 |
| 1989 | 1115 | 1754 | 0.415 |
| 1990 | 1122 | 1882 | 0.402 |
| 1991 | 1222 | 2572 | 0.316 |
| 1992 | 1382 | 3034 | 0.357 |
| 1993 | 1785 | 4459 | 0.318 |
| 1994 | 1492 | 4626 | 0.267 |
| 1995 | 1630 | 5352 | 0.259 |
| 1996 | 1083 | 2318 | 0.229 |

Table 4.6. Canada-U.S. landings statistics for rock sole in Area 5A, 1954-96.

| Year | Landings (t) | Effort (h) ${ }^{\text {a }}$ | CPUE (t/h) ${ }^{6}$ | CPUE (t/h) ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 54 | 52 | 175 | 0.141 | 0.216 |
| 55 | 119 | 274 | 0.208 | 0.287 |
| 50 | 551 | 1441 | 0.230 | 0.241 |
| 57 | 511 | 1633 | 0.176 | 0.219 |
| 58 | 501 | 2204 | 0.160 | 0.181 |
| 59 | 212 | 834 | 0.124 | 0.162 |
| 60 | 397 | 1588 | 0.159 | 0.148 |
| 61 | 237 | 757 | 0.159 | 0.212 |
| 62 | 196 | 910 | 0.099 | 0.120 |
| 63 | 161 | 456 | 0.118 | 0.170 |
| 64 | 156 | 346 | 0.137 | 0.195 |
| 65 | 157 | 350 | 0.152 | 0.203 |
| 66 | 330 | 651 | 0.243 | 0.283 |
| 67 | 252 | 822 | 0.174 | 0.233 |
| 68 | 435 | 1224 | 0.196 | 0.233 |
| 69 | 293 | 1230 | 0.111 | 0.115 |
| 70 | 167 | 566 | 0.140 | 0.159 |
| 71 | 135 | 392 | 0.165 | 0.162 |
| 72 | 58 | 117 | 0.119 | 0.168 |
| 73 | 57 | 68 | 0.245 | 0.352 |
| 74 | 74 | 50 | 0.206 | 0.351 |
| 75 | 37 | 191 | 0.071 | 0.111 |
| 76 | 182 | 466 | 0.107 | 0.185 |
| 77 | 83 | 197 | 0.124 | 0.209 |
| 78 | 79 | 230 | 0.101 | 0.134 |
| 79 | 202 | 526 | 0.166 | 0.216 |
| 80 | 238 | 810 | 0.143 | 0.206 |
| 81 | 114 | 404 | 0.125 | 0.181 |
| 82 | 189 | 548 | 0.176 | 0.261 |
| 83 | 124 | 195 | 0.152 | 0.266 |
| 84 | 142 | 348 | 0.133 | 0.217 |
| 85 | 56 | 115 | 0.121 | 0.156 |
| 86 | 23 | 12 | 0.065 | 0.112 |
| 87 | 80 | 74 | 0.171 | 0.249 |
| 88 | 128 | 330 | 0.118 | 0.180 |
| 89 | 143 | 425 | 0.112 | 0.164 |
| 90 | 190 | 554 | 0.129 | 0.134 |
| 91 | 200 | 608 | 0.127 | 0.159 |
| 92 | 290 | 731 | 0.158 | 0.231 |
| 93 | 462 | 1864 | 0.135 | 0.199 |
| 94 | 311 | 1399 | 0.108 | 0.150 |
| 95 | 212 | 939 | 0.104 | 0.138 |
| 96 | 87 | 540 | 0.102 | 0.102 |

${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
${ }^{\text {b }}$ Adjusted for changes in vessel horsepower over time
${ }^{\text {c }}$ Median CPUE for $25 \%$ qualified landings

Table 4.7. Canada-U.S. landing statistics for rock sole in Area 5B, 1954-96.

| Year | Landings (t) | Effort ( ${ }^{\text {a }}{ }^{\text {a }}$ | CPUE (th) ${ }^{\text {b }}$ | CPUE (th) ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 54 | 203 | 133 | 0.518 | 0.295 |
| 55 | 267 | 259 | 0.280 | 0.247 |
| 56 | 307 | 614 | 0.294 | 0.270 |
| 57 | 206 | 531 | 0.249 | 0.302 |
| 58 | 379 | 1338 | 0.186 | 0.206 |
| 59 | 344 | 945 | 0.253 | 0.213 |
| 60 | 503 | 1444 | 0.227 | 0.203 |
| 61 | 416 | 1167 | 0.180 | 0.189 |
| 62 | 531 | 1345 | 0.222 | 0.227 |
| 63 | 517 | 947 | 0.233 | 0.225 |
| 64 | 482 | 559 | 0.186 | 0.193 |
| 65 | 568 | 729 | 0.216 | 0.226 |
| 66 | 772 | 794 | 0.261 | 0.253 |
| 67 | 741 | 423 | 0.324 | 0.280 |
| 68 | 392 | 492 | 0.244 | 0.246 |
| 69 | 652 | 1028 | 0.192 | 0.211 |
| 70 | 245 | 319 | 0.195 | 0.192 |
| 71 | 368 | 790 | 0.186 | 0.203 |
| 72 | 382 | 518 | 0.244 | 0.189 |
| 73 | 324 | 245 | 0.223 | 0.238 |
| 74 | 371 | 165 | 0.165 | 0.232 |
| 75 | 408 | 497 | 0.209 | 0.276 |
| 76 | 368 | 879 | 0.199 | 0.218 |
| 77 | 188 | 351 | 0.179 | 0.182 |
| 78 | 217 | 279 | 0.327 | 0.265 |
| 79 | 208 | 425 | 0.165 | 0.209 |
| 80 | 410 | 846 | 0.322 | 0.263 |
| 81 | 220 | 570 | 0.193 | 0.211 |
| 82 | 155 | 314 | 0.262 | 0.287 |
| 83 | 206 | 447 | 0.268 | 0.245 |
| 84 | 87 | 116 | 0.171 | 0.238 |
| 85 | 170 | 358 | 0.177 | 0.269 |
| 86 | 135 | 178 | 0.144 | 0.171 |
| 87 | 205 | 165 | 0.325 | 0.295 |
| 88 | 272 | 302 | 0.261 | 0.329 |
| 89 | 260 | 520 | 0.186 | 0.269 |
| 90 | 419 | 843 | 0.195 | 0.217 |
| 91 | 437 | 922 | 0.235 | 0.284 |
| 92 | 416 | 1203 | 0.164 | 0.227 |
| 93 | 343 | 1155 | 0.152 | 0.224 |
| 94 | 323 | 1023 | 0.168 | 0.215 |
| 95 | 252 | 848 | 0.142 | 0.150 |
| 96 | 231 | 842 | 0.176 | 0.176 |

${ }^{2}$ Annual effort for $25 \%$ qualified landings.
${ }^{\mathrm{b}}$ Adjusted for changes in vessel horsepower over time
${ }^{\text {c }}$ Median CPUE for $25 \%$ qualified landings

Table 4.8. Canada-U.S. landing statistics for Hecate Strait rock sole, 1945-96.

| Year | Landings (t) | Effort (h) ${ }^{\text {a }}$ | CPUE (th) ${ }^{\text {b }}$ | CPUE (th) ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 45 | 121 | 434 | 0.279 | 0.279 |
| 46 | 410 | 2228 | 0.184 | 0.184 |
| 47 | 1181 | 1946 | 0.607 | 0.607 |
| 48 | 901 | 1753 | 0.514 | 0.514 |
| 49 | 657 | 1352 | 0.486 | 0.486 |
| 50 | 784 | 1452 | 0.540 | 0.540 |
| 51 | 1024 | 944 | 1.085 | 1.085 |
| 52 | 2292 | 2014 | 1.138 | 1.138 |
| 53 | 779 | 1227 | 0.635 | 0.635 |
| 54 | 926 | 840 | 0.889 | 0.938 |
| 55 | 1560 | 1558 | 0.640 | 0.680 |
| 56 | 1160 | 1484 | 0.548 | 0.644 |
| 57 | 1151 | 2019 | 0.392 | 0.443 |
| 58 | 1256 | 1331 | 0.775 | 0.650 |
| 59 | 416 | 636 | 0.499 | 0.403 |
| 60 | 1127 | 1100 | 0.852 | 0.680 |
| 61 | 744 | 694 | 1.104 | 0.900 |
| 62 | 829 | 849 | 0.702 | 0.735 |
| 63 | 881 | 735 | 0.685 | 0.737 |
| 64 | 743 | 835 | 0.507 | 0.531 |
| 65 | 879 | 629 | 0.994 | 0.545 |
| 66 | 2544 | 2491 | 0.691 | 0.598 |
| 67 | 2162 | 2324 | 0.734 | 0.511 |
| 68 | 2366 | 4209 | 0.370 | 0.386 |
| 69 | 1461 | 4485 | 0.417 | 0.314 |
| 70 | 1403 | 3660 | 0.256 | 0.326 |
| 71 | 1503 | 3587 | 0.264 | 0.255 |
| 72 | 515 | 650 | 0.334 | 0.337 |
| 73 | 507 | 619 | 0.334 | 0.435 |
| 74 | 622 | 603 | 0.312 | 0.475 |
| 75 | 1204 | 1912 | 0.301 | 0.360 |
| 76 | 1438 | 1830 | 0.449 | 0.402 |
| 77 | 846 | 1896 | 0.237 | 0.285 |
| 78 | 874 | 1662 | 0.297 | 0.336 |
| 79 | 1313 | 1943 | 0.333 | 0.330 |
| 80 | 977 | 2420 | 0.166 | 0.254 |
| 81 | 584 | 806 | 0.228 | 0.287 |
| 82 | 291 | 841 | 0.174 | 0.209 |
| 83 | 247 | 499 | 0.194 | 0.286 |
| 84 | 188 | 573 | 0.127 | 0.188 |
| 85 | 112 | 276 | 0.201 | 0.242 |
| 86 | 219 | 470 | 0.219 | 0.345 |
| 87 | 536 | 577 | 0.262 | 0.389 |
| 88 | 1402 | 2520 | 0.322 | 0.41 |
| 89 | 1422 | 3757 | 0.228 | 0.288 |
| 90 | 1519 | 3948 | 0.239 | 0.319 |
| 91 | 2666 | 6552 | 0.224 | 0.295 |
| 92 | 2226 | 5777 | 0.241 | 0.289 |
| 93 | 2080 | 5851 | 0.224 | 0.301 |
| 94 | 1384 | 4282 | 0.201 | 0.275 |
| 95 | 1294 | 3538 | 0.234 | 0.322 |
| 96 | 670 | 2336 | 0.207 | 0.207 |

${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
${ }^{\text {b }}$ Adjusted for changes in vessel horsepower over time
${ }^{\mathrm{c}}$ Median CPUE for $25 \%$ qualified landings

Table 4.9. Canada-U.S. landing statistics for Hecate Strait English sole, 1944-96.

| Year | Landings (t) | Effort (h) ${ }^{\text {a }}$ | CPUE (t/h) ${ }^{\text {b }}$ | CPUE (th) ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 44 | 152 | 215 | 0.707 | 0.707 |
| 45 | 304 | 365 | 0.832 | 0.832 |
| 46 | 470 | 809 | 0.581 | 0.581 |
| 47 | 350 | 538 | 0.651 | 0.651 |
| 48 | 937 | 2740 | 0.342 | 0.342 |
| 49 | 795 | 1893 | 0.420 | 0.420 |
| 50 | 2622 | 4910 | 0.534 | 0.534 |
| 51 | 1024 | 2142 | 0.478 | 0.478 |
| 52 | 1347 | 3293 | 0.409 | 0.409 |
| 53 | 871 | 2084 | 0.418 | 0.418 |
| 54 | 455 | 563 | 0.245 | 0.362 |
| 55 | 875 | 744 | 0.332 | 0.401 |
| 56 | 956 | 1344 | 0.308 | 0.349 |
| 57 | 552 | 640 | 0.180 | 0.244 |
| 58 | 693 | 617 | 0.251 | 0.337 |
| 59 | 940 | 772 | 0.279 | 0.315 |
| 60 | 1147 | 1058 | 0.307 | 0.333 |
| 61 | 871 | 1615 | 0.230 | 0.298 |
| 62 | 459 | 903 | 0.212 | 0.247 |
| 63 | 408 | 568 | 0.154 | 0.207 |
| 64 | 436 | 441 | 0.243 | 0.272 |
| 65 | 414 | 326 | 0.198 | 0.317 |
| 66 | 362 | 354 | 0.190 | 0.302 |
| 67 | 534 | 535 | 0.373 | 0.411 |
| 68 | 671 | 844 | 0.285 | 0.302 |
| 69 | 819 | 1314 | 0.276 | 0.390 |
| 70 | 1002 | 2042 | 0.262 | 0.312 |
| 71 | 488 | 1585 | 0.168 | 0.192 |
| 72 | 371 | 550 | 0.197 | 0.230 |
| 73 | 667 | 514 | 0.294 | 0.411 |
| 74 | 500 | 519 | 0.372 | 0.519 |
| 75 | 938 | 1015 | 0.383 | 0.466 |
| 76 | 1133 | 1627 | 0.234 | 0.275 |
| 77 | 1179 | 2201 | 0.200 | 0.310 |
| 78 | 559 | 944 | 0.161 | 0.246 |
| 79 | 864 | 980 | 0.237 | 0.337 |
| 80 | 995 | 1105 | 0.220 | 0.327 |
| 81 | 1327 | 2149 | 0.204 | 0.249 |
| 82 | 428 | 1062 | 0.155 | 0.219 |
| 83 | 430 | 834 | 0.163 | 0.240 |
| 84 | 658 | 1129 | 0.221 | 0.290 |
| 85 | 585 | 1520 | 0.166 | 0.226 |
| 86 | 335 | 469 | 0.244 | 0.365 |
| 87 | 630 | 396 | 0.336 | 0.347 |
| 88 | 688 | 540 | 0.324 | 0.493 |
| 89 | 826 | 925 | 0.294 | 0.385 |
| 90 | 992 | 1335 | 0.224 | 0.383 |
| 91 | 913 | 940 | 0.208 | 0.308 |
| 92 | 987 | 1602 | 0.239 | 0.307 |
| 93 | 1421 | 2636 | 0.191 | 0.295 |
| 94 | 1000 | 1860 | 0.200 | 0.343 |
| 95 | 1190 | 2321 | 0.219 | 0.320 |
| 96 | 455 | 570 | 0.310 | 0.310 |

[^0]Table 4.10. Canada-U.S. landing statistics for Dover sole, Areas 5C-E, 1970-96.

| Year | Landings (t) | Effort (h) ${ }^{\text {a }}$ | CPUE (t/h) ${ }^{\text {b }}$ | CPUE (t/h) ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 70 | 965 |  | 0.432 | 0.590 |
| 71 | 903 | 1367 | 0.380 | 0.556 |
| 72 | 922 | 1495 | 0.331 | 0.543 |
| 73 | 768 | 910 | 0.481 | 0.679 |
| 74 | 767 | 878 | 0.531 | 0.687 |
| 75 | 882 | 1135 | 0.339 | 0.573 |
| 76 | 1022 | 1465 | 0.297 | 0.440 |
| 77 | 577 | 900 | 0.217 | 0.319 |
| 78 | 483 | 650 | 0.252 | 0.497 |
| 79 | 697 | 1057 | 0.193 | 0.333 |
| 80 | 807 | 724 | 0.237 | 0.416 |
| 81 | 840 | 1079 | 0.275 | 0.428 |
| 82 | 512 | 894 | 0.242 | 0.433 |
| 83 | 693 | 544 | 0.284 | 0.568 |
| 84 | 953 | 1526 | 0.265 | 0.448 |
| 85 | 830 | 1039 | 0.258 | 0.485 |
| 86 | 1040 | 931 | 0.313 | 0.562 |
| 87 | 503 | 432 | 0.426 | 0.549 |
| 88 | 649 | 652 | 0.360 | 0.594 |
| 89 | 696 | 775 | 0.407 | 0.567 |
| 90 | 787 | 1181 | 0.279 | 0.542 |
| 91 | 649 | 1041 | 0.250 | 0.428 |
| 92 | 883 | 1444 | 0.206 | 0.381 |
| 93 | 1508 | 2767 | 0.225 | 0.414 |
| 94 | 1418 | 3117 | 0.243 | 0.371 |
| 95 | 1587 | 4220 | 0.199 | 0.320 |
| 96 | 1133 | 2245 | 0.308 | 0.308 |

${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
${ }^{\mathrm{b}}$ Adjusted for changes in vessel horsepower over time
${ }^{\text {c }}$ Median CPUE for $25 \%$ qualified landings

Table 4.11 Estimates of the instantaneous total mortality rate, Z, for Area 5C-E Dover sole.

| Age range | Z | $\mathrm{r}^{2}$ |
| :--- | :--- | :--- |
|  |  |  |
|  | 0.27 | 0.89 |
| $8-17$ | 0.22 | 0.89 |
| $8-22$ | 0.22 | 0.89 |
| $8-46$ | 0.15 | 0.81 |
| $9-17$ | 0.16 | 0.69 |
| $10-17$ | 0.25 | $0.8: 2$ |

The catch-age model used for the assessments of Hecate Strait rock and English soles is an application of the state space model developed by Schnute and Richards (1995). The model attempts to reconstruct the population history from known controls and observations. In this context, the catch biomass acts as a known control on the population dynamics. Observations, including proportions at age in the catch and a biomass index from survey CPUE values, describe the current state of the system. The model relates the observations, measured with error, to unknown numbers of fish in the population.

Table A. 1 contains a deterministic version of the model, with notation described in Table A.2. Equations in Table A. 1 are tailored for each application. In particular, the SchnuteRichards model is based on numbers of fish; we use known weights, $w_{a t}$ of fish of age $a$ in time $t$ and the maturity ogive $m_{a}$ to determine the spawning biomass $S_{t}$ and exploitable population biomass $B_{t}$.

Similar to other stochastic catch-age models, our analysis contains a separability assumption. The two parameters $\alpha$ and $\beta_{l}$ describe a selectivity function which is timeindependent and asymptotic with age. The quantity $\beta_{a}$ in equation (A.2) denotes the proportion of age $a$ fish that are vulnerable to the fishery.

Other quantities in the parameter vector $\Theta$ are the natural mortality $M$, the survey catchability $q$, and the time series of recruitments $R_{t}$. We treat the recruitments as parameters to be estimated from the data; our analysis does not contain an explicit stock-recruitment function.

The prediction equations (A.14) and (A.15) relate quantities $\bar{I}_{t}$ and $\bar{p}_{a t}$ obtained from the model dynamics to observations $I_{t}$ and $p_{a t}$ of survey CPUE and age proportions, respectively. (We use the convention of a bar over a quantity to denote a prediction for that quantity.) We assume in (A.14) that the survey CPUE indexes the population biomass after half of the annual catch has been removed. The catchability $q$ converts units of population biomass into units of CPUE. Although the relationship (A.14) could be made age-specific, age composition data are
not available for the early surveys. The predicted age proportions in the catch are obtained from the underlying population age structure in equation (A.4).

Schnute and Richards (1995) specify stochastic counterparts of the de:erministic equations (Table A.1), model residuals, and the model likelihood function. They impose three sources of error: (1) autoregressive lognormal process error among the recruitments $R_{t}$; (2) lognormal error in CPUE; and (3) multivariate logistic error in the observed proportions $p_{a t}$. These error structures lead to residual functions

$$
\begin{aligned}
& \xi_{t}=\log I_{t}-\log \bar{I}_{t} \\
& \eta_{a t}=\log p_{a t}-\log \bar{p}_{a t}-\frac{1}{A} \sum_{a=1}^{A}\left[\log p_{a t}-\log \bar{p}_{a t}\right]
\end{aligned}
$$

that describe model relationships between predictions and observations of survey CPUE and age proportions, respectively.

The likelihood for this catch-age model conforms to the errors-in-variables paradigm (Schnute 1994); apparent variations in abundance can be explained through high process error $\sigma$ in recruitment or high measurement error $\tau$ in CPUE. Schnute and Richards (1995) resolve this ambiguity by fixing the model variance ratio

$$
\rho=\frac{\sigma^{2}}{\sigma^{2}+\tau^{2}}
$$

between recruitment variance and total variance $\left(\sigma^{2}+\tau^{2}\right)$.
For the catch-age analysis, we fix the variance ratio $\rho=0.7$, a value that represents moderate levels of error in both recruitment and survey CPUE. Similar stock reconstructions were obtained for a range of reasonable choices of $\rho$ in preliminary model runs. 'Ne also employ a fixed natural mortality rate of $M=0.2$. Age classes in the model range from reciuits to the fishery at age 4 to an accumulator age class for age 12 and older. To reduce the influence on the model likelihood of very small age proportion observations (obtained from a small number of
fish), we group consecutive age classes such that $p_{a t} \geq 0.02$ for each age $a$ and time $t$ (Richards et al. 1997).

The model was implemented using AD Model Builder software (Otter Research Ltd. 1994). Standard errors for the model parameters and other quantities were obtained from the model hessian matrix. These allow calculation of symmetric confidence intervals, assuming that the parameter estimates have a multivariate normal distribution. In particular, we used AD Model Builder to compute standard errors for $\log$ recruitment, log spawner biomass and log exploitable biomass. The asymmetric confidence intervals illustrated in the figures were obtained by back transformation.

Table A.1. Deterministic catch-age model. Calculations begin with the parameter vector $\Theta$ and proceed recursively to define all states and observations.

## Parameters

(A.1)

$$
\Theta=\left(\alpha, \beta_{1}, M, q,\left\{R_{t}\right\}_{t=2-A}^{T}\right)
$$

Selectivity

$$
\begin{equation*}
\beta_{a}=1-\left(1-\beta_{1}\right)\left(\frac{A-a}{A-1}\right)^{a} \tag{A.2}
\end{equation*}
$$

## State moments

$$
\begin{equation*}
P_{t}=\sum_{a=1}^{A} \beta_{a} N_{a t} \tag{A.3}
\end{equation*}
$$

$$
\begin{equation*}
u_{a t}=\beta_{a} N_{a t} / P_{t} \tag{A.4}
\end{equation*}
$$

$$
\begin{equation*}
B_{t}=\sum_{a=1}^{A} \beta_{a} w_{a t} N_{a t} \tag{A.5}
\end{equation*}
$$

$$
\begin{equation*}
S_{t}=\sum_{a=1}^{A} m_{a} w_{a t} N_{a t} \tag{A.6}
\end{equation*}
$$

$$
\begin{equation*}
C_{t}=D_{t} / \sum_{a=1}^{A} u_{a t} w_{a t} \tag{A.7}
\end{equation*}
$$

(

Initial states

$$
\begin{equation*}
N_{a 1}=R_{2-a} e^{-M(a-1)} ; 1 \leq \mathrm{a}<\mathrm{A} \tag{A.9}
\end{equation*}
$$

(A.10)

$$
N_{A 1}=R_{2-A}\left(\frac{e^{-M(A-1)}}{1-e^{-M}}\right)
$$

State Dynamics

$$
\begin{equation*}
N_{1 t}=R_{t} \tag{A.11}
\end{equation*}
$$

$$
\begin{equation*}
N_{a t}=e^{-M}\left[N_{a-1, t-1}-u_{a-1, t-1} C_{t-1}\right] ; 2 \leq \mathrm{a}<\mathrm{A} \tag{A.12}
\end{equation*}
$$

(A.13)

$$
N_{A t}=e^{-M}\left[N_{A-1, t-1}+N_{A, t-1}-\left(u_{A-1, t-1}+u_{A, t-1}\right) C_{t-1}\right]
$$

Predicted Observations
(A.14)
$\bar{I}_{t}=q\left(B_{t}-0.5 D_{t}\right)$
(A.15)
$\bar{p}_{a t}=u_{a t} ; 2 \leq \mathrm{a} \leq \mathrm{A}$

Appendix table 4.1. Description of the notation for the input data, parameters, and other calculated model quantities in Table A.1.

Symbol
Description

Index quantities
$a$
$t$

D
$I_{t}$
$m_{a}$
$p_{a t}$
$w_{a t}$
$\Theta$
$\alpha$
$\beta_{l}$
M
$q$
$R_{t}$
$\beta_{a}$
$B_{t}$
$C_{t}$
$F_{t}$
$N_{a t}$
$P_{t}$
$S_{t}$
$u_{a t}$
age-class from 1 to $A$
year from 1 to $T$
Input data
observed catch biomass in year $t$
observed survey CPUE in year $t$
proportion of age-class $a$ fish which are mature observed proportion of age-class $a$ fish in the year $t$ catch weight of age-class $a$ fish in year $t$

Parameters
parameter vector
selectivity slope parameter
selectivity of age-class 1
natural mortality rate
catchability for survey CPUE age-class 1 recruitment in year $t$

Calculated quantities
selectivity for age-class $a$
exploitable population biomass at the start of year $t$
catch number in year $t$
fishing mortality rate in year $t$
number of age-class $a$ fish at the start of year $t$
exploitable population numbers at the start of year $t$
spawning biomass at the start of year $t$
exploitable proportion of age-class $a$ fish in year $t$ catch

Appendix A.2. The dynamic Fox surplus production model used in the Area 5C-E Dover sole assessment
$\ln \left(U_{t+1}\right)=(2 r /(2+r)) \ln (K q)+((2-r) /(2+r)) \ln \left(U_{t}\right)-(q /(2+r))\left(E_{t}+E_{t+1}^{\prime}\right)$
where:
$\mathrm{U}=$ median $\operatorname{CPUE}(\mathrm{t} / \mathrm{h})$
$\mathrm{E}=\mathrm{effort}(\mathrm{h})$
$r=$ natural growth rate
$\mathrm{q}=$ catchability coefficient
$\mathrm{K}=$ environmental carrying capacity

The model fit to the data was expressed as:
$\ln \left(U_{t+1}\right)=c 1+c 2 \ln \left(U_{t}\right)+c 3\left(E_{t}+E_{t+1}\right)$
where $\mathrm{r}, \mathrm{q}$, and K were:
$r=2(1-c 2) /(1+c 2)$
$q=-c 3(2+r)$
$K=(1 / q) \exp (c 1(2+r) / 2 r)$

Optimum effort, $E_{\text {opt }}$, was estimated as
the slope in the regression:
$U=a-E_{o p t} E$

Maximum sustainable yield, $M S Y$, was estimated as:

$$
M S Y=q k E_{o p t}\left(\exp \left(\left(-q E_{o p t}\right) / r\right)\right)
$$



Figure 4.1. Landings (t) from the 'southern' stock of petrale sole (top panel) in British Columbia, 1942-96 and from the 'northern' stock of petrale sole (bottom panel) in British Columbia, 1944-96.


Figure 4.2. Age composition of petrale sole caught in the 1993-94 trawl fishery in British Columbia.


Figure 4.3 Estimates of the instantaneous total mortality rate for west coast Vancouver Island Petrale sole. The filled squares indicate data used to fit the regression.


Figure 4.4. Landing statistics for Dover sole in Areas 3C-D, 1980,96. Top panel - bars $=$ landings, line $=$ effort ( h ). Bottom panel - Median CPUE and $9 \leq \%$ confidence interval.


Figure 4.5. Canada-U.S. landing statistics for rock sole in Area 5A. top panel - bars $=$ landings, line $=$ effort. Bottom panel - Median CPUE and $95 \%$ confidence interval.


Figure 4.6. Canada-U.S. landing statistics for rock sole in Area 5B. top panel - bars $=$ landings, line $=$ effort. Bottom panel - Median CPUE and $95 \%$ confidence interval.


Figure 4.7. Canada-U.S. landing statistics for rock sole in Hecate Strait. Top panel - bars $=$ landings, line $=$ effort. Bottom panel - Median CPUE and $95 \%$ confidence interval.


Figure 4.8. Input data for the catch-age analysis of Hecate Strait rock sole, 1945-96. The circle radii are proportional to values for individual age proportions.


Figure 4.9. Rock sole CPUE from research surveys conducted in Hecate Strait between 1984 and 1996.


Figure 4.10. Symbol plot of residuals for the age proportion data for the catcl.-age analysis of Hecate Strait rock sole by year. Circles represent negative residuc.ls winile squares represent positive residuals. Circle radii are scaled to the maximum negative residual while the area of each square is scaled to the maximum positive residual. A blank space represents a small age proportion that is grouped with the next older age. Residuals are defined in Appendix Table A.1.


Figure 4.11. Biomass and Recruitment trajectories and $95 \%$ confidence interval from the catchage model results for Hecate Strait rock sole, 1945-96.



Figure 4.12. Canada-U.S. landing statistics for English sole in Hecate Strait. Top panel - bars $=$ landings, line $=$ effort. Bottom panel - Median CPUE and $9 \leftrightarrows \%$ confidence interval.


Figure 4.13. Input data for the catch-age analysis of Hecate Strait English sole (ages 4-12+), 1945-96. The circle radii are proportional to values for individual age proportions.


Figure 4.14. English sole CPUE from research surveys conducted in Hecate Stra: between 1984 and 1996.


Figure 4.15. Symbol plot of residuals for the age proportion data for the catch-age analysis of Hecate Strait English sole by year. Circles represent negative residuals while squares represent positive residuals. Circle radii are scaled to the maximum negative residual while the area of each square is scaled to the maximum positive residual. A blank space represents a small age proportion that is grouped with the next older age. Residuals are defined in Appendix Table A.1.



Year

Figure 4.16. Biomass and Recruitment trajectories from the catch-age model results for Hecate Strait English sole, 1944-96.



Figure 4.17. Canada-U.S. landing statistics for Dover sole in Areas 5C-E. Top panel - bars $=$ landings, line $=$ Effort (h). Bottom panel - Median CPUE and $95 \%$ confidence interval.


Figure 4.18. Age composition for Dover sole caught in the 1996 commercial trawl fishery in Area 5D.


Figure 4.19 Estimates of the instantaneous total mortality rate for Area 5C-E Dover sole. The filled squares indicate data used to fit the linear regression.


[^0]:    ${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
    ${ }^{\mathrm{b}}$ Adjusted for changes in vessel horsepower over time
    ${ }^{\text {c }}$ Median CPUE for $25 \%$ qualified landings

